

**Receptor model source attributions for Utah's Salt Lake City
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ammonium nitrate and ammonium chloride aerosol**

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IMPLICATIONS

The study suggests that secondary ammonium chloride aerosol can be a significant source of wintertime $PM_{2.5}$ in an ammonia-rich environment, like the Wasatch Front, if sufficient sources of atmospheric chlorine exist. During winter-time, cold-air-pool events, the source attribution results generally agree with the county emission inventories with the exception of wood smoke and cooking sources. In Salt Lake City, the estimated contributions from wood smoke and cooking are nearly double those of the corresponding inventory, suggesting that they are nearly as important as gasoline emissions at this monitoring station.

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Title: Receptor model source attributions for Utah's Salt Lake City airshed and the impacts of wintertime secondary ammonium nitrate and ammonium chloride aerosol

Kerry E. Kelly^a, Robert Kotchenruther^b, Roman Kuprov^c, Geoffrey D. Silcox^a

a Dept. of Chemical Engineering, University of Utah, 380 INSCC, Salt Lake City, UT, 84112, USA, Kelly@eng.utah.edu; (801) 587 7601 (tel); (801) 585-9291 (fax)

Corresponding author

b U.S. EPA Region-10 Office of Environmental Assessment, 1200 Sixth Avenue, Suite 900, OEA-095, Seattle, WA, 98101-3140, USA

Kotchenruther.Robert@epa.gov; (206) 553-6218 (tel)

c Utah Division of Air Quality, 195 North 1950 West, Salt Lake City, UT, 84116, USA

rkuprov@utah.gov; 801-536-4254 (tel); 801-536-0601 (fax)

d Dept. of Chemical Engineering, University of Utah, 3290 MEB, Salt Lake City, UT, 84112, USA, geoff@che.utah.edu; (801) 581-8820 (tel); (801) 585-9291 (fax)

ABSTRACT

Communities along Utah's Wasatch Front are currently developing strategies to reduce daily average PM_{2.5} levels to below National Ambient Air Quality Standards during wintertime, persistent, multi-day stable atmospheric conditions or cold-air pools. Speciated PM_{2.5} data from the Wasatch Front airshed indicates that wintertime exceedances of the PM_{2.5} standard are mainly driven by high levels of ammonium nitrate. Stable wintertime conditions foster the formation of ammonium nitrate aerosol, if there are sufficient sources of NO_x, ammonia, and oxidative capacity. However, this work demonstrates that secondary ammonium chloride aerosol can also be a significant source of secondary wintertime PM_{2.5} if sufficient sources of atmospheric chlorine exist. Two factor analysis techniques, positive matrix factorization (PMF) and Unmix, were used to identify contributors to PM_{2.5} at three monitoring stations along Utah's Wasatch Front: Bountiful, Lindon, and Salt Lake City. The monitoring data included chemically speciated PM_{2.5} data for 227, 227, and 429 days at each location, respectively, during the period from May 2007 through May 2011. PMF identified 10 – 12 factors, and Unmix identified 4 - 5 factors for each of the locations. The wintertime PMF and Unmix results showed large contributions from secondary PM_{2.5} when PM_{2.5} concentrations exceeded 20 µg/m³. PMF identified both ammonium nitrate and ammonium chloride aerosol as significant secondary contributors to PM_{2.5} (10%-15% of total PM_{2.5} from ammonium chloride) during wintertime pollution episodes. Subsequent ion balance analysis of the monitoring data confirmed the

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3 presence of significant ammonium chloride aerosol on these highly polluted days at all three
4 monitoring sites. The primary PM_{2.5} portion of the source attribution results were further
5 compared to county-level emissions inventories and showed generally good agreement for Salt
6 Lake City and Lindon during wintertime except for wood smoke and fugitive dust, which have
7 higher contributions in the receptor modeling results than in the emissions inventories.
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12 13 **IMPLICATIONS**

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16 The study suggests that secondary ammonium chloride aerosol can be a significant source of
17 wintertime PM_{2.5} in an ammonia-rich environment, like the Wasatch Front, if sufficient sources
18 of atmospheric chlorine exist. During winter-time, cold-air-pool events, the source attribution
19 results generally agree with the county emission inventories with the exception of wood smoke
20 and cooking sources. In Salt Lake City, the estimated contributions from wood smoke and
21 cooking are nearly double those of the corresponding inventory, suggesting that they are nearly
22 as important as gasoline emissions at this monitoring station.
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29 30 **INTRODUCTION**

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32 Exposure to fine particulate matter (PM_{2.5}, particles with an aerodynamic diameter <2.5 μm) has
33 been linked to adverse human health effects, including increases in cardiovascular and
34 pulmonary disease (Baliff et al., 2000; Nicolai et al., 2003), and morbidity and mortality
35 (Dockery et al., 1993; Pope et al., 1991). PM_{2.5} also contributes to impaired visibility (Watson
36 2002) and changes in the global radiative balance (Chung and Seinfeld, 2002). In 2006, EPA
37 issued an updated 24-hour standard for PM_{2.5} of 35 μg/m³, and as a result in 2009 they declared
38 three regions in northern Utah along the Wasatch Front as nonattainment areas for 24-hr average
39 PM_{2.5}. The state of Utah is currently developing a plan to bring the PM_{2.5} concentrations to
40 attainment levels.
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50 The Wasatch Front typically experiences elevated levels of PM_{2.5} during wintertime, when high-
51 pressure weather systems and a high solar zenith angle lead to cold-air pools that periodically
52 trap aerosols in mountain valleys. These elevated PM_{2.5} levels cause adverse health effects
53 locally. For example, the State of Utah found that the odds of an emergency department visit in
54 Salt Lake County, with a primary diagnosis of asthma, are 42% greater during the 5th – 7th days
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3 of prolonged inversions than for non-inversion days (UDAQ, 2010). During a particularly
4 extreme cold-air pool event in 2004 PM_{2.5} concentrations exceeded 100 µg/m³ (Malek et al.,
5 2006), while more recent maximum concentrations have been in the range of 50 – 70 µg/m³.
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10 Several previous studies have examined the sources of fine particulate matter along the Wasatch
11 Front; however, compared to this work these studies used data that spanned a brief time period,
12 on the order of weeks. Hansen et al. (2010) collected speciated hourly PM_{2.5} data at the Utah
13 Division of Air Quality (UDAQ) Lindon air monitoring station for 10 days during the winter of
14 2007 and performed a source apportionment study using the Positive Matrix Factorization (PMF)
15 model. Their model results identified the following four primary sources: mobile diesel, mobile
16 gasoline, wood smoke, road dust and the following secondary sources: sulfate, nitrate, organic
17 matter, and aged wood smoke. Grover et al. (2006) performed an intensive air monitoring
18 campaign in August 2002 at the UDAQ Lindon air monitoring station and complementary source
19 apportionment using the Unmix model. From this small sample set, they found three main
20 contributors to PM_{2.5}: gasoline emissions, diesel emissions, and secondary aerosols. Their
21 analysis did not include inorganic species, so it is not surprising that they did not identify sources
22 of crustal material. The limited duration of these previous studies makes it difficult to draw
23 general conclusions about the sources of fine particulate matter along the Wasatch Front.
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37 This study investigates the sources of PM_{2.5} impacting three monitoring sites along the Wasatch
38 Front using ambient data collected from 2007 to 2011. Source apportionment is performed using
39 two receptor models, PMF and Unmix, and model results are compared to each other as well as
40 to emissions inventories. Analysis of chemically speciated PM_{2.5} in this work and in previous
41 studies (Hansen et al., 2010, Mangelson et al., 1997) indicates that the largest contribution to
42 PM_{2.5} on winter days that exceed the National Ambient Air Quality Standard comes from
43 secondary ammonium nitrate aerosol. However, the source apportionment and monitoring data
44 analyses presented in this work also demonstrates that secondary ammonium chloride aerosol is
45 a significant source of wintertime PM_{2.5} at all three monitoring locations. The presence of
46 significant concentrations of ammonium chloride aerosol in the Wasatch Front airshed has not
47 previously been identified. The study by Hansen et al. (2010) did identify elevated aerosol
48 chlorine in their winter 2007 measurements, but lacking cation measurements, they made the
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3 assumption that the chlorine was associated with aerosolized sodium chloride from entrained
4 road salt. In the work presented here, we determine that the majority of the chlorine is associated
5 with secondary ammonium chloride rather than primary sodium chloride.
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10 The physical and thermodynamic properties of ammonium chloride aerosol have been previously
11 investigated by Pio and Harrison (1987a). They found that the thermodynamics of aerosol
12 ammonium chloride are similar to that of ammonium nitrate. Both ammonium nitrate and
13 ammonium chloride aerosols exist in a reversible phase equilibrium with their gaseous
14 precursors (ammonia and the respective acids); however, the volatility of ammonium chloride
15 has a somewhat stronger temperature dependence than ammonium nitrate.
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23 While published observations of ambient ammonium chloride aerosol are unusual, a number of
24 previous studies have reported or inferred its presence. Du et al. (2010) measured chemically
25 speciated hourly PM_{2.5} concentrations in Shanghai for a two day period in 2009. They observed
26 that the hourly sum of ammonium sulfate, ammonium nitrate, and ammonium chloride ranged
27 from between 2.6 – 101.2 µg/m³ with ammonium nitrate being the dominant species. Hourly
28 ammonium nitrate and ammonium chloride observations were well correlated, with ammonium
29 chloride being approximately one tenth the concentration of ammonium nitrate. Chang and
30 Allen (2006) compared measured ammonium chloride aerosol with photochemical model
31 predictions during a photochemical episode in southeast Texas from August to September 2000.
32 Both model and measurements indicated that significant ammonium chloride aerosol formation
33 only occurred at times and in areas that were both ammonia rich and had sufficient atmospheric
34 chlorine. They found that anthropogenic chlorine emissions in southeast Texas had the potential
35 to enhance PM_{2.5} up to 9 µg/m³, but only in localized areas for brief periods given the
36 summertime temperatures and the mostly ammonia-poor conditions. Possanzini et al., (1992)
37 conducted field experiments in Rome, Italy and found that ammonium chloride aerosol
38 represented approximately one-fifth of observed ammonium nitrate and one-tenth of the total
39 ammonium species by mass. Pio and Harrison (1987b) determined ammonium chloride aerosol
40 concentrations of up to 10 µg/m³ in Northwest England, and Yoshizumi and Okita (1983)
41 examined aerosol filter data from 1975 collected in Riverside, CA and determined that
42 ammonium chloride concentrations ranged between 7.76 - 15.5 µg/m³.
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5 In this work we provide evidence that ammonium chloride aerosol adds a significant contribution
6 to PM_{2.5} in the Wasatch Front airshed during wintertime days when PM_{2.5} is elevated and the
7 airshed is more generally dominated by ammonium nitrate aerosol.
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10 11 MATERIAL AND METHODS

12 13 14 **PM_{2.5} samples**

15 The UDAQ operates three urban PM_{2.5} chemical speciation monitors as part of the U. S.
16 Environmental Protection Agency (EPA) Speciation and Trends Network (STN). The three STN
17 monitors are located in Salt Lake City, Bountiful, and Lindon, and Figure 1 shows a map
18 indicating the locations of the three monitors along the Wasatch Front. The 24-hour average
19 PM_{2.5} data used in this study were downloaded from EPA's Air Quality System (AQS) database
20 of quality-assured data. The date range of data extracted was 5/6/2007 – 5/9/2011 resulting in
21 429 sample days for Salt Lake City, 227 days for Bountiful, and 228 days for Lindon. The
22 sampling frequency at the Salt Lake City monitor was twice that of the other monitors, which
23 accounts for approximately twice the number of samples available from the Salt Lake City site.
24 The beginning of this date range coincided with the installation of an updated carbon sampler
25 (URG 3000N, May 2007), and the end date represents the latest data available at the time data
26 were extracted.
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38 After preprocessing the datasets (described in Section 2.2), the PMF analysis included the
39 following chemical species: Al, Br, Ca, Cr, Cu, Cl, Fe, Pb, Mn, Ni, Mg, Ti, V, Si, Zn, K, Na,
40 NH₄, NO₃, OC1, OC2, OC3, OC4, OP, EC1, EC2, EC3, SO₄ (note, OC and EC data were
41 exclusively from the URG 3000N). The Unmix analysis required removal of additional species.
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46 *Pre-processing*

47 Prior to analysis, the datasets were preprocessed to correct for field blanks, missing/negative
48 values, incomplete values, and poor signal-to-noise (S/N) ratios.
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53 Analytical data from EPA's STN monitoring network in the AQS database were not blank
54 corrected, but sample blank data were available. Field blanks were generally collected less
55 frequently than the monitoring frequency, so many samples did not have an associated field
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3 blank. For samples with coincident field blanks, the measured blank concentration was
4 subtracted from the reported concentration. For samples without coincident blanks, the median
5 value from the previous three blanks was used as an estimate for blank correction.
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10 The PMF model requires that each measurement be assigned an uncertainty. Most
11 measurements reported in the AQS database also report the analytical uncertainty. The overall
12 measurement uncertainty was estimated from a combination of the measured analytical
13 uncertainty and the method detection limit (MDL). The uncertainty of measurements reported
14 below the MDL were set to either $5/6 \cdot \text{MDL}$ or the reported uncertainty, whichever was larger.
15 The uncertainty of measurements above the MDL were set to the measured analytical uncertainty
16 plus $1/3 \cdot \text{MDL}$ (Reff et al., 2007). For those chemical species where uncertainties were not
17 available, the uncertainties were estimated based on Kim et al. (2005).
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26 If the total $\text{PM}_{2.5}$ mass of a sample was missing or if an entire analytical channel was missing
27 (e.g., carbon data, ions, metals), the sample was removed from the modeling dataset for both
28 PMF and Unmix datasets. In instances where individual chemical species were missing, the
29 missing observations in the PMF and the Unmix dataset were replaced with the species median
30 concentration, and the uncertainty for PMF was set to four times the species median
31 concentration to minimize the influence of the replaced data on the model solution. Chemical
32 species were removed from the modeling dataset if more than 50% of the samples had missing
33 data.
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42 PMF and Unmix do not allow negative data, which can sometimes be reported if species
43 concentrations are close to zero. For the PMF dataset, negative concentrations were reset to
44 zero. For the Unmix dataset, negative or zero values were replaced with $1/2 \cdot \text{MDL}$ (US EPA
45 2007).
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51 Data with low S/N can occur when many samples are reported near or below MDLs. Using data
52 with excessive noise has been found to negatively impact that quality of receptor modeling
53 results (Paatero and Hopke, 2003; Reff et al., 2007). In this study, chemical species with $\text{S/N} <$
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0.46 were discarded, and the influence of chemical species with S/N less than 2.0 was diminished in PMF by multiplying the uncertainties of these data by a factor of three.

To avoid overweighting the influence of duplicate chemical species in the model results, one of the duplicate species was removed to avoid double counting. In this analysis, S, Na⁺, K⁺, and OP by the thermal optical transmittance method were removed, and SO₄, Na, K, and OP by the thermal optical reflectance method were retained. The chemical species that were retained were chosen based on data completeness and S/N ratio.

In order to obtain a feasible Unmix solution with PM_{2.5} as the normalizing species, additional days were removed from the analysis based on plots of PM_{2.5} versus the sum of species (Figure S-1, supplementary material) and an evaluation of days affected by fireworks. The following days were removed from the Unmix analysis:

- Days when the PM_{2.5} concentration exceeded the sum of species by more than a factor of 2.
- Days when the PM_{2.5} sum of species exceeded the PM_{2.5} concentration by more than 20%. These days had low PM_{2.5} concentrations (less than 7 µg/m³).
- Days affected by fireworks. Preliminary Unmix runs were unable to identify a fireworks factor. Thus, the high potassium concentrations associated with fireworks would unnecessarily contribute to noise in this species concentration. Therefore, days were removed when K concentrations exceeded the average concentration by a factor of 5 or more and were within 2 days of New Year's Eve, the 4th of July, or Pioneer Day (July 24th).

A complete list of the days removed can be found in the supplementary material (Table S-1).

In addition, Unmix could not provide a feasible solution with all of the chemical species included in PMF. The subset of species for the Unmix analysis was selected by first identifying species with an average annual concentration of 0.2 µg/m³ and adding species that are suggestive of potentially relevant sources, such as Si for crustal material and K for biomass combustion. The final Unmix species were then selected by trial and error to yield the best solution. The selection criteria included species that led to a solution: having the best fit to the measured data (R² value)

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3 with PM_{2.5} set as the normalizing species, containing most of the species with annual an average
4 concentrations of 0.2 µg/m³ or greater, and including species that helped to identify sources of
5 interest.
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8 9 10 **PM_{2.5} emission inventory**

11 UDAQ provided the 2008 winter-adjusted and annual Salt Lake County, Utah County (Lindon
12 monitor location), and Davis County (Bountiful monitor location) PM_{2.5} emission inventories.
13 UDAQ is currently using these inventories to develop control scenarios as part of the Clean Air
14 Act requirement to develop a State Implementation Plan to bring the area into attainment of the
15 24-hour PM_{2.5} National Ambient Air Quality Standard. Consequently, this inventory has been
16 subject to high levels of review and quality assurance.
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24 In addition to daily emissions, the inventories contained categories for source type (point,
25 mobile, and area), county, source classification code, and a description of the sources. Based on
26 this description, the source types were grouped into five categories of primary PM_{2.5} sources:
27 gasoline, diesel, fugitive/ industrial/urban dust, and wood smoke/cooking. Mobile sources of
28 gasoline and diesel emissions were estimated using MOVES 2010A (on-road emissions) and
29 NONROAD 2008A (non-road emission). For Salt Lake County, diesel contributed 24% and
30 gasoline contributed 74% to mobile emissions with approximately 2% from other sources (i.e.,
31 aircraft and natural gas). The inventories did not identify any stationary point sources of diesel or
32 gasoline emissions.
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42 The emission inventories included the following fugitive dust sources: paved and unpaved roads,
43 mining/mineral processing operations, sand and gravel operations, and agricultural operations.
44 They did not include wind-blown dust that is associated with undisturbed land, land outside the
45 county, or other activities. The emission factors for fugitive dust estimates were developed from
46 population census numbers, EPA emission factors, and UDAQ-developed, state-specific
47 emission factors. The inventories also included emissions from wood stoves, fireplaces and
48 cooking (i.e., commercial char broiling and frying). These emissions were also based on
49 population census numbers and EPA emission factors. The inventories did not include emissions
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3 from prescribed burns, agricultural burns, wildfires, or secondary formation of PM_{2.5} from the
4 other products of wood burning, i.e., volatile organic compounds.
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9 The inventories do not include fireworks or estimates of secondary sources of PM_{2.5}. In addition,
10 not all categories in the inventory match the five primary source categories from the source
11 attributions, which are discussed in the results section. Emission inventory source categories not
12 matched to the source apportionment results include emissions from the use of natural gas,
13 propane, and unspecified fuels, and they accounted for 11% of the Salt Lake County Inventory,
14 15% of the Utah County Inventory (Lindon), and 25% of the Davis County Inventory
15 (Bountiful).
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22 Theory/calculation

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24 Both PMF and Unmix solve the general receptor modeling problem, given by
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$$28 \quad C_{ij} = \sum_{k=1}^m a_{jk} S_{ik} + e_{ij} \quad i = 1, \dots, N; j = 1, \dots, n \quad (1)$$

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30 For airborne particles C_{ij} is the j^{th} species concentration ($\mu\text{g}/\text{m}^3$) measured in the i^{th} sample, S_{ik} is
31 the mass contribution from source k in the i^{th} sample (e.g., source contribution, $\mu\text{g}/\text{m}^3$), and e is
32 the error, which results from variations in the source contribution and analytical uncertainties.
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38 *PMF*

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40 PMF source apportionment modeling was performed using EPA PMF 3.0 (Norris et al., 2008).
41 A thorough discussion of the mathematical equations underlying EPA PMF can be found in
42 Paatero (1997), Paatero (1999), Paatero and Hopke (2003), Reff et al. (2007), and Norris et al.
43 (2008). The model was run in the robust mode using a non-random seed value of 10 (so that
44 results were reproducible) and 20 repeat runs to ensure the model least-squares solution
45 represented a global rather than local minimum. The rotational FPEAK variable was held at the
46 default value of 0.0. The model solution with the optimum number of factors was determined
47 somewhat subjectively based on inspection of the factors in each solution, but also from the
48 quality of the least-squares fit (analysis of Q_{Robust} and Q_{True} values) in the model output. The
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3 scaled residuals for final model solutions were normally distributed and generally fell into the
4 recommended range of +3 to -3.
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7 8 *Unmix*

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10 The U.S. EPA's Unmix model 6.0 (US EPA, 2007) was used in this study. The mathematical
11 details of Unmix can be found in Henry (1997, 2002, 2003). Unmix solves the receptor modeling
12 problem (eqn. 1) using a self-modeling curve resolution algorithm that searches for "edges" in
13 the data that define the j^{th} species mass contribution from the k^{th} source. These edges occur when
14 some samples lack contributions from at least one source or are dominated by contributions from
15 one source for a group of species. The number and direction of the edges derived from Unmix
16 depend on the set of species used, and the solution is sensitive to the species included in the
17 model. It has been successfully employed in various source attribution studies and generates
18 results that compare well with other techniques (Hu et al., 2006; Chakraborty and Gupta, 2010;
19 Mukerjee et al., 2004; Song et al., 2006).
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28 29 **RESULTS AND DISCUSSION**

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31 This section primarily focuses on the results for Salt Lake City with a brief discussion of the
32 Lindon and Bountiful locations. Additional detail on the factor profiles for Bountiful and Lindon
33 can be found in the supplementary material (Figure S-2 and S-3). The reader should note that
34 factors determined in a source-attribution analysis using factor analysis methods like PMF and
35 Unmix are not necessarily mutually exclusive. For example, a factor identified as predominantly
36 ammonium nitrate aerosol is likely from a combination of primary and secondary sources
37 including gasoline engines, diesel engines, and industrial facilities, even though these sources
38 may also have separately identified factors. Hence, caution should be used in interpreting factor
39 classifications too literally or with exclusivity.
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49 **PMF Results**

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51 Figure 2 illustrates the species contributions to the PMF factors for Salt Lake City, and the
52 following sections detail each of the PMF factors. Figure S-4 shows the monthly PMF factor
53 contributions for Salt Lake. The PMF factors account for 98% of the PM_{2.5} mass in Salt Lake
54 City and Lindon, and 97% of the PM_{2.5} mass in Bountiful.
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Ammonium Chloride Factor

A PMF factor predominantly comprising ammonium chloride was identified at all three STN monitoring stations along the Wasatch Front. During winter PM_{2.5} pollution episodes this factor was significant contributing 10%–15% of total PM_{2.5} on days when the 24-hour PM_{2.5} exceeded 30 µg/m³.

Independent of the PMF results, we performed an ion charge balance analysis on each STN dataset that confirmed the presence of ammonium chloride aerosol. This analysis showed that the total measured cationic charge in each sample (sum of charge from Na⁺, K⁺, NH₄⁺) was closely balanced by the total measured anionic charge (sum of charge from Cl⁻, NO₃⁻, and SO₄²⁻), indicating there were no major missing ionic species (Figure S-5). Note that the chloride ion was not directly measured, but total chlorine was assumed to be mostly chloride ion. The ion balance on days with high chlorine mass indicated that the overwhelming majority of chlorine was associated with ammonium due to very low concentrations of both Na⁺ and K⁺ (see Figures S-6 –S-8 in supplementary materials).

Chang and Allen (2006) describe the equilibrium partitioning between the gas and particle phases when ammonia, nitric acid, sulfuric acid and chloride are all present in an airshed. Ammonia-poor conditions occur when there is insufficient ammonia to neutralize all acid components, and under these conditions equilibrium partitioning will favor the formation of ammonium sulfate over other ammonium salts. Ammonia-rich conditions occur when there is excess ammonia, and there is sufficient ammonia to neutralize all of the acids in the aerosol phase. Several studies have demonstrated that the airshed in and near Salt Lake City is ammonia-rich during winter PM_{2.5} pollution episodes (UDAQ, 2011; Martin, 2006; Mangelson et al., 1997). Also, the charge balance between aerosol anions and cations described above indicates that the majority of aerosol acids were neutralized by ammonium.

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3 As previously noted, the work of Pio and Harrison (1987a) reported that the thermodynamics of
4 ammonium chloride aerosol are similar to that of ammonium nitrate. They also estimated that
5 under typical atmospheric concentrations of gaseous ammonia and hydrogen chloride,
6 ammonium chloride aerosol should not be prevalent at temperatures above 10°C. The Wasatch
7 Front results reported here are consistent with those estimates. In our results the majority of
8 ammonium chloride factor mass above 1 $\mu\text{g}/\text{m}^3$ occurred when the maximum daily temperature
9 was below 10°C, and all instances of this factor above 3 $\mu\text{g}/\text{m}^3$ occur below this temperature
10 (Figure S-9).
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19 Possible sources for chlorine include wintertime road salting, industrial emissions, and biological
20 activity linked to the Great Salt Lake and surrounding marshes. Of the possible sources, road
21 salting was deemed an unlikely source of chlorine given the relative absence of sodium in the
22 observations as well as the absence of reports of ammonium chloride aerosol in other urban areas
23 that conduct road salting in wintertime. Sodium chloride is, however, a well-known source of
24 gas-phase chlorine in coastal and marine areas through the reaction of nitric acid with sodium
25 chloride and the release of hydrogen chloride. Because measured levels of ammonia in the
26 atmosphere during wintertime along the Wasatch Front are high (UDAQ, 2011), nitric acid levels
27 should be extremely low, making the reaction of nitric acid with sodium chloride unlikely.
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37 To address possible sources of chlorine from industrial activity, regional emissions inventories
38 were evaluated and indicated one significant industrial source of chlorine in the Wasatch Front
39 airshed. US Magnesium, Inc., a magnesium production facility that lies approximately 70 km
40 due west of Salt Lake City. The 2008 emissions estimates for US Magnesium, Inc. compiled by
41 the State of Utah for their PM_{2.5} State Implementation Plan, reported 579 tons/year of chlorine
42 air emissions, which made up 96.6% of the total reported anthropogenic chlorine emissions for
43 the airshed. At this reported emission level, screening-level plume dispersion modeling using the
44 NOAA ARL Hysplit on-line modeling system (Draxler and Rolph, 2012) indicates that chlorine
45 emissions from US Magnesium might be a significant contributor to aerosol chlorine observed at
46 the three STN PM_{2.5} samplers along the Wasatch Front. While more refined plume dispersion
47 modeling is beyond the scope of this study, these screening-level results suggest that a more
48 refined modeling investigation is warranted.
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5 A third possible source for the atmospheric chlorine is biological activity related to the Great Salt
6 Lake. Methyl chloride has been found to be a product of microbial activity in salt marshes and
7 the ocean. Rhew et al. (2000) reported significant fluxes of bromomethane and chloromethane in
8 two coastal salt marshes and suggested that salt marshes may contribute up to 10% of the global
9 emissions budget for these methyl halides. While chlorine emissions from salt marshes
10 associated with the Great Salt Lake is an intriguing hypothesis for a contributing source of
11 atmospheric chlorine in the Salt Lake airshed, more investigation is necessary to quantify all the
12 contributing sources of chlorine..
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19 20 *Ammonium Nitrate Factor* 21

22 This factor was identified by the dominance of nitrate and ammonium. Ammonium nitrate
23 accounts for a significant portion of the average winter $PM_{2.5}$, approximately 50%, and
24 contributions were as high as 75% on some highly polluted winter days. Figure S-4 indicates that
25 both ammonium nitrate mass and mass fraction were elevated during the winter months and
26 mostly absent during summer. UDAQ has performed monitoring to determine whether ammonia
27 is a limiting factor in the formation of ammonium nitrate and other inorganic secondary PM and
28 has determined that ammonia is present in excess during wintertime, cold-air pool events along
29 the Wasatch Front (UDAQ, 2011).
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37 38 *Ammonium Sulfate* 39

40 This factor was identified by the dominance of sulfate and ammonium. Figure S-4 shows the
41 average monthly mass impacts from this factor and indicates no strong seasonal pattern. This
42 suggests that primary sulfate emissions may be playing a significant role in this factor, rather
43 than secondary photochemical production where one would expect a pattern of summertime
44 maxima and winter minima.
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Diesel and Gasoline Factors

Both gasoline and diesel factors were identified by the relative contribution of OC and EC to the total mass. For gasoline vehicles, the factor was dominated by OC2, OC3 and OC4, with the EC component mostly EC1. Previous work has also linked similar factors to gasoline vehicles (Hwang and Hopke, 2007; Kim et al., 2004; Maykut et al., 2003; Zhao and Hopke, 2004, 2006; Kim and Hopke, 2006; Kim and Hopke, 2008b). For diesel emissions, the relatively high proportions of Mn and Fe along with OC and EC are similar to findings in other major urban areas (Kim et al., 2004; Maykut et al., 2003; Ramadan et al., 2000; Kim and Hopke, 2008a; Wu et al., 2007; Amato and Hopke 2011; Kim et al., 2010; Lewis et al., 2003). The diesel factor showed a statistically significant higher average mean contribution on weekdays vs. weekend (student t-test 90% confidence level).

The monthly impact (Figure S-4) shows that both factors have highest impacts during the cold months of the year: November, December, and January. However, these impacts are likely due to the higher frequency of air-stagnation events and stable atmospheric boundary layers in winter and are not associated with an actual increase in gasoline and diesel emissions in those months. The monthly average fractional contribution of the diesel factor to $PM_{2.5}$ (Figure S-4) fluctuated around 10% throughout the year. The monthly average fractional contribution of the gasoline factor to $PM_{2.5}$ ranged between 10-20%.

Fugitive Dust I and II Factors

Two factors possessing markers associated with airborne crustal materials were identified as Fugitive Dust I and II. Both contained a large fraction of Al, Si, Mg, Ca, Fe, and Ti. Fugitive Dust II was enriched in Ca compared to Fugitive Dust I. While there are similarities in Fugitive Dust I and Fugitive Dust II chemical signatures, Fugitive Dust I and II factors tended to be present in different degrees depending on the time of year. Fugitive Dust I tended to be most active in July – September, while Fugitive Dust II was mostly active during March through July (Figure S-4). Superimposing the wind-speed data on the time series graphs of the fugitive dust factors shows a moderate correlation between elevated wind speeds and the magnitudes of the factor values. For example, when the average hourly wind speed was 5.4 m/s (12 mph) or greater, the average fugitive dust contribution was significantly greater than on days with an average hourly wind speed below 5.4 m/s (student t-test 95% confidence level). It is worthy to note, however, that the exceptionally high wind speeds recorded during winter months resulted in no increase in elevated fugitive dust factor. This is consistent with the understanding that snow cover and high soil-moisture levels prevent significant entrainment of crustal material.

The presence of two separate factors associated with the entrainment of fugitive dust material is most likely associated with the seasonal change in the wind pattern. Similar to the results presented in this work, Chen et al. (2010) reported finding two separate dust factors, one calcium-rich and one not, in PMF modeling for receptors in southeastern Minnesota. In their work they attributed potential sources of calcium-rich aerosol to limestone outcroppings in the karst areas of southeastern Minnesota and/or concrete dust from construction activities. In a separate study, PMF modeling conducted by Desert Research Institute (DRI) as part of the Causes of Haze Assessment project (<http://www.coha.dri.edu/>) found two separate soil factors, also one Ca-rich and one not, for monitoring sites at Nevada's Great Basin National Park and Jarbidge Wilderness Area and Utah's Lone Peak Wilderness Area. The findings of two soil factors at the three STN monitors in the greater Salt Lake area in this study as well as DRI's results for three remote locations in Utah and Nevada suggest these two crustal factors represent different soil types, possibly from differing source regions.

Fireworks Factor

The chemical fingerprint of the fireworks factor is composed of metals commonly used in pyrotechnics (Vecchi et al., 2008; Joly et al., 2010). The time series of PM_{2.5} associated with this factor also matched that expected for fireworks use. The largest contributions from this factor fell on sampling days on or near the national holidays of New Years Day (Jan 1st), Independence Day (July 4th), and the state holiday of Pioneer Day (July 24th).

Industrial/Urban Factor

This factor is dominated by a mixture of OC, EC, SO₄ and NO₃, with trace but significant contributions from Na, V, Ni, Pb and Cr. The EC2 fraction was particularly important in the model's identification of this factor. Assigning the source of this factor is unclear. Previous work has linked similar factors to a variety of sources including railroad traffic or other combustion-related emissions (Kim et al., 2004; Zhao and Hopke, 2004; Kim and Hopke, 2006; Han et al., 2007). It is likely that this factor is in some way connected to fuel combustion. This factor's time series of monthly average impacts shows a bimodal shape with the two maxima in March and October.

OP-Rich Factor

This factor was identified by the dominance of the OP species. Numerous other PMF studies that have used OC and EC fraction data based on the thermal evolution protocols developed under the IMPROVE program have also found OP-rich factors (Hwang and Hopke, 2007; Zhao and Hopke, 2006; Kim and Hopke, 2006) and have suggested this factor is linked to secondary organic aerosol production (Jang et al., 2003).

Wood-Smoke Factor

The wood-smoke factor was identified by its large OC content, smaller EC content, and majority of mass impacts occurring during the winter months of December and January. It is interesting to note that K, commonly associated with combustion of biomaterials, was not identified as a part of this factor. It is possible that potassium's role in other factors (e.g., fugitive dust, fireworks) combined with the relatively small contribution of wood smoke in the overall dataset influenced the models lack of attribution of potassium to the wood smoke factor.

Unmix Results

For Salt Lake City, the Unmix solution included 411 observations and 13 species: Fe, Ni, Si, Zn, K, NH₄, NO₃, OC1, OC2, OC3, OP, EC1, and PM_{2.5}. This solution had a minimum R² of 0.92, indicating that at least 92% of the variance of each species can be explained by five sources. The Unmix factors accounted for 99% of the PM_{2.5} mass in Salt Lake City, 98% of the mass in Lindon, and 97% of the mass in Bountiful. Figure 3 summarizes the species contributions to each of the five factors in Salt Lake City; Figures S-10 and S-11 summarize the species contributions for Bountiful and Lindon, respectively. The standardized residuals were normally distributed and generally fell into the recommended range of +3 to -3. Note that chlorine was not included in the Unmix analysis because either Unmix could not find a feasible solution or other important species, like nitrate or several of the OC fractions, had to be omitted in order to obtain a feasible solution. Thus, Unmix could not identify an ammonium chloride factor for any of the Wasatch Front locations.

Diesel Factor

This factor comprises EC and OC with trace elements of Fe, Zn and/or Ca, depending on the species included in the analysis (Figure 3, S-10, S-11). The EC/OC ratio ranged from 0.67 to 1.1 at the three locations. This factor was slightly more prevalent during winter. However, the weekday contributions from this factor were significantly higher than the weekend contributions (90% confidence interval, unpaired t-test, equal variance). Furthermore when examining the most important days for this factor, none of the top-20 days occurred on a Sunday, suggesting that this factor is related to vehicle traffic.

Fugitive Dust Factor

This factor comprises Si, small amounts of K, Fe, and Al, and some EC and OC. Because the best solution for each location required a slightly different mix of species, the factor profile for fugitive dust differed slightly at the three monitoring stations. Days when this factor was most important were somewhat associated with higher wind speeds. When the average hourly windspeed was 5.4 m/s (12 mph) or greater, the average fugitive dust contribution was significantly greater than on days with an average hourly windspeed below 5.4 m/s (student t-test 95% confidence level). Based on the factor composition and the seasonal pattern, showing low contributions during the winter, this factor appears to be mostly wind-driven fugitive dust from natural sources, paved roads, and potentially industrial sources. The composition and the seasonal pattern of this factor agree with the PMF factors for Fugitive Dust I and Fugitive Dust II.

Gasoline Factor

This factor contained EC and OC with trace amounts of Zn (Figure 3). It contained less Fe and Ca compared to the diesel factor, and it had a lower EC/OC ratio, ranging from 0.18 to 0.65. It was more prevalent during winter and exhibited a weekday/weekend trend. Although the differences were not statistically significant, when examining the days when this factor is most important, none of the top 20 days included Sundays, suggesting that this factor has some relationship to traffic.

Secondary Inorganic Aerosol Factor. Based on the factor composition, with significant contributions of ammonium and nitrate (Figure 3), and the strong seasonal pattern (much more common in winter, Figure S-12) this factor appears to represent secondary inorganic pollutant formation, primarily ammonium nitrate. Both the composition and the seasonal contribution of this factor agree with the PMF results.

Wood Smoke/Aged Wood Smoke. The importance of K, EC, OC and OP (Figure 3) as well as the seasonal pattern suggest that this factor is related to wood smoke. The potassium contribution to this factor was 1-2%, and the EC contribution was 14-20%. However, this factor shows an OP contribution of 6-10%, which does not agree with the PMF results. It is possible that this factor

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3 represents wood smoke and a portion of the PMF OP-rich factor, which may include aged wood
4 smoke. OP can be a marker for wood smoke and biomass burning; Chow et al. (2004) report OP
5 factors for a variety of wood and vegetative burns range from 2 – 25%. Jeong et al. (2011)
6 identified a biomass-burning factor based on high concentrations of K, OP and EC1. Chow et al.
7 also suggest that the K/Fe ratio can be used as a tracer for wood smoke, with a range of 3 – 230.
8 In Salt Lake City, the ratio of K/Fe is 7, which is at least a factor of 10 greater than the other four
9 factors. In Lindon, the ratio is higher than any of the other four factors but was difficult to
10 quantify because the Fe contribution was very close to zero.
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19 The seasonal pattern was also important in identifying this factor (Figure S-12). This factor
20 tended to be important on days associated with wildfires. The contribution of this factor was
21 also higher in the summers of 2007 and 2008 when high wildfire activity occurred in Utah and
22 California. In the summer of 2010, wildfire activity was much lower in this region as was the
23 contributions from this factor. The species found in this factor may also be indicative of wood
24 burning for wintertime heating, cooking, charbroiling/smoking, and frying foods (Chow et al.,
25 2004).
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33 **Comparison of PMF, Unmix, and the Inventory Results**

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36 The source attribution results were in close agreement with respect to the large contributions of
37 inorganic secondary PM to PM_{2.5} along the Wasatch Front. During winter when PM_{2.5}
38 concentrations exceeded 20 µg/m³ the PMF and Unmix results showed that secondary PM was
39 the largest contributor to PM_{2.5} (60 – 67% for Salt Lake City, 72 -73% for Lindon, and 64 – 80%
40 for Bountiful). Examining the year-round results also showed that secondary PM was important
41 with contributions ranging from 41–43% in Salt Lake City, 41–44% in Lindon, and 40–45% in
42 Bountiful.
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50 Figures 4 and 5 compare the source contributions of primary PM_{2.5} obtained with PMF, Unmix,
51 and the county emission inventories during winter when PM_{2.5} concentrations exceeded 20 µg/m³
52 and for the year-round results, respectively. Note that Salt Lake, Bountiful, and Lindon are
53 located in different counties. The county inventories included only primary PM_{2.5} emissions and
54 did not attempt to estimate secondary PM_{2.5}. Consequently in order to compare the source-
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3 attribution results to the inventories, the source-attribution results were normalized to remove the
4 secondary PM_{2.5} factors (including the PMF factors ammonium chloride, ammonium nitrate, and
5 ammonium sulfate, and the secondary Unmix factor). Inventory sources that could not be
6 categorized into one of the five primary PM_{2.5} categories were grouped into the category “other”.
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8 The PMF firework factor was grouped in the “other” category.
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14 Although the Unmix solution could not resolve as many factors as the PMF solution, once the
15 PMF factors are grouped, the winter-time Salt Lake City results agreed well with each other,
16 with the exception of the contribution from wood smoke. Both PMF and Unmix suggested a
17 greater contribution from wood smoke and OP than the inventory. For the purpose of this
18 analysis, aged wood smoke is considered to be the main contributor to the wintertime OP-rich
19 PMF factor mass. So the wood smoke contribution here is the sum of the PMF wood smoke and
20 OP-rich factors in order to make the PMF results comparable to the Unmix results. The absence
21 of any secondary PM contributions from wood smoke in the inventory partially, but not fully,
22 explains the discrepancy. It is possible that cooking could be related to the wood-smoke factors.
23 However even if one sums the wood smoke emissions and the cooking emissions from the
24 inventory, the inventory accounts for less than half of the PMF- or Unmix-predicted wood-
25 smoke and OP contribution. It is also possible that the location of the Salt Lake City monitor in
26 an older residential area is more influenced by wood burning than the county as a whole. As Salt
27 Lake County seeks strategies for meeting the PM_{2.5} national ambient air quality standard, in
28 addition to strategies to control secondary PM_{2.5}, additional restrictions on residential wood
29 burning may be considered because the monitoring station Salt Lake City used in this study tends
30 to exhibit the highest PM_{2.5} concentrations in Salt Lake County during winter-time, cold-pool
31 conditions. When comparing the year-round results (Figure 5), both PMF and Unmix identified
32 a greater contribution from fugitive dust than the inventory, which is understandable because the
33 inventory does not include wind-blown dust. The inventory also predicts greater contributions
34 from diesel and gasoline vehicles than the source-attribution methods.
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52 **Comparison of Primary PM_{2.5} Results for Lindon and Bountiful**

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55 Unmix and PMF identified similar factors in Salt Lake, Bountiful, and Lindon; however, neither
56 PMF nor Unmix identified primary PM_{2.5} gasoline exhaust contributions in Bountiful. Note that
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3 Unmix was not able to resolve as many factors as PMF.
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7 *Lindon*

8 During winter when $PM_{2.5}$ concentrations exceeded $20 \mu\text{g}/\text{m}^3$, the Unmix, PMF, and inventory
9 results for primary $PM_{2.5}$ agreed reasonably well with contributions from fugitive dust of
10 approximately 11% of primary $PM_{2.5}$, from diesel emissions ranging from 18–30%, and from
11 gasoline ranging from 31–41%. In addition, PMF results suggest larger contributions from wood
12 smoke (40% of primary $PM_{2.5}$) compared to the inventory (21%) or Unmix 19%. When
13 considering the year-round emissions, PMF and Unmix suggest greater contributions from
14 fugitive dust compared to the inventory, likely the result of wind-blown dust not being included
15 in the inventory.
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25 *Bountiful*

26 The Bountiful results for primary $PM_{2.5}$ (PMF, Unmix, and inventory) do not agree as well as
27 either the Salt Lake City or Lindon locations. During winter when $PM_{2.5}$ concentrations
28 exceeded $20 \mu\text{g}/\text{m}^3$, the contributions ranged from 6–40% of the primary portion of $PM_{2.5}$ for
29 fugitive dust, 2–20% for diesel, 0–51% for gasoline, and 20–70% for wood smoke. It is unclear
30 whether the lack of agreement is due to the inability to include OP among the species in the
31 Unmix solution, or the need to remove more dates to obtain a feasible solution with $PM_{2.5}$ as the
32 normalizing species. Neither Unmix nor PMF could identify a primary gasoline factor at this
33 location, although the inventory identifies 50% and 38% of contributions from gasoline sources
34 during the winter and year-round, respectively. The source attribution methods suggest a greater
35 contribution from fugitive dust sources, ranging from double to greater than a factor of six (year
36 round).
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47 **Conclusions**

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49 The two factor analysis techniques, PMF and Unmix, were used to identify contributors to $PM_{2.5}$
50 at three monitoring stations along Utah's Wasatch Front: Bountiful, Lindon, and Salt Lake City.
51 During wintertime, cold-pool events, both source attribution methods identified inorganic
52 secondary aerosols as the main contributors to $PM_{2.5}$ concentrations (60–80%, all locations), with
53 secondary ammonium nitrate being the dominant contributor. In addition to ammonium nitrate,
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3 PMF identified a separate factor linked to secondary ammonium chloride, and this factor had a
4 minor but significant contribution to the total PM_{2.5} aerosol mass during wintertime PM_{2.5}
5 pollution events along the Wasatch Front (10%–15% of total PM_{2.5} on days when the 24-hour
6 PM_{2.5} exceeded 30 µg/m³). Subsequent ion balance analysis of the monitoring data confirmed
7 the presence of significant ammonium chloride aerosol on these highly polluted days at all three
8 monitoring sites. The presence of significant ammonium chloride aerosol is unusual and has not
9 previously been reported for this airshed. Sources of chlorine in the airshed are not well
10 understood, but could be related to industrial activity and/or biogenic emissions from the Salt
11 Lake.
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21 Although Unmix identified fewer factors than PMF, after grouping some of the PMF factors, the
22 source attribution results for the primary portion of PM_{2.5} generally agreed with the emissions
23 inventories for the region. The source attributions for primary PM_{2.5} suggest that wood smoke
24 and cooking may be as important as gasoline emissions during the winter-time, cold-pool events
25 in Salt Lake City, perhaps due to the location of the sampling stations in residential areas.
26 During spring through fall, fugitive dust emissions may be more important contributors to
27 primary PM_{2.5} emissions than the inventory suggests. This may be due to the omission of wind-
28 blown dust in the inventories and the difficulty in obtaining accurate estimates of fugitive dust
29 emissions in an arid region when emissions are highly dependent on meteorological conditions
30 as well as activity levels.
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11

12 **About the Authors**

13 Kerry Kelly is a Research Associate with the Department of Chemical Engineering and Institute
14 for Clean and Secure Energy as well as a PhD student in Environmental Engineering at the
15 University of Utah.
16

17 Robert Kotchenruther is an Environmental Scientist with U.S. EPA Region-10.
18

19 Roman Kuprov is an Environmental Scientist with the Utah Division of Air Quality.
20

21 Geoffrey D. Silcox is a lecturing professor and associate chair of Chemical Engineering at the
22 University of Utah.
23

24 **Keywords**

25 Source attribution, PM_{2.5}, cold pool, nonattainment, inversion
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32 inventory and guidance on its interpretation. Many thanks to Dr. Cristina Jaramillo, University
33 of Utah Department of Chemical Engineering, for her insights on factor analysis.
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37 **Figure captions**

38 Figure 1. Locations of the three sampling stations along the Wasatch Front.
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41 Figure 2. Species contributions to PMF factor profiles for Salt Lake City.
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44 Figure 3. Species contributions to Unmix factor profiles for Salt Lake City.
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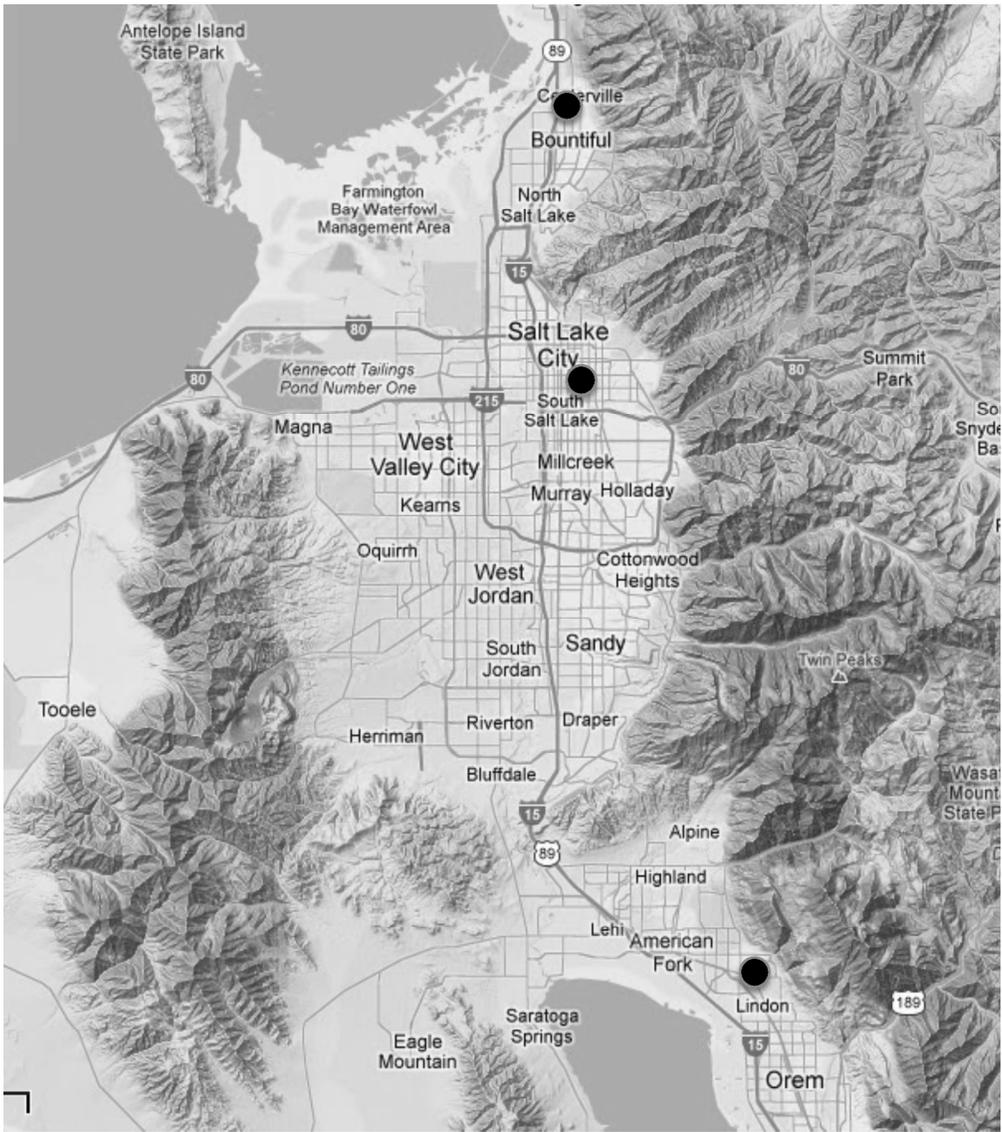
46 Figure 4. Comparison of primary PM_{2.5} factor contributions (%) from PMF, Unmix, and the
47 UDAQ inventory, during winter when the average contributions to PM_{2.5} on days with PM_{2.5} >
48 20 µg/m³ (October – March). Because the inventory does not include secondary PM_{2.5}, for this
49 comparison the PMF and Unmix results were normalized to remove the secondary PM_{2.5}
50 contributions. This inventory represents a typical winter day, i.e., January or February (Pennell,
51 personal communication). Note that cooking includes frying and charbroiling.
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53 Figure 5. Comparison of primary PM_{2.5} year-round factor contributions (%) from PMF, Unmix,
54 and the UDAQ's emission inventory for PM_{2.5}. Because the inventory does not include
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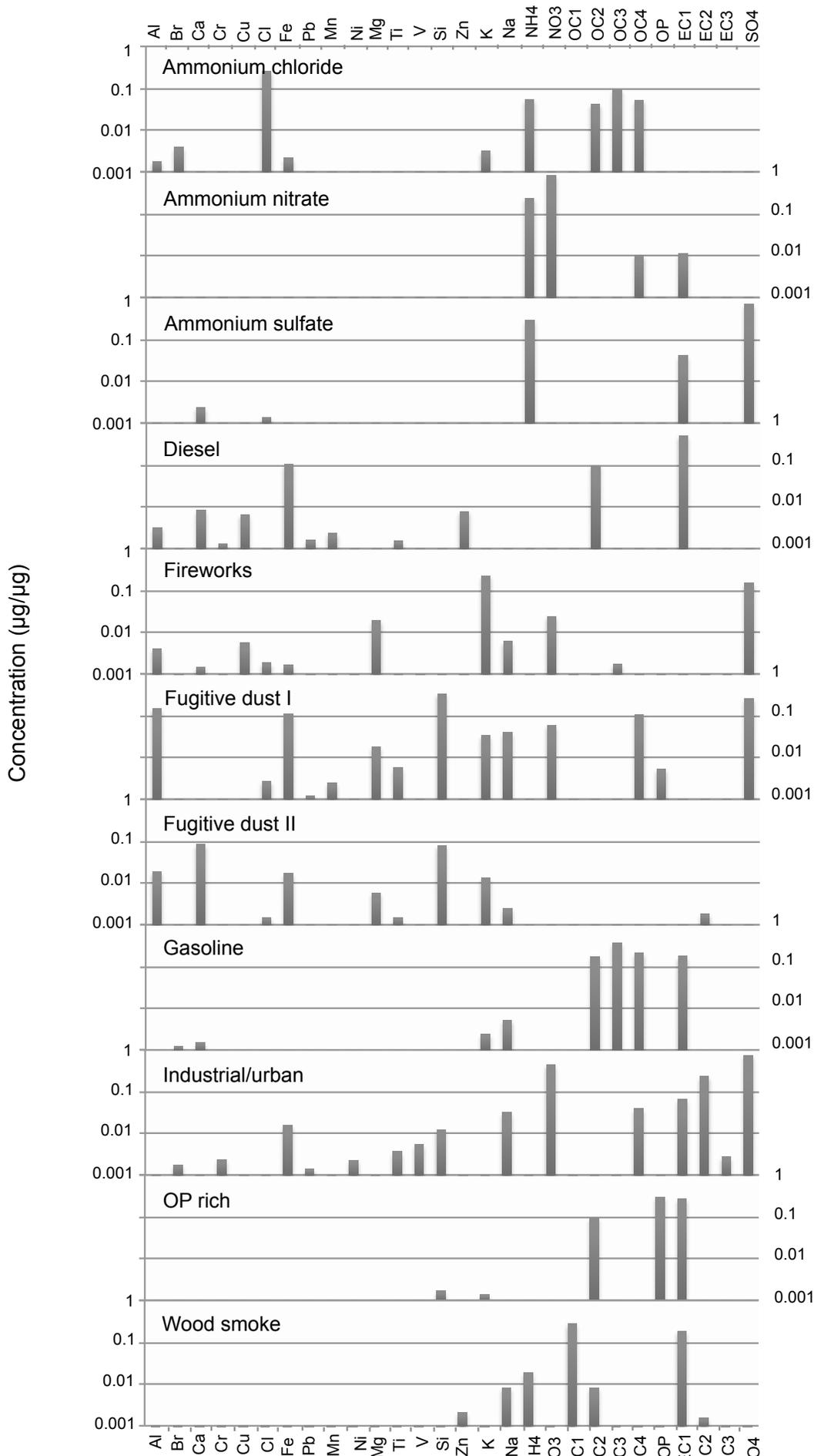
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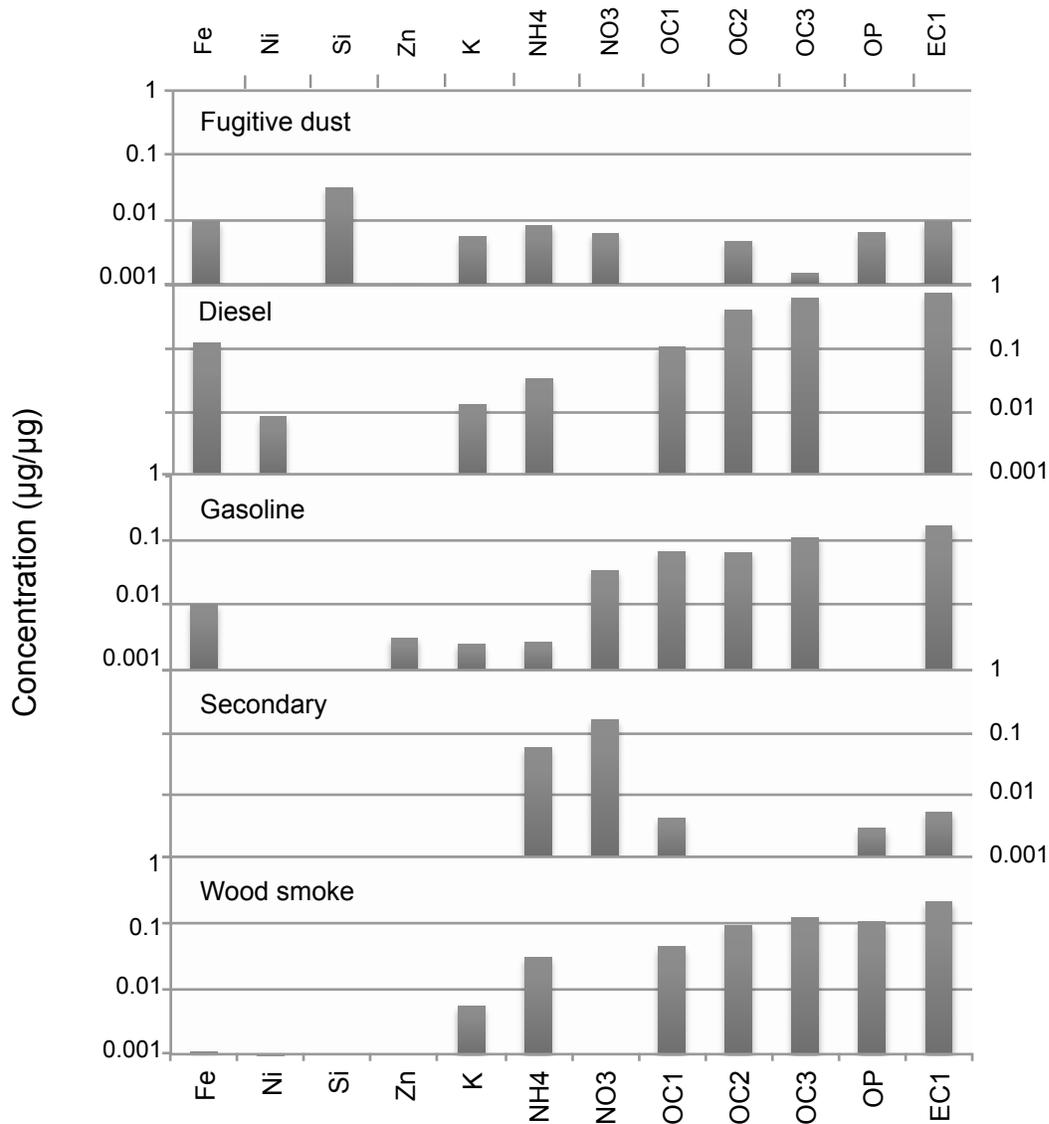
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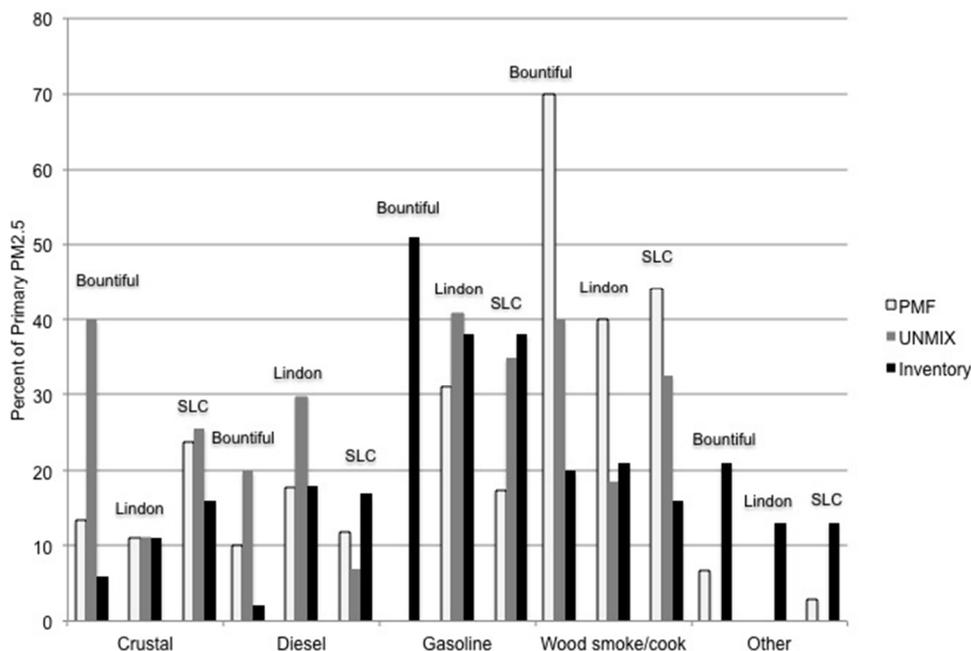


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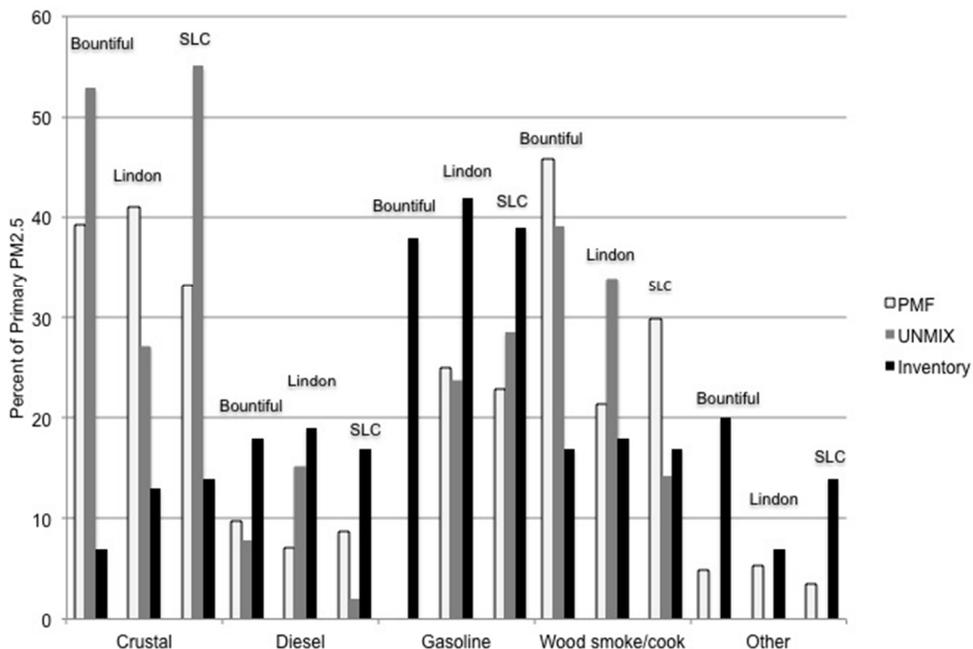






Comparison of primary PM_{2.5} factor contributions (%) from PMF, Unmix, and the UDAQ inventory, during winter when the average contributions to PM_{2.5} on days with PM_{2.5} > 20 µg/m³ (October – March). Because the inventory does not include secondary PM_{2.5}, for this comparison the PMF and Unmix results were normalized to remove the secondary PM_{2.5} contributions. This inventory represents a typical winter day, i.e., January or February (Pennell, personal communication). Note that cooking includes frying and charbroiling.

238x161mm (72 x 72 DPI)



Comparison of primary PM2.5 year-round factor contributions (%) from PMF, Unmix, and the UDAQ's emission inventory for PM2.5. Because the inventory does not include secondary PM2.5, for this comparison the PMF and Unmix results were normalized to remove the secondary PM2.5 contributions.
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Supplementary Material

Figure S-1. Sum of species vs. $PM_{2.5}$ concentration for the Salt Lake City monitoring site.

Figure S-2. Species contributions to PMF factor profiles for Bountiful.

Figure S-3. Species contributions to PMF factor profiles for Lindon.

Figure S-4. Monthly PMF factor contributions for Salt Lake City. The box indicates the 25th and 75th percentiles, and the line within the box indicates the mean value. The lines indicate the maximum and minimum values, excluding outliers, which are denoted by an 'x' and exceed the interquartile range by more than a factor of 1.5.

Figure S-5. Aerosol charge balance between anions and cations at three STN monitoring locations along Utah's Wasatch Front. Anion equivalence was calculated as the sum $[Cl^-]/35.453 + [NO_3^-]/62.005 + [SO_4^{2-}]/48.03$, cation equivalence was calculated as the sum $[Na^+]/23.0 + [K^+]/39.098 + [NH_4^+]/18.04$.

Figure S-6. Cation and anion charge equivalence for the highest measured aerosol Cl days in Salt lake City.

Figure S-7. Cation and anion charge equivalence for the highest measured aerosol Cl days in Bountiful.

Figure S-8. Cation and anion charge equivalence for the highest measured aerosol Cl days in Lindon.

Figure S-9. PMF factor for ammonium chloride plotted against maximum daily temperature at three STN monitoring locations along Utah's Wasatch Front.

Figure S-10. Species contributions to Unmix factor profiles for Bountiful.

Figure S-11. Species contributions to Unmix factor profiles for Lindon.

Figure S-12. Monthly Unmix factor contributions for Salt Lake City. The box indicates the 25th and 75th percentiles, and the line within the box indicates the mean value. The lines indicate the maximum and minimum values, excluding outliers, which are denoted by an 'x' and exceed the interquartile range by more than a factor of 1.5.

Table S-1. Days removed from the Unmix analysis for Salt Lake City.

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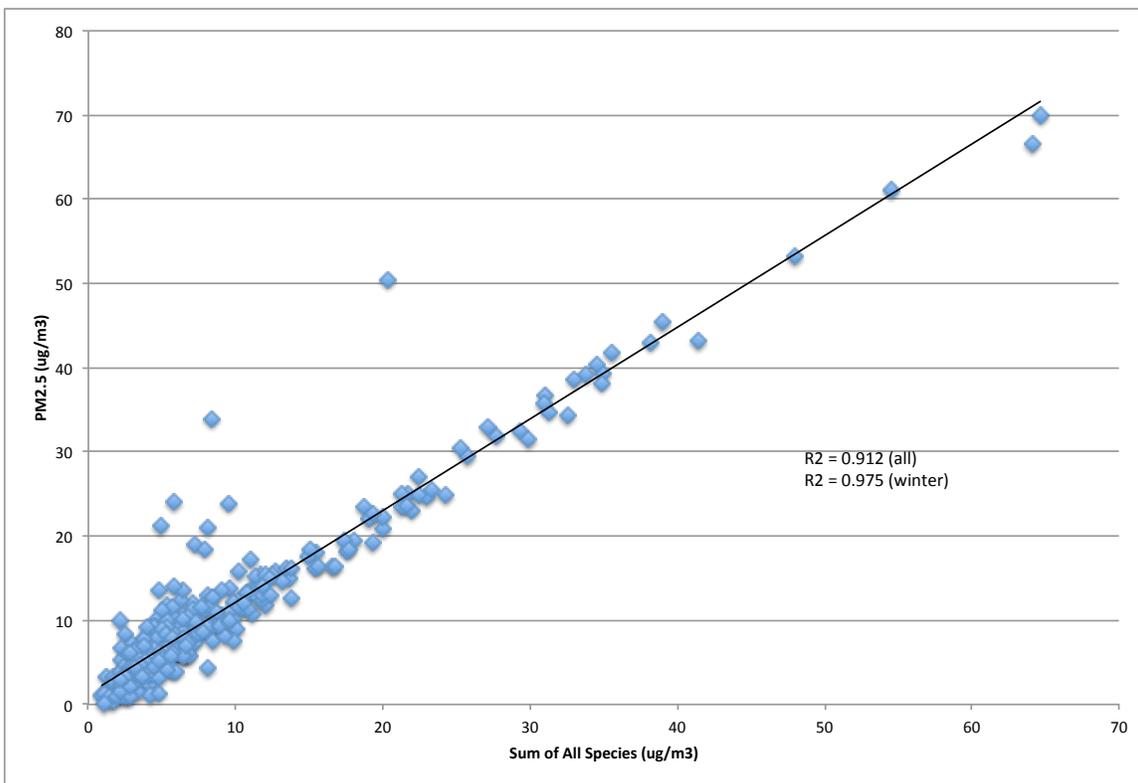


Figure S-1. Sum of species vs. PM_{2.5} concentration for the Salt Lake City monitoring site.

Table S-1: Days removed from the Unmix analysis for Salt Lake City.

Location	Date	PM _{2.5} (ug/m ³)	Reason for exclusion
Bountiful	7/5/07	10.4	Fireworks
Bountiful	1/1/08	16	Fireworks
Bountiful	7/5/08	15.7	Fireworks
Bountiful	1/1/09	26.8	Fireworks
Bountiful	7/24/09	7.4	Fireworks
Bountiful	7/25/10	8.4	Fireworks
Bountiful	10/21/07	1	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	11/2/07	5.1	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	12/2/07	5.7	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	4/12/08	4.1	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	5/24/08	1.2	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	10/15/08	5.5	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	11/14/08	1.9	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	2/24/09	2.2	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	3/14/09	6.7	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	9/16/09	1.5	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	10/28/09	1.6	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	11/3/09	6.8	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	12/3/09	2.3	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	2/25/10	3.9	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	4/14/10	2.5	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	4/26/10	2.5	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	5/14/10	6.9	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	11/10/10	2.9	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	2/20/11	1.5	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	4/21/11	2.2	Sum of species exceeded PM _{2.5} by more than a 20%
Bountiful	9/9/08	18.5	Sum of species was less than 50% of PM _{2.5}
Bountiful	6/11/08	4.1	Sum of species was less than 50% of PM _{2.5}
Lindon	7/5/07	23.2	Fireworks
Lindon	1/1/08	16.8	Fireworks
Lindon	7/5/08	18.5	Fireworks
Lindon	7/24/09	10.7	Fireworks
Lindon	7/25/10	12.7	Fireworks
Lindon	8/28/08	1.7	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	9/16/09	2	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	3/28/11	0.6	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	1/25/08	3.2	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	9/22/09	2.6	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	10/16/09	4.2	Sum of species exceeded PM _{2.5} by more than a 20%
Lindon	2/6/08	34.9	Sum of species was less than 50% of PM _{2.5}

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3	Lindon	7/11/07	36.5	Sum of species was less than 50% of PM _{2.5}
4	Salt Lake City	1/1/08	19.5	Fireworks
5	Salt Lake City	7/5/08	12.6	Fireworks
6	Salt Lake City	1/1/09	30.4	Fireworks
7	Salt Lake City	7/24/09	13.5	Fireworks
8	Salt Lake City	7/4/10	13.6	Fireworks
9	Salt Lake City	7/25/10	15.8	Fireworks
10	Salt Lake City	1/1/11	5.8	Fireworks
11	Salt Lake City	2/9/09	0.1	Sum of species exceeded PM _{2.5} by more than a 20%
12	Salt Lake City	5/9/11	0.2	Sum of species exceeded PM _{2.5} by more than a 20%
13	Salt Lake City	8/31/07	1.0	Sum of species exceeded PM _{2.5} by more than a 20%
14	Salt Lake City	9/18/07	1.3	Sum of species exceeded PM _{2.5} by more than a 20%
15	Salt Lake City	9/24/07	0.8	Sum of species exceeded PM _{2.5} by more than a 20%
16	Salt Lake City	4/30/11	0.9	Sum of species exceeded PM _{2.5} by more than a 20%
17	Salt Lake City	4/18/11	0.9	Sum of species exceeded PM _{2.5} by more than a 20%
18	Salt Lake City	10/3/07	1.4	Sum of species exceeded PM _{2.5} by more than a 20%
19	Salt Lake City	3/5/09	0.9	Sum of species exceeded PM _{2.5} by more than a 20%
20	Salt Lake City	3/26/09	1.3	Sum of species exceeded PM _{2.5} by more than a 20%
21	Salt Lake City	3/30/10	50.4	Sum of species was less than 50% of PM _{2.5}
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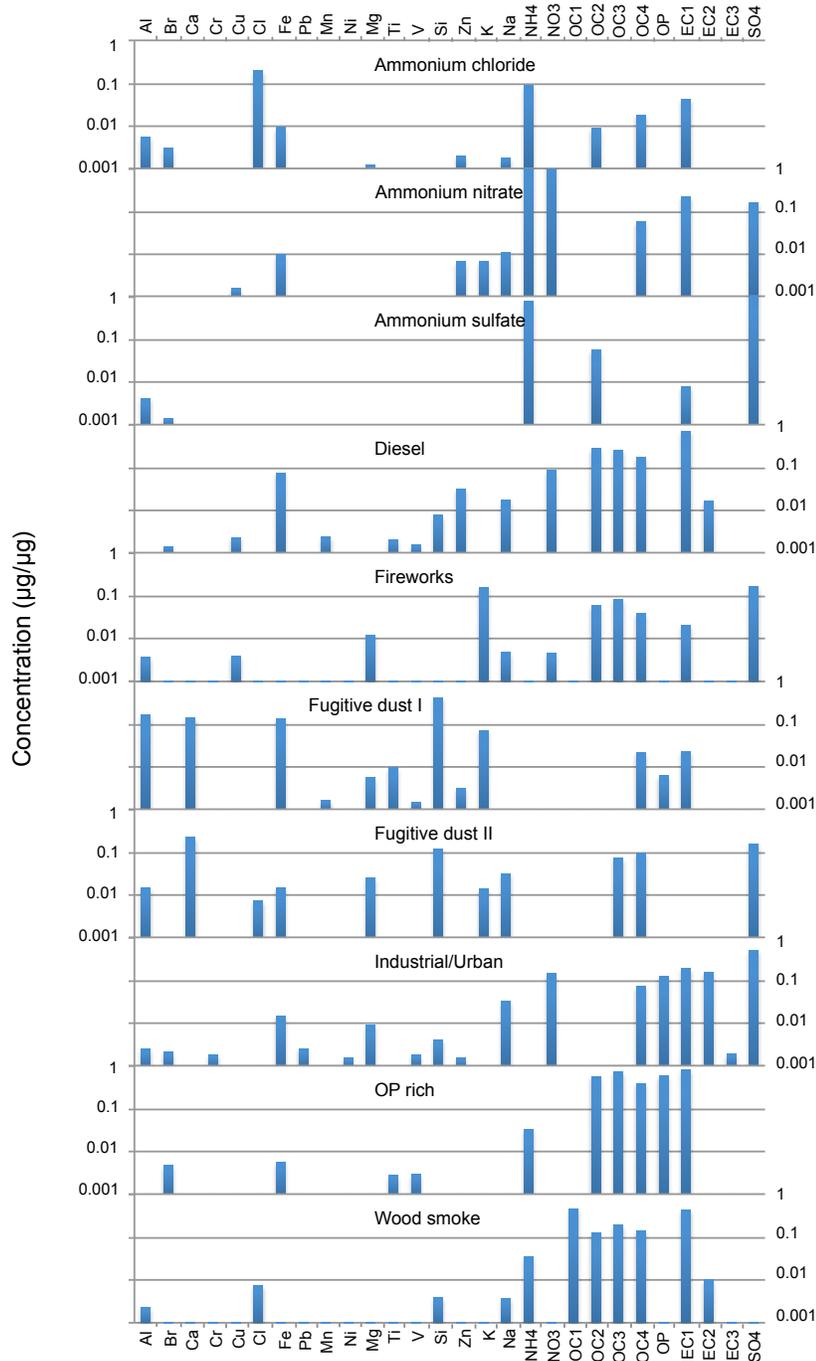


Figure S-2. Species contributions to PMF factor profiles for Bountiful.

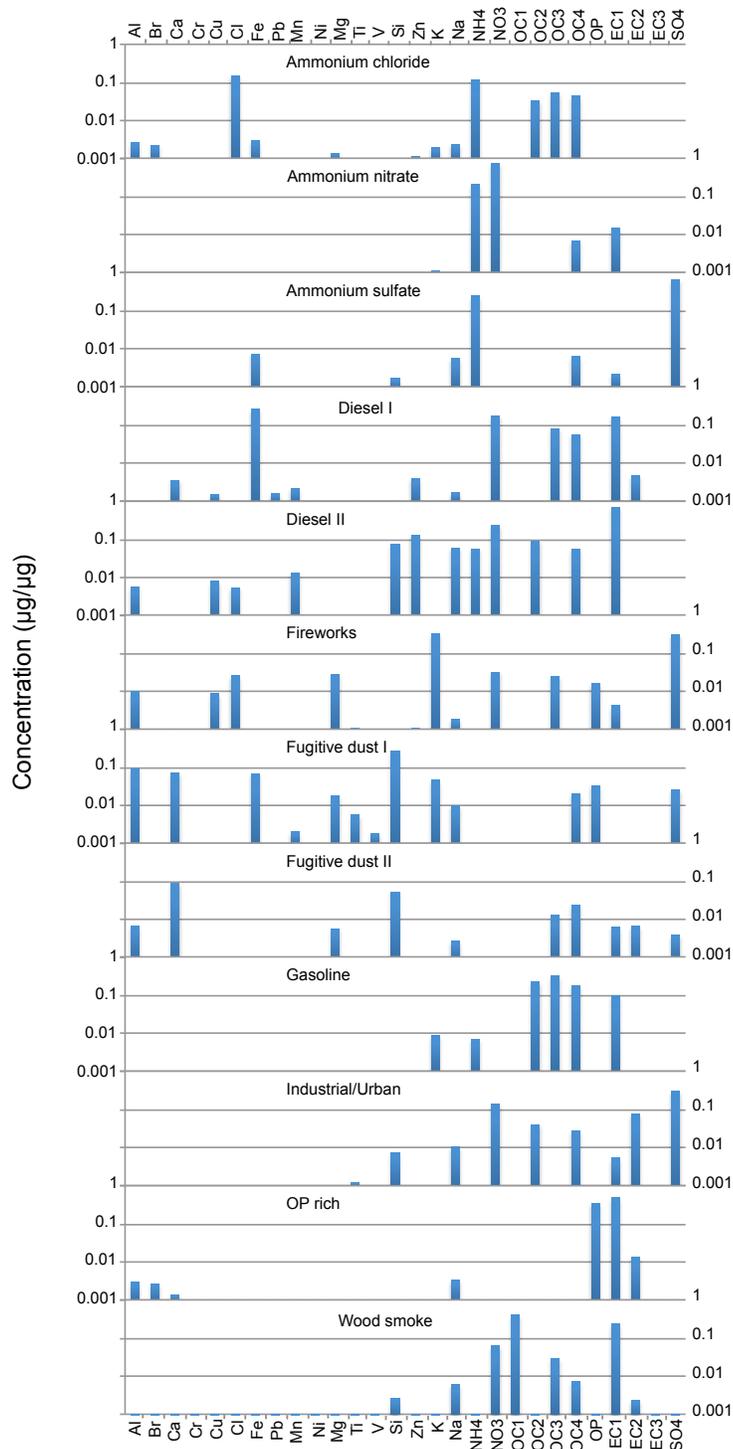
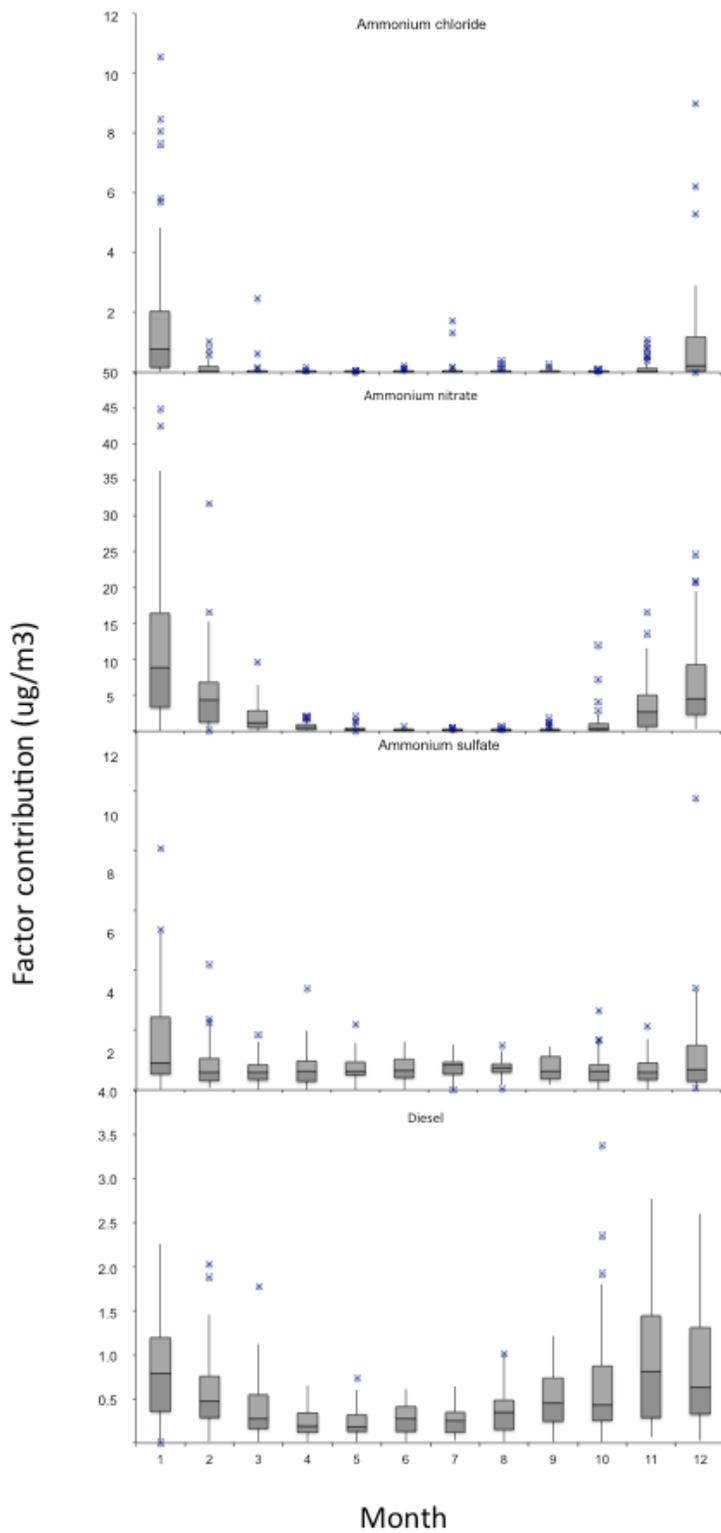
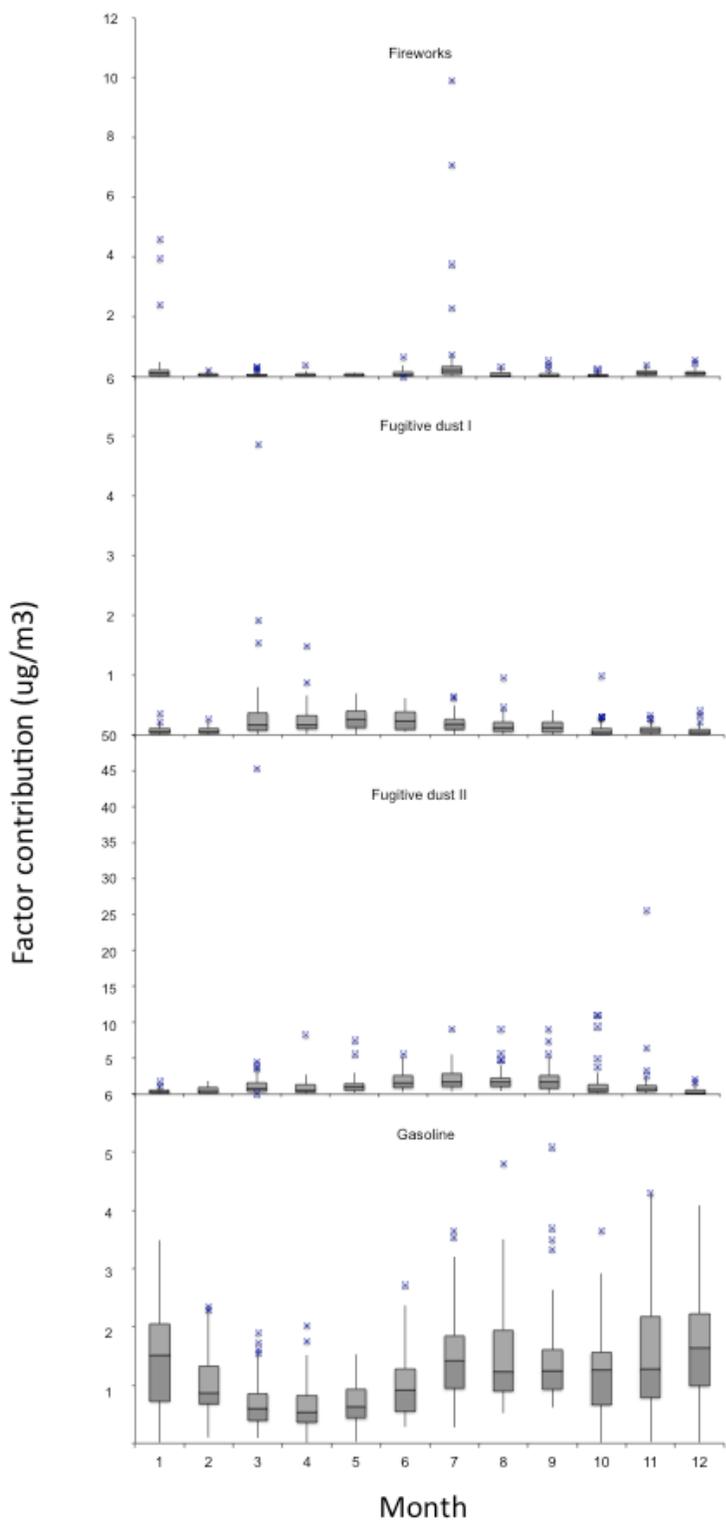


Figure S-3. Species contributions to PMF factor profiles for Lindon.



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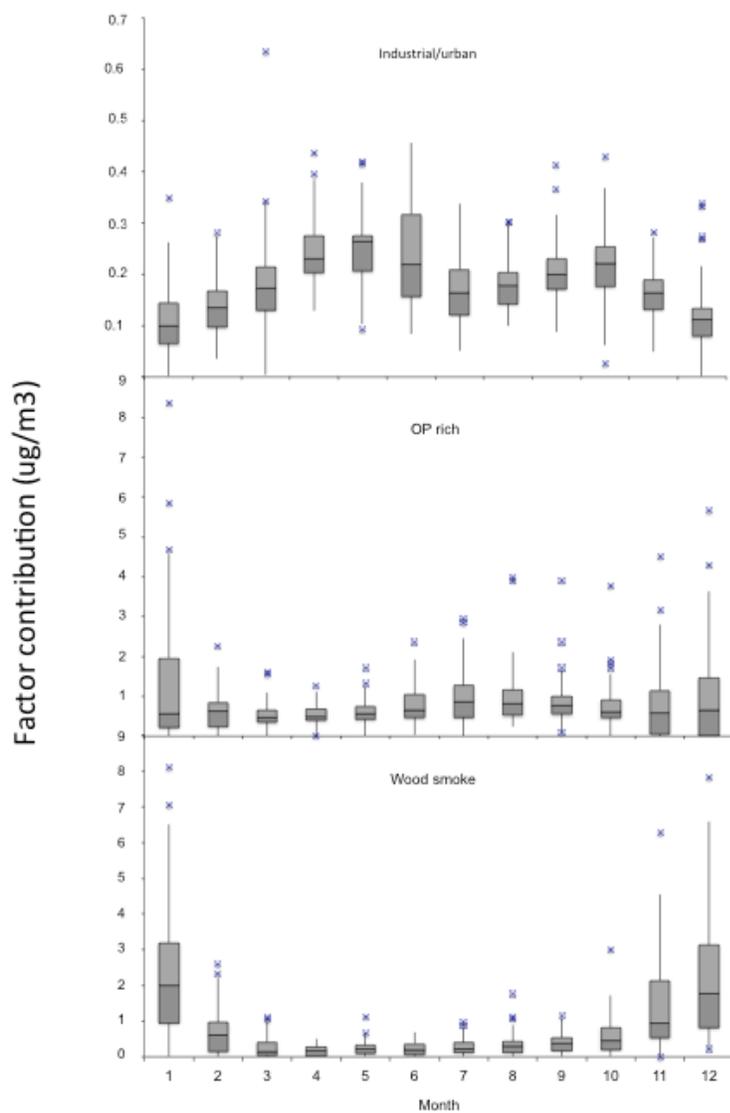


Figure S-4. Monthly PMF factor contributions for Salt Lake City. The box indicates the 25th and 75th percentiles, and the line within the box indicates the mean value. The lines indicate the maximum and minimum values, excluding outliers, which are denoted by an 'x' and exceed the interquartile range by more than a factor of 1.5.

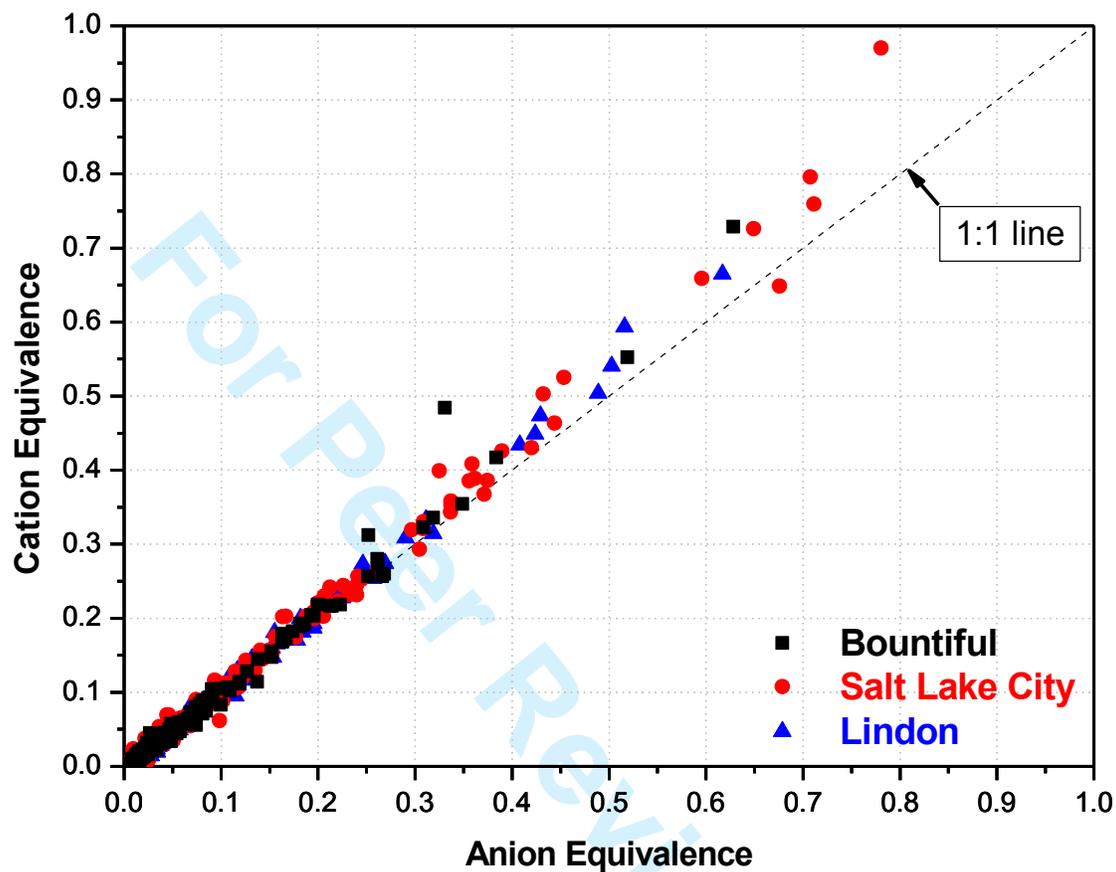


Figure S-5. Aerosol charge balance between anions and cations at three STN monitoring locations along Utah's Wasatch Front. Anion equivalence was calculated as the sum $[Cl^-]/35.453 + [NO_3^-]/62.005 + [SO_4^{2-}]/48.03$, cation equivalence was calculated as the sum $[Na^+]/23.0 + [K^+]/39.098 + [NH_4^+]/18.04$.

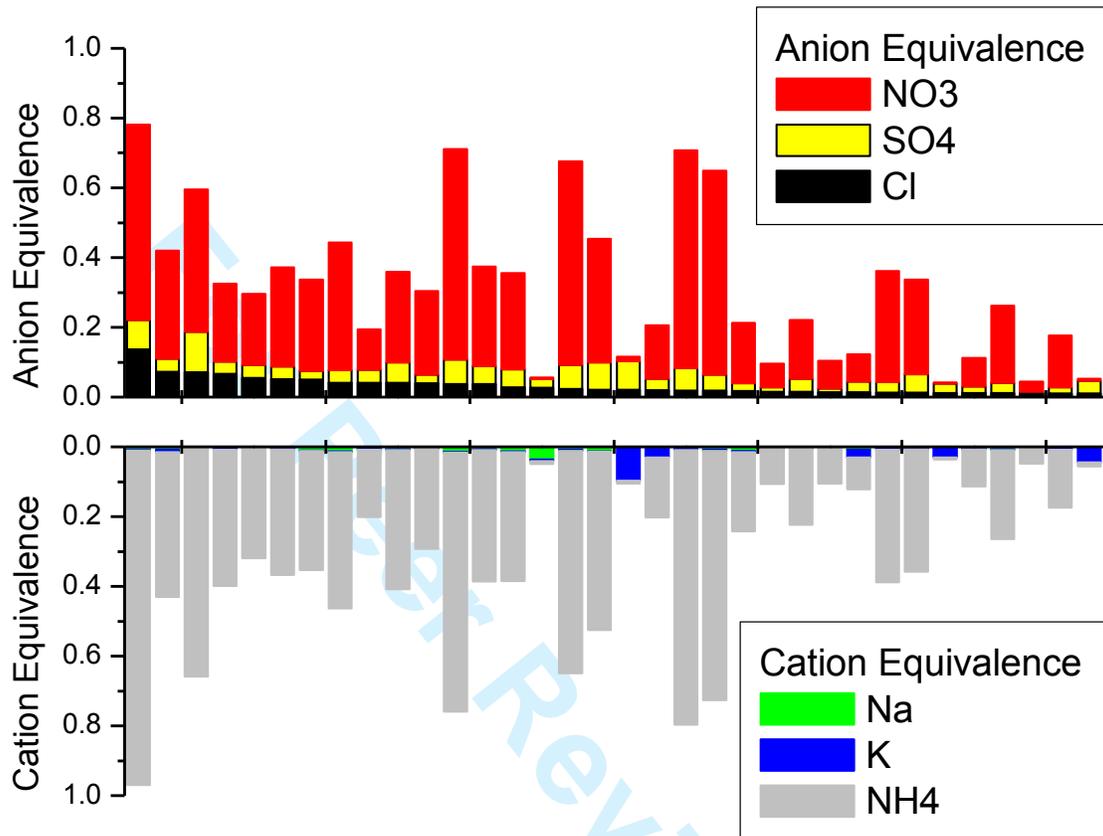


Figure S-6. Cation and anion charge equivalence for the highest measured aerosol Cl days in Salt Lake City.

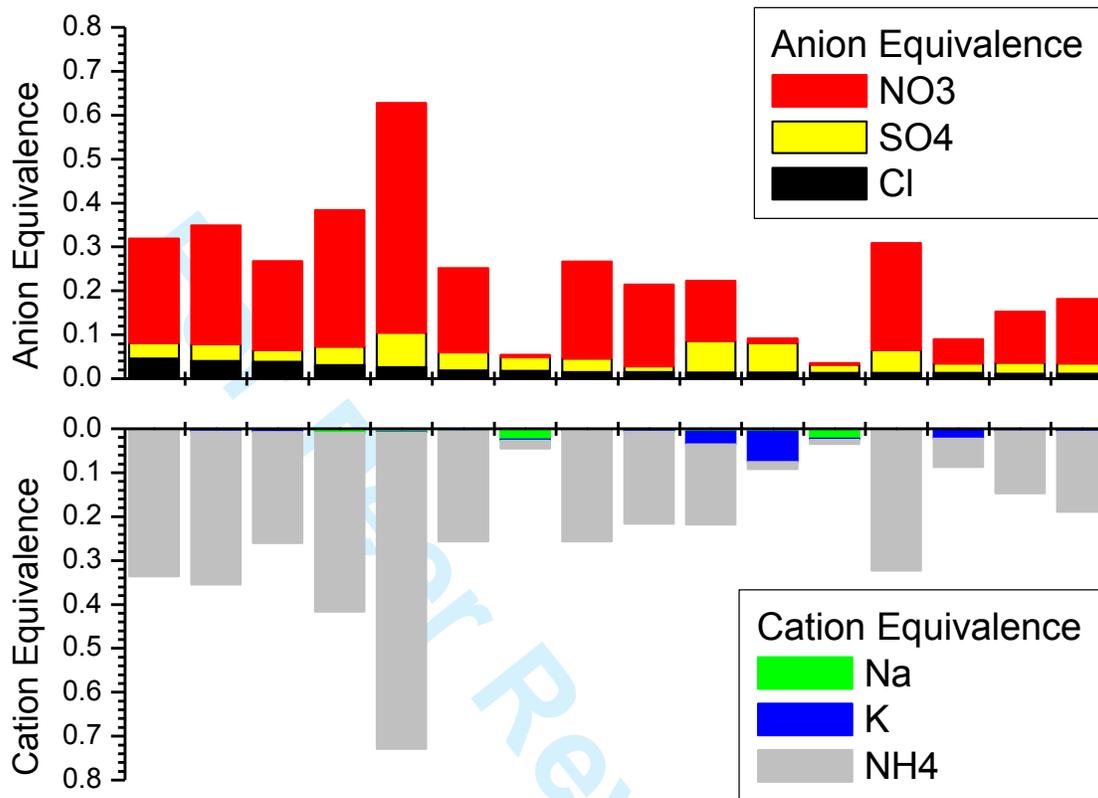


Figure S-7. Cation and anion charge equivalence for the highest measured aerosol Cl days in Bountiful.

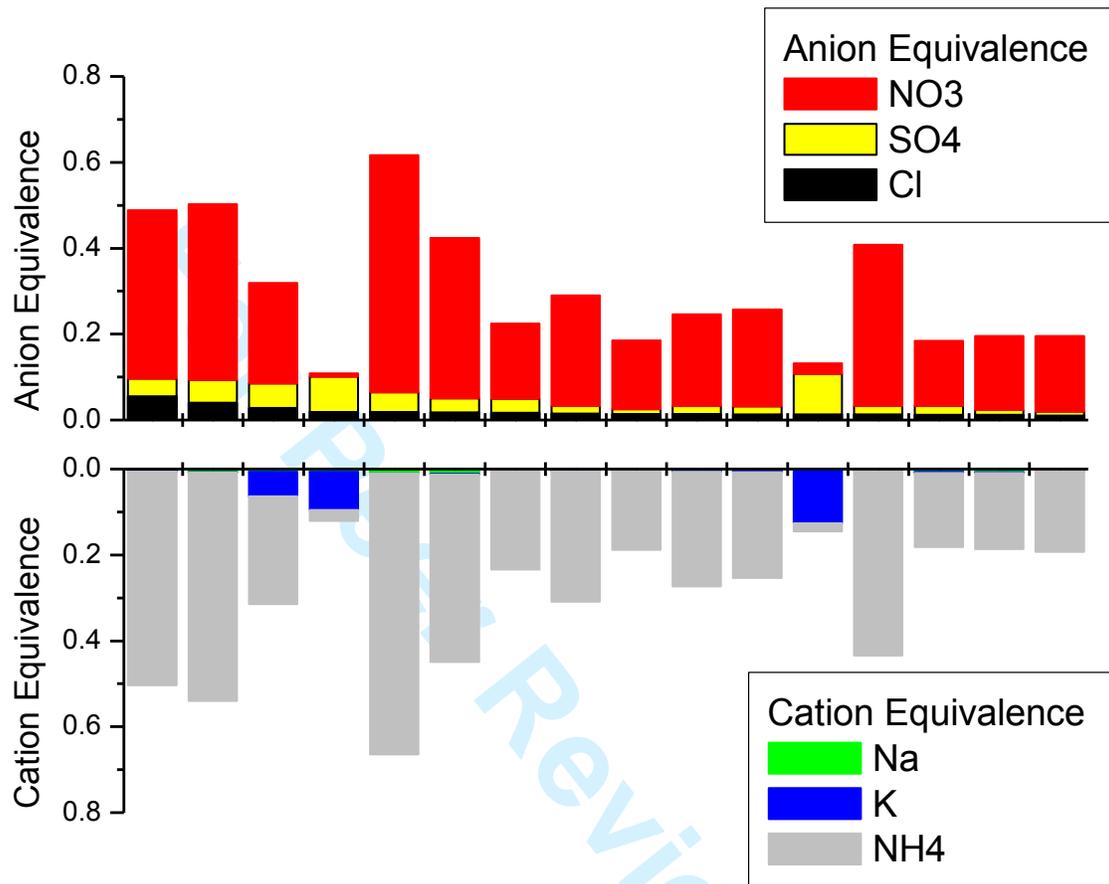


Figure S-8. Cation and anion charge equivalence for the highest measured aerosol Cl days in Lindon.

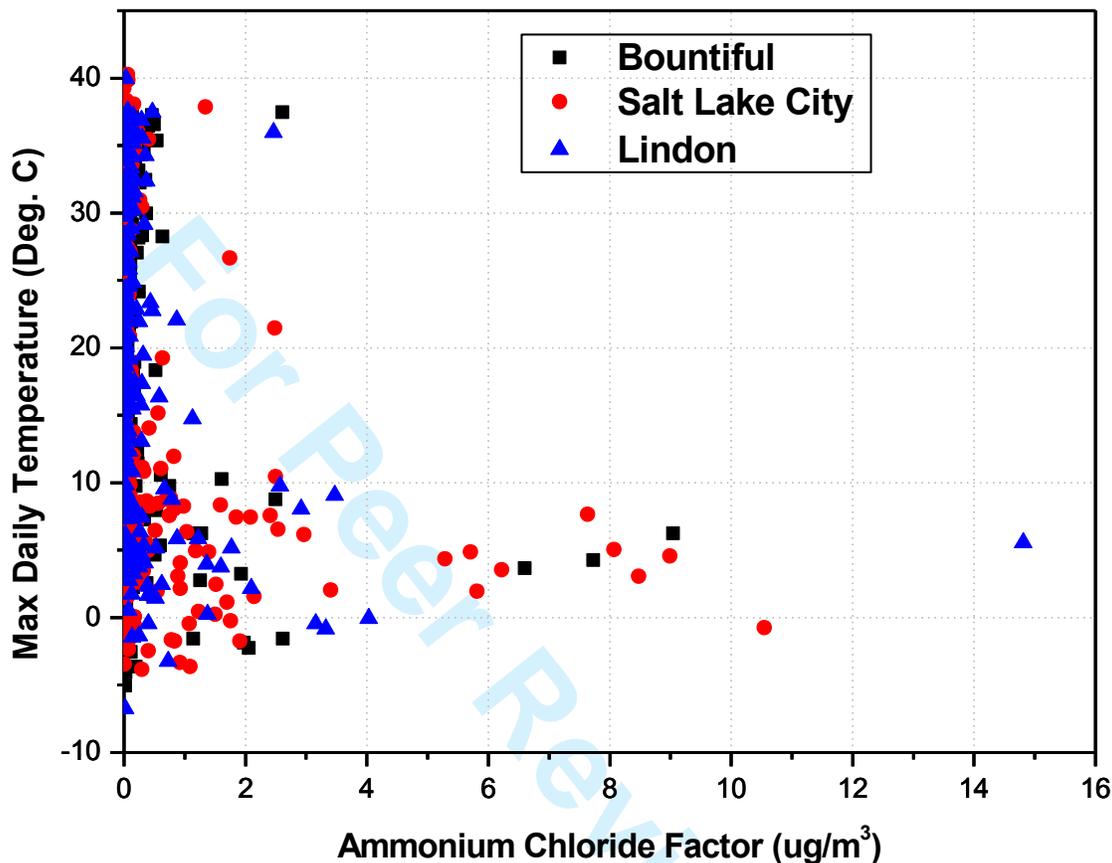


Figure S-9. PMF factor for ammonium chloride plotted against maximum daily temperature at three STN monitoring locations along Utah's Wasatch Front.

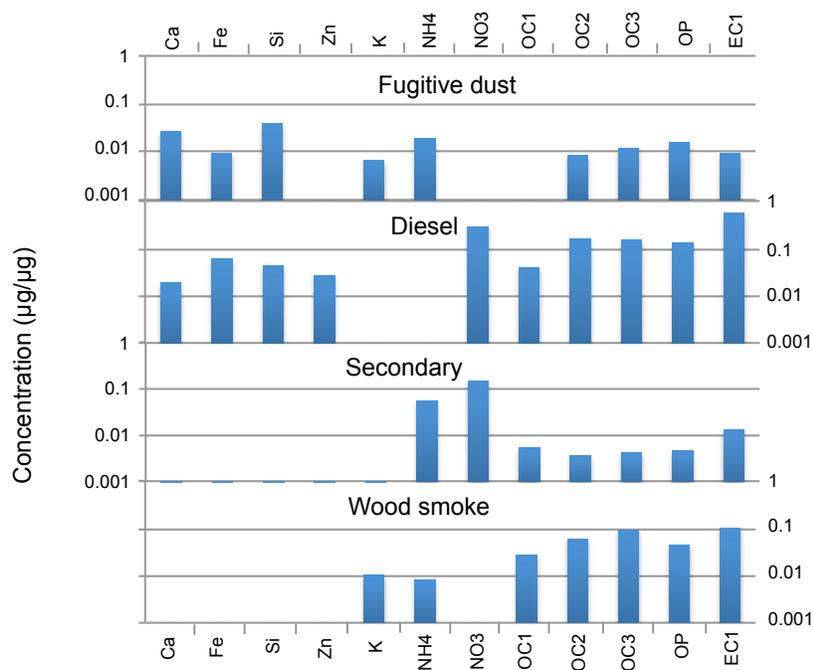


Figure S-10. Species contributions to Unmix factor profiles for Bountiful.

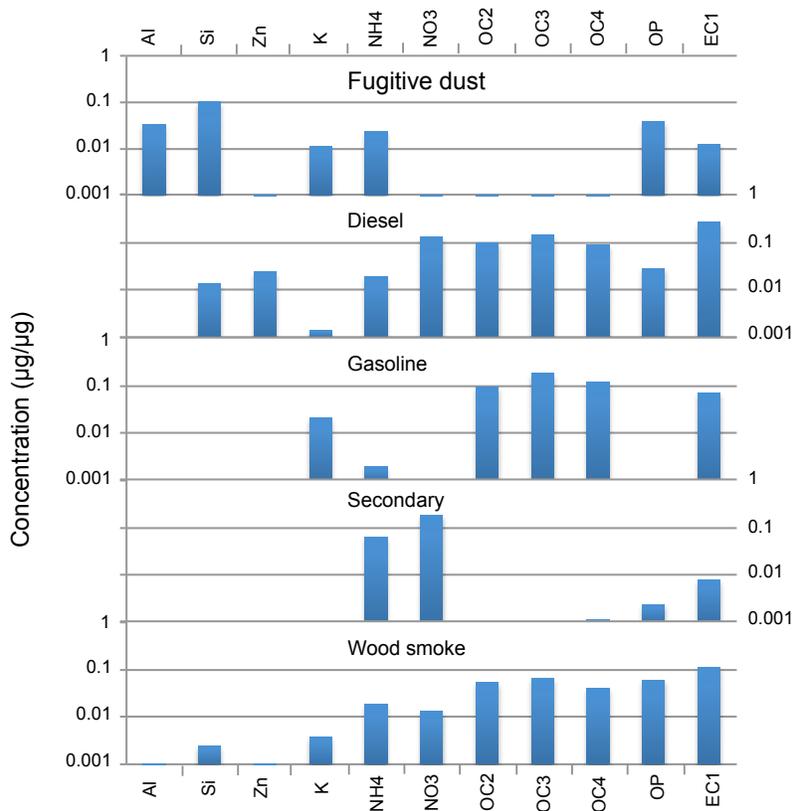
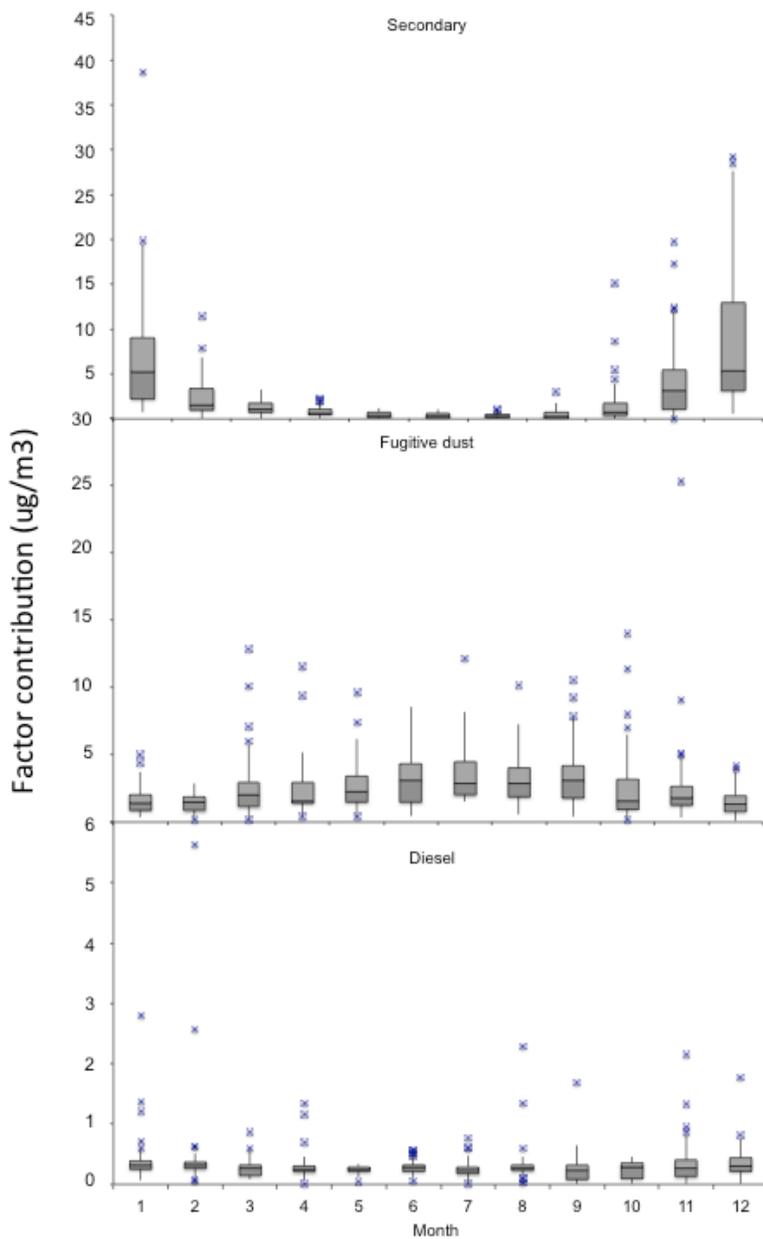


Figure S-11. Species contributions to Unmix factor profiles for Lindon.

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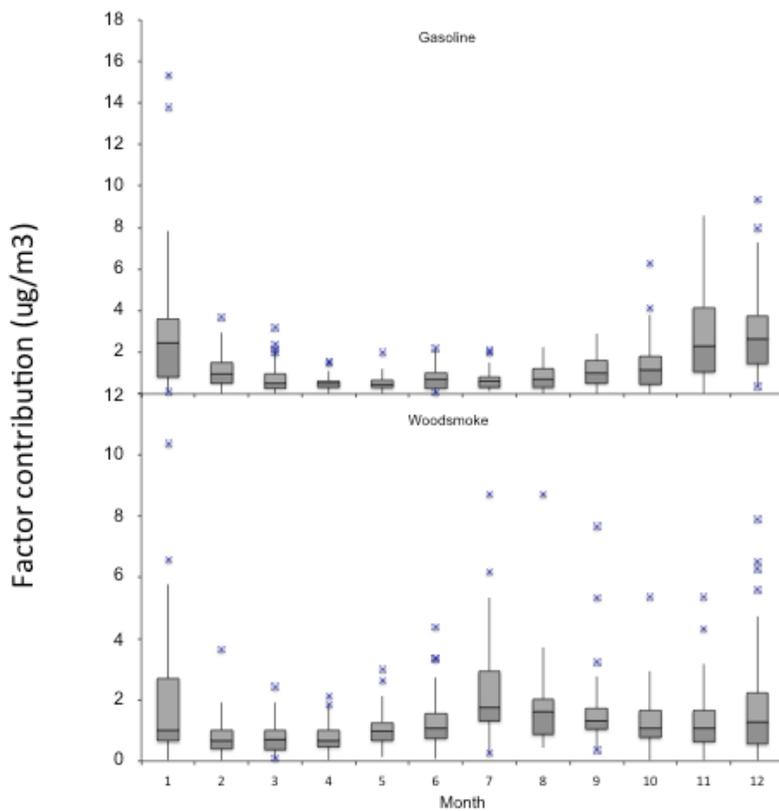


Figure S-12. Monthly Unmix factor contributions for Salt Lake City. The box indicates the 25th and 75th percentiles, and the line within the box indicates the mean value. The lines indicate the maximum and minimum values, excluding outliers, which are denoted by an 'x' and exceed the interquartile range by more than a factor of 1.5.