

Spatially refined aerosol direct radiative forcing efficiencies

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Global aerosol direct radiative forcing (DRF) is an important metric for assessing potential climate impacts of future emissions changes. However, the radiative consequences of emissions perturbations are not readily quantified nor well understood at the level of detail necessary to assess realistic policy options. To address this challenge, here we show how adjoint model sensitivities can be used to provide highly spatially resolved estimates of the DRF from emissions of black carbon (BC), primary organic carbon (OC), sulfur dioxide (SO_2) and ammonia (NH_3), using the example of emissions from each sector and country following multiple Representative Concentration Pathways (RCPs). The radiative forcing efficiencies of many individual emissions are found to differ considerably from regional or sectoral averages for NH_3 , SO_2 from the power sector, and BC from domestic, industrial, transportation and biomass burning sources. Consequently, the amount of emissions controls required to attain a specific DRF varies at intra-continental scales by up to a factor of four. These results thus demonstrate both a need and means for incorporating spatially refined aerosol DRF into analysis of future emissions scenario and design of air quality and climate change mitigation policies.

1 Introduction

Anthropogenic enhancements to aerosol abundances have significantly influenced climate since pre-industrial times through their combined direct radiative forcing (DRF) of $-0.5 \pm -0.4 \text{ Wm}^{-2}$ (1). Changes to aerosol and aerosol precursor emissions are expected, either intentionally or inadvertently, to exert further influence on climate in coming decades (2–4). Such emissions changes may encompass a range of impacts to multiple source sectors, in multiple regions, altering the abundances of different types of aerosols and greenhouse gases with several competing consequences. Policy makers are thus concerned with refining our understanding of the link between emissions changes and radiative forcing.

However, estimating the radiative forcing of numerous, detailed future emissions scenarios is challenging for several reasons. First, radiative effects of aerosols and other short-lived climate forcers (SLCFs) are spatially variable, so unlike emissions of long-lived greenhouse gases, the location of their sources modulates their impact. Hence, recent studies have moved from abundance-based metrics of radiative forcing to quantifying the forcing of emissions from specific regions and sectors (5–14). Second, accounting for the full range of aerosol interactions and feedbacks, from the aerosol indirect (15, 16) and semi-direct (17, 18) effects to aerosol-gas interactions (19, 20), requires lengthy model calculations. Lastly, emissions projections for future scenarios themselves are inherently uncertain. To address all of these issues, ensembles of detailed climate model calculations could be performed across a wide spectrum of regional, sector and species specific emissions scenarios. However, using standard approaches, it is too computationally expensive to separate impacts at a source-specific level across numerous scenarios in detailed models.

Here we present, validate and apply an entirely new approach to quantifying the radiative forcing impacts of emissions scenarios at a resolution several orders of magnitude

finer than previously considered. Presently we focus exclusively on aerosol direct radiative forcing as a necessary first step towards understanding the broader impacts of aerosols. Sector and region-specific direct global aerosol radiative forcing may be more important than global indirect effects (14), hence the DRF alone is significant enough to provide valuable guidance towards more detailed assessments. Previous studies attributing radiative forcing to sources at regional scales have relied on multiple model evaluations wherein subsets of emissions are perturbed sequentially, a method that becomes prohibitively expensive as the number of emissions sets considered becomes large (>100). In contrast, adjoint modeling is a means by which variations in a model response function are propagated backwards in time through an auxiliary (adjoint) set of model equations, ultimately yielding the sensitivities of this function with respect to all model inputs simultaneously at a computational cost independent of the number of inputs considered (21, 22). Here we introduce this approach as a means of calculating the sensitivities of direct aerosol radiative effects with respect to emissions. This provides an estimate of the instantaneous global aerosol DRF from emissions of every aerosol and aerosol precursor emission, from each source sector, in each grid cell of the model several orders of magnitude faster than conventional methods.

2 Methods

In the following sections we describe how a chemical transport model and its adjoint are used to calculate DRF sensitivities.

2.1 GEOS-Chem

GEOS-Chem¹ is a chemical transport model driven with assimilated meteorology from the Goddard Earth Observing System (GEOS) of the NASA Global Modeling and Assimilation Office (GMAO). GEOS-5 meteorological data sets are down-sampled to facilitate detailed simulation of tropospheric gas-phase chemistry (23), here run at $2^\circ \times 2.5^\circ$ resolution globally using model version 8-02-01 with relevant patches and updates up through v9. Aerosols in GEOS-Chem (24, 25) are treated as an external mixture. The size distributions are log normal with fixed width and dry mode diameters, subject to hygroscopic growth. Bulk partitioning of secondary inorganic species is calculated with the RPMARES thermodynamic scheme (26). Aging of primary black carbon from hydrophobic to hydrophilic occurs with an e -folding time of 1.15 days. Aerosols are coupled with oxidant simulations through heterogeneous chemistry (27, 28) and aerosol effects on photolysis rates (29). Global anthropogenic emissions of NO_x and SO_x are from EDGAR (30) and from Bond et. al (31) for carbonaceous aerosol, overwritten by regional inventories where available (32). NH_3 emissions are described in Park et. al (25), monthly biomass burning emissions are from GFEDv2 (33) and biofuel emissions from Yevich and Logan (34). Dry deposition is calculated using a resistance in series approach (35) and wet losses include in-cloud and below-cloud rainout and convective scavenging (27, 36).

Aerosol optical properties in GEOS-Chem (29, 37) follow the Global Aerosol Data Set (GADS) (38, 39). All aerosol species are assumed to be externally mixed, which will likely lead to underestimation of aerosol absorption (40). To compensate, we enforce a BC mass absorption of 1.5 (41). Calculation of aerosol radiative forcing employs five spectral bands spanning 4400 to 32260 cm^{-1} (37). The LIDORT radiative transfer model (42, 43) is used to calculate the top of the atmosphere (TOA) radiative flux, using surface reflectances

¹<http://www.geos-chem.org>

from GOME (44) that span from 12987 cm^{-1} to 27027 cm^{-1} and are extrapolated to cover the shortwave spectrum considered here. Unless noted otherwise, calculations are for all-sky conditions wherein forcing is masked by the GMAO cloud fraction of each column. The resulting combined global yearly average AOD at 500 nm from ammonium sulfate, ammonium nitrate, BC, and OC of 0.078 is consistent with previous modeling studies (45), which reported global average AOD at 550 nm for sulfate, BC, and primary organic matter ($= 1.4 \times \text{OC}$) of 0.057 and ranging by more than a factor of two amongst models with unified emissions. The pre-industrial (1850) to present (2000) aerosol all-sky TOA DRF is calculated to be -0.47 Wm^{-2} using emissions from the Climate Model Intercomparison Program (CMIP5) (46) in support of the fifth Intergovernmental Panel on Climate Change (IPCC) report. The breakdown of this forcing into contributions from specific regions and sectors is provided in Fig. S4. Excluding biomass burning, the entire aerosol DRF due to constant 2000 emissions is only slightly (3%) greater than the 2000 relative to 1850 DRF, so that these results also provide a useful perspective on the impacts of current emissions on future climate. These values are broadly consistent with those in the fourth IPCC report (1), considering (a) that we likely overestimate NH_3 forcing (see Fig. S3) and (b) that the IPCC reports abundance-based forcings, while here we consider emissions based forcings, which, for example, are lower for SO_2 than for aerosol sulfate (20).

2.2 GEOS-Chem adjoint sensitivities of radiative forcing

An adjoint model is a set of equations auxiliary to a forward model that are used to efficiently calculate sensitivities of a scalar model response function with respect to numerous model inputs (21). The adjoint of GEOS-Chem is presently the only such model to include gas-phase chemistry, heterogeneous chemistry, black and organic primary aerosol,

and sulfate-ammonium-nitrate formation chemistry and thermodynamics (47), with code updates following the relevant parts of the GEOS-Chem forward model up through version v9. In this work we extend the GEOS-Chem adjoint to include calculation of sensitivities of a model response, $\mathcal{J}(\sigma)$, defined to be the direct radiative forcing of aerosols from present-day anthropogenic emissions, σ ,

$$\mathcal{J}(\sigma) = \uparrow F(\sigma^0) - \uparrow F(\sigma^0 + \sigma), \quad (1)$$

where $\uparrow F$ is the global upward radiative flux at the top of the atmosphere and σ^0 is the vector of pre-industrial emissions. Calculation of adjoint sensitivities (λ_σ) is initiated with expansion of the derivative of the forcing with respect to aerosol concentrations at each time step using the chain-rule,

$$\left(\frac{\partial \mathcal{J}}{\partial \mathbf{c}^n} \right)^T = (\mathbf{M}_a^n)^T (\mathbf{M}_{opt}^n)^T (\mathbf{M}_{rtm}^n)^T \mathbf{m}_w \frac{\partial \mathcal{J}(\sigma)}{\partial \uparrow F(\sigma^i + \sigma)} \quad (2)$$

Evaluating this equation from right to left, the $\frac{\partial \mathcal{J}(\sigma)}{\partial \uparrow F(\sigma^i + \sigma)}$ term is -1 . \mathbf{m}_w is a weighting vector that accounts for the contribution of the radiative flux in a particular column at a single time step to the global yearly average flux, including the column cloud fraction for all-sky calculations. \mathbf{M}_{rtm} is the Jacobian of the instantaneous TOA upward radiative flux in a single column with respect to the aerosol optical thickness, scattering albedo and phase function in each vertical level of that column, calculated analytically by LIDORT (42, 43). \mathbf{M}_{opt} is the Jacobian of these optical properties with respect to the aerosol wet diameter for each species in each cell. \mathbf{M}_a is the Jacobian of the aerosol wet diameter with respect to the aerosol mass concentrations, \mathbf{c}^n , where \mathbf{c}^n is the vector of all K tracer concentrations, $\mathbf{c}^n = [c_1^n, \dots, c_K^n]^T$ at time step $n=0, \dots, N$. \mathbf{M}_{opt} is calculated using analytic derivatives from a linearized Mie scattering calculation. The product of $(\mathbf{M}_a^n)^T$ with the vector $(\mathbf{M}_{opt}^n)^T (\mathbf{M}_{rtm}^n)^T \mathbf{m}_w \frac{\partial \mathcal{J}(\sigma)}{\partial \uparrow F(\sigma^i + \sigma)}$ is calculated using adjoints. The adjoint model then

takes as input $\left(\frac{\partial \mathcal{J}}{\partial \mathbf{c}^n}\right)^T$; after integrating backwards in time the solution of the adjoint model is the vector of sensitivities with respect to emissions $\boldsymbol{\lambda}_\sigma = \nabla_\sigma \mathcal{J}$.

To consider how radiative forcing will change owing to the difference between anthropogenic emissions in the future, $\boldsymbol{\sigma}^f$, and the present, $\boldsymbol{\sigma}$, the change in forcing, $\Delta \mathcal{J} = \mathcal{J}(\boldsymbol{\sigma}^f) - \mathcal{J}(\boldsymbol{\sigma})$, can be approximated as

$$\Delta \mathcal{J} \approx \boldsymbol{\lambda}_\sigma^T \Delta \boldsymbol{\sigma} \quad (3)$$

where $\Delta \boldsymbol{\sigma} = \boldsymbol{\sigma}^f - \boldsymbol{\sigma}$, $\Delta \mathcal{J}$ is the corresponding DRF, and $\boldsymbol{\lambda}_\sigma$ is the vector of sensitivities of \mathcal{J} to these emissions from the solution of the adjoint model. Thus, $\boldsymbol{\lambda}_\sigma$ represents a fundamental radiative forcing efficiency ($\text{Wm}^{-2}/\text{emission}$) that is used to linearly estimate DRF responses for any arbitrary set of emissions perturbations. The accuracy of the adjoint-based estimates is extensively validated (see Supporting Information and Figs. S1 and S2); grid-level radiative forcing estimated from the adjoint sensitivities correspond well ($r^2 > 0.97$) with changes in radiative forcing calculated explicitly using the full forward model. We also consider the extent to which linear extrapolations of adjoint sensitivities to estimate DRF are valid over a range of spatially aggregated perturbations to emissions. While nonlinearities triggered by perturbing global emissions of NO_x diminish the relationship between globally aggregated forward and adjoint sensitivities, similar tests for SO_2 , BC, OC and NH_3 are well correlated ($r^2 = 0.98$) and have low bias, hence we restrict our assessment to these species.

3 Radiative forcing efficiencies

Application of the adjoint model yields radiative forcing efficiencies for each aerosol precursor emission in each grid cell, which indicate how arbitrary changes to emissions can have location-dependent impacts. Yearly average efficiencies, in $(\text{Wm}^{-2})/(\text{kgm}^{-2} \text{ yr}^{-1})$,

are constructed from twelve separate week-long adjoint model calculations, one per month, for which the response function (Eq. 1) is the average DRF over a 24 h period and sensitivities are integrated backward in time for a week prior to this period. The resulting efficiencies (Fig. 1) show the change in global DRF with respect to emissions perturbations in each location. The variability in the efficiencies is driven by several factors, the largest being the degree to which emissions from a particular grid cell contribute to aerosol concentrations over surfaces of low or high albedo. Variability in loss rates also plays a role; these two factors entirely govern the efficiency distribution of BC and OC. For NH_3 , the efficiency is further regulated by the degree of neutralization of the aerosols (48). For SO_2 , emissions that form sulfate aerosol in dry, oxidative conditions have a higher radiative forcing efficiency than emissions leading to formation of sulfate in clouds, where sulfate is more readily scavenged. Over snow, even reflective aerosols can have a positive forcing through effective enhancement of the radiative path length. Overall, the heterogeneity of the DRF efficiencies underscores the importance of spatially refined treatment of aerosol DRF.

4 Radiative forcing of future emissions scenarios

The spatial variability in the radiative forcing efficiencies shown in Fig. 1 indicates that regionally or sectorally aggregated radiative forcing estimates could misrepresent the impacts of finer scale trends in emissions. To investigate the radiative effects of such trends, we multiply these radiative forcing efficiencies by spatially resolved estimates of emission changes from 2000 to 2050. These changes follow Representative Concentration Pathways (RCPs (49)), a set of stabilization scenarios employing different land-use and greenhouse gas emissions projections to meet radiative forcing targets in the year 2100 that are progressively more aggressive, with RCP 8.5 leading to the highest greenhouse

gas emissions (50), followed by RCP 6.0 (51, 52), RCP 4.5 (53–55), and RCP 2.6 with the largest reductions in greenhouse gases (56). However, it is important to note that the RCPs are not derived from a common framework, and the progression of the emissions of aerosols amongst them are only indirectly coupled to their radiative forcing targets. Thus, differences between the RCPs can not directly be interpreted as a result of specific climate policies or socioeconomic developments.

4.1 Variability in efficiency vs emissions magnitude

The extent to which variability in DRF is driven by the magnitude of emissions changes vs the efficiency of such changes is analyzed in detail for the mid-range RCP 4.5 scenario in the year 2050 relative to 2000. Figure 2 shows the percent change in emissions in each grid cell, by species and sector, vs the percent by which RCP 4.5 emissions changes contribute to the total DRF of that particular species; the total efficiency of each sector across all locations is included as supplementary material (Fig. S3). Any negative DRF from BC or positive DRF from SO_2 , OC or NH_3 indicates a location where the RCP 4.5 emissions are lower in 2050 than 2000, and vice versa.

Changes to precursor emissions of secondary aerosols have widely variable radiative impacts. There are two distinct response regimes for NH_3 where DRF is not commensurate with emissions changes. Figure 3 shows the actual emissions changes and the corresponding DRFs for the NH_3 agricultural sector. These two regimes correspond to increases in NH_3 in parts of Africa and India where aerosol formation is not NH_3 limited, compared to areas in China, Europe and the U.S., where small amounts of NH_3 can lead to efficient formation of ammonium nitrate. The DRF for SO_2 emissions are shown in Fig. 3 to be highly variable on a grid-cell by grid-cell basis; the DRF of 0.5% changes to total SO_2 emissions via the power sector may vary by a factor of four (from 0.25% to 1%).

However, the globally aggregated DRF of each SO_2 sector closely follows their emissions change (see Fig. S3). Thus, there could be substantial error in assuming a uniform DRF across all locations for NH_3 and SO_2 from the agriculture and energy sectors, respectively.

For primary carbonaceous aerosol, increases to BC emissions from the industry, domestic and transportation sectors occur mostly over sub-Saharan Africa, where BC DRF efficiency is low, leading to consistently muted DRFs. For BC emissions reductions in China, India, Europe and the U.S., the efficiency is less consistent. In total, BC emissions changes of 35%, 19% and 2%, from industry, domestic, and transportation sectors, respectively, contribute to 46%, 34% and 10% of the total BC DRF, a response that is a third to a factor of five times larger than the corresponding emissions changes. Changes to biomass burning emissions of BC account for 33.4% of total reduction to BC, but the contribution to BC DRF of the biomass burning sector as a whole is nearly neutral, as small increases to BC emissions in Canada and Russia of 2.3% and 1.5%, respectively, have a large impact (15% combined) on BC DRF, which leads to the steeply sloping cluster of red points in the lower left of the BC panel of Fig. 2. The impact of OC emissions is more closely tied to emissions rates, although the total DRF efficiency of domestic and industry emissions is slightly greater than the response to biomass burning.

To further investigate both regional and sector variability, DRFs aggregated over 24 regions² are shown in Fig. 4. Changes to emissions from the domestic, industry and transportation sectors have potentially large radiative forcing impacts, though the balance of effects from the absorbing vs scattering aerosols leads to net changes near zero in most regions for this scenario. Exceptions are changes to industry emissions in S. Asia and the U.S. having a net positive DRF, and transportation having a positive DRF in China and negative DRF in Europe and the Middle East. An additional factor contributing to the

²<http://themasites.pbl.nl/en/themasites/image/background/regions/index.html>

total aerosol DRF is the impact of agricultural NH_3 emissions in China, the U.S. and S. Asia through promotion of secondary inorganic aerosol formation. We note this is more substantial than commonly recognized (1) for a few reasons. The NH_3 forcings effectively include the impacts of aerosol nitrate, as excess ammonia levels promote ammonium nitrate formation. This may be exaggerated in the GEOS-Chem adjoint, which presently does not account for formation of sodium nitrate or uptake of nitric acid by dust. Secondly, as discussed earlier, NH_3 DRFs based on sensitivities calculated around present-day conditions do not account for any reductions in availability of sulfate and nitrate with which to form aerosol, and as such are likely upper estimates in many regions.

4.2 Comparison across scenarios

We next consider how influences from individual regions, sectors and species govern the aerosol DRFs across RCP 8.5, 6.0, 4.5 and 2.6, estimated here to have global DRFs of -0.07, -0.08, 0.05 and -0.02 Wm^{-2} , respectively. This multi-scenario analysis is afforded through repeated application of the high-resolution DRF efficiencies without the need for numerous additional forward model calculations. The variability in the DRFs across each RCP within individual regions is provided in Fig. S5. Previous works have noted the potential benefit of reducing transportation emissions (7, 57), though we find that for RCP 4.5, the regional analysis of which was shown previously in Fig. 5, the impacts of transportation BC in regions such as the US is negated by co-reduction of reflective aerosols. However, we find a net negative forcing from the transportation sector following the three other RCPs in the US, Canada, Europe, Russia, and Japan, thus underscoring the importance of considering multiple future scenarios. The largest variability between RCPs occurs for the biomass burning DRFs; the domestic and industry emissions from China also have large DRFs which are highly variable across RCPs. Energy and agricultural DRFs are

consistently positive and negative, respectively, across nearly all regions and RCPs.

The total global DRFs from each sector and their contributions by species are shown in Fig. 5. DRFs from the industrial sector are largest in magnitude for the lowest greenhouse gas scenario (RCP 2.6) for BC, OC and SO₂. In contrast, BC DRFs for the transportation and domestic sectors are smaller in RCP 2.6 than RCP 8.5. RCP 6.0 has smaller DRFs for BC, OC and SO₂ relative to the other scenarios. The total global DRF from SO₂, BC and OC throughout many of the sectors combines to yield a net DRF close to zero, with the exception of energy and forest fire emissions. A significant driver of the overall DRF across all sectors is NH₃ from agricultural emissions, modulation of which contributes to RCP 8.5 having a negative DRF and RCP 4.5 a positive DRF. These are again likely uppers estimate of NH₃ DRF.

5 Discussion

The results presented here are limited in several regards. The DRF efficiencies come from a single modeling framework, and thus do not reflect uncertainty in the underlying model and emissions (58–61), the former aspect alone can have a large impact on aerosol DRF even under unified emissions (62). Application of yearly average RCP emissions does not account for key seasonalities, such as the biomass burning season in boreal areas being out-of-phase with seasons of peak albedo. Ignoring the potential for organic aerosol to form semi-absorptive (i.e., brown) particles, in addition to secondary formation of organic aerosol, is a further simplification. The GEOS-Chem model’s simple bulk-partitioning aerosol scheme assumes an external aerosol mixing state, which will not account for inter-species impacts on aerosol lifetime (63). Additionally, these results consider only direct aerosol effects, not accounting for indirect effects or impacts on greenhouse gases.

Without disregard for these caveats, we can still emphasize insights gleaned from

evaluating radiative impacts at an unprecedented level of spatial, species and sectoral detail. That the DRF efficiencies ($\text{Wm}^{-2}/\text{emission}$) calculated for SO_2 , BC, OC and NH_3 in each model grid cell demonstrate considerable variability at a spatial resolution much finer than previously considered in regionally-based analysis is itself noteworthy, particularly given the commensurate variability in estimates of future emissions changes. Variations in the DRF efficiency can be considered a measure of the inequity of any mitigations strategy enacted on a per-emissions basis; such inequities have been shown here to be quite large for several sectors and species. Accounting for these variations, we assess the 2050 aerosol DRFs relative to present for multiple RCPs. The resulting range of DRFs indicates only a weak synergy between aerosol DRF and long-term radiative forcing targets attained primarily based on greenhouse gas targets. The balance of impacts from absorbing vs scattering aerosols is highly dependent upon sector, region and, in cases such as transportation emissions, which future scenario is considered. Further, inclusion of continually increasing NH_3 emissions can significantly offset reductions of SO_2 emissions (via replacement of ammonium sulfate with ammonium nitrate).

Efforts to mitigate impacts from individual species (64, 65) must thus be optimally targeted with regard to co-emitted species. Resolving radiative forcing at resolutions closer to that at which emissions controls are actually enacted makes analysis of realistic policy options more viable, thereby affording inclusion of SLCF effects into the design of future scenarios.

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7 Supporting Information Available

Supporting Information is available for details of the adjoint sensitivity validation and additional estimates of radiative forcings for pre-industrial to present and future scenarios.

This information is available free of charge via the Internet at <http://pubs.acs.org/>.

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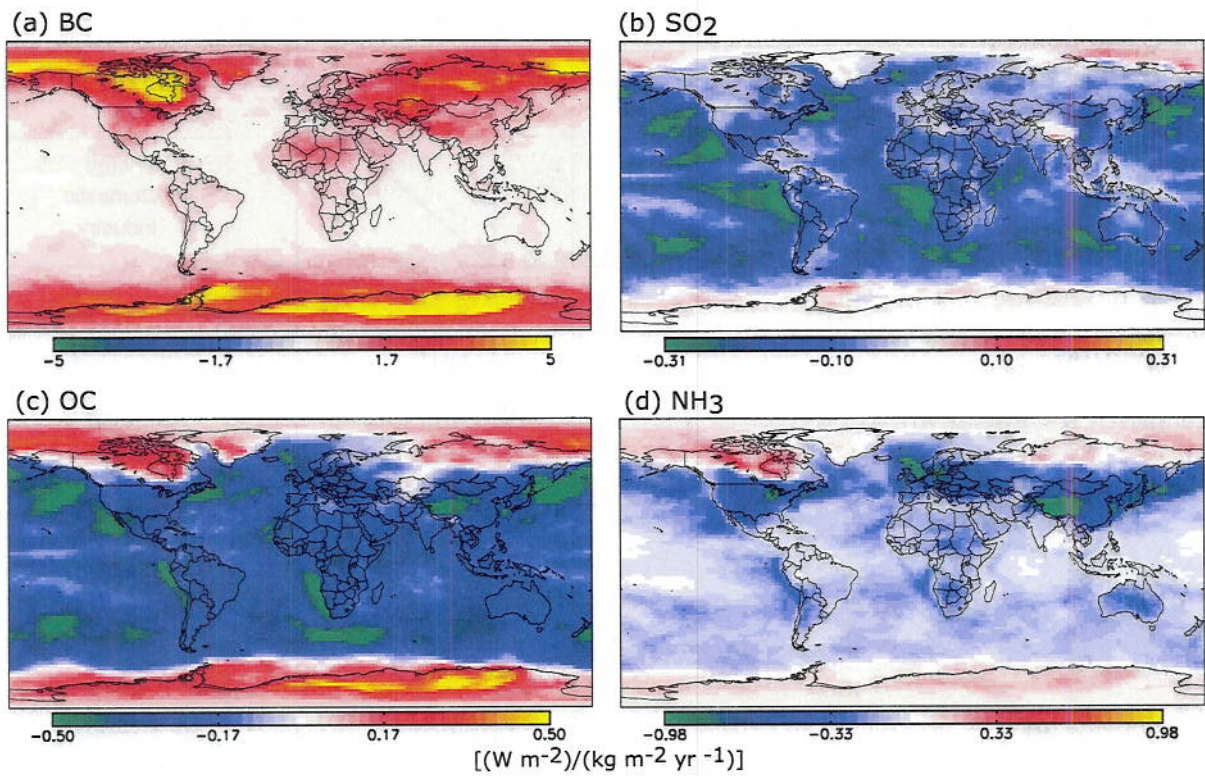


Figure 1: Yearly average radiative forcing efficiencies for (a) BC, (b) SO₂, (c) OC and (d) NH₃. Values in a particular grid cell show the response of global aerosol DRF to perturbations of emissions in that grid cell.

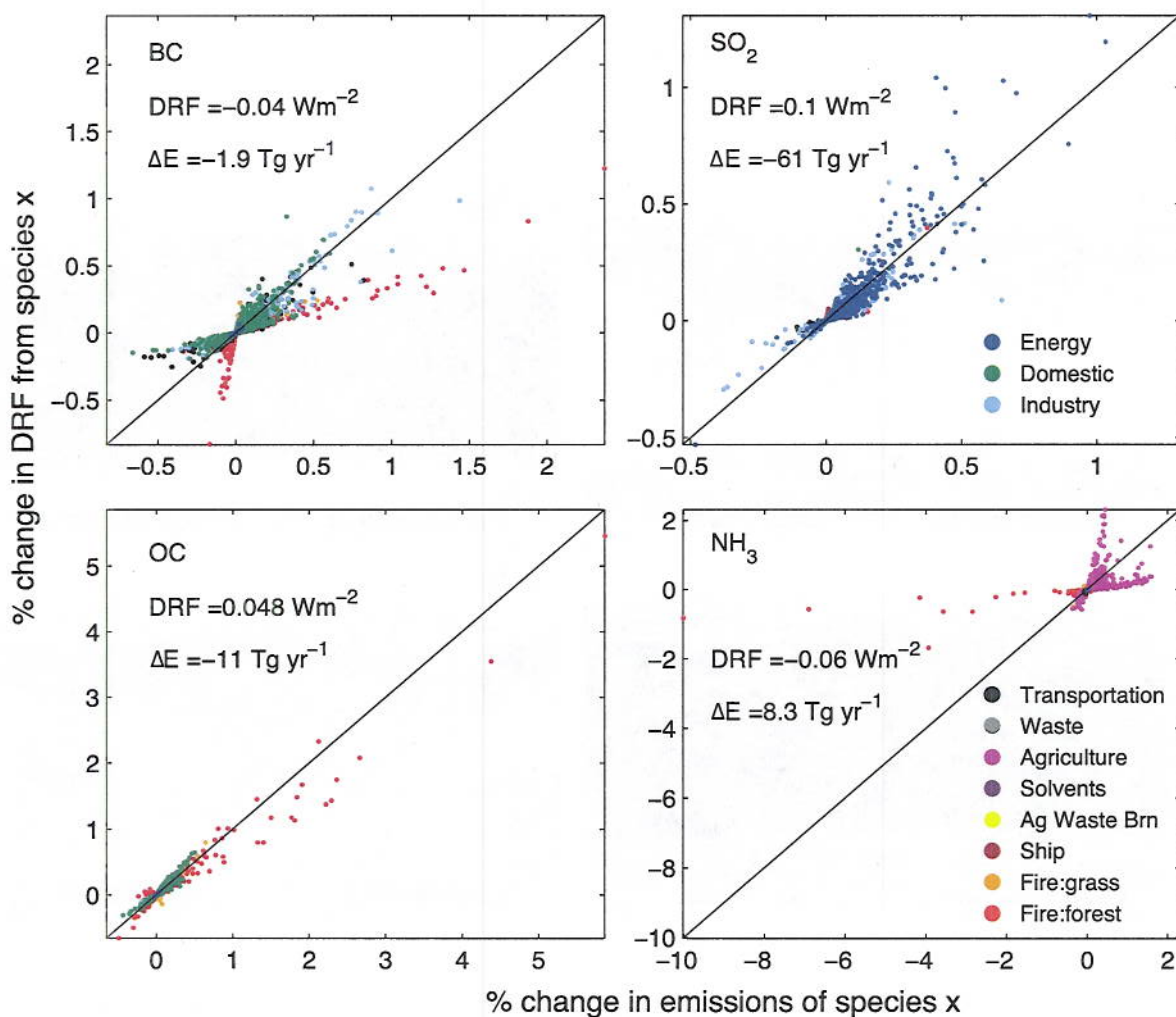
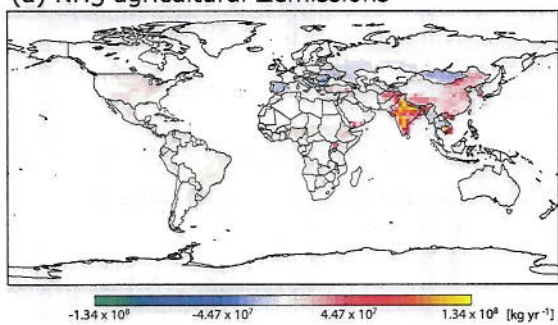


Figure 2: For RCP 4.5 in 2050 relative to 2000: percent change in emissions for a specific species in each grid cell (x-axis) vs the corresponding percent change in the aerosol DRF for that specific species alone (y-axis). Each panel also shows the DRF and change in emissions (ΔE) on which these percentages are based. Note that positive percentages indicate enhancements to the absolute magnitude of net DRF or ΔE , while negative percentages indicate changes in the direction opposite the net DRF or ΔE .

(a) NH_3 agricultural Δ emissions



(b) DRF for NH_3 agricultural emissions

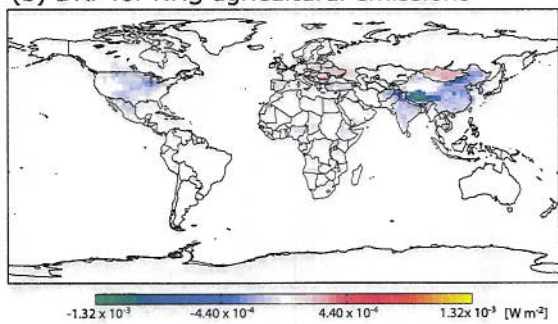


Figure 3: Left: changes to NH_3 agricultural emissions for RCP 4.5 in 2050 relative to 2000. Right: corresponding aerosol DRF for emissions change in each grid cell.

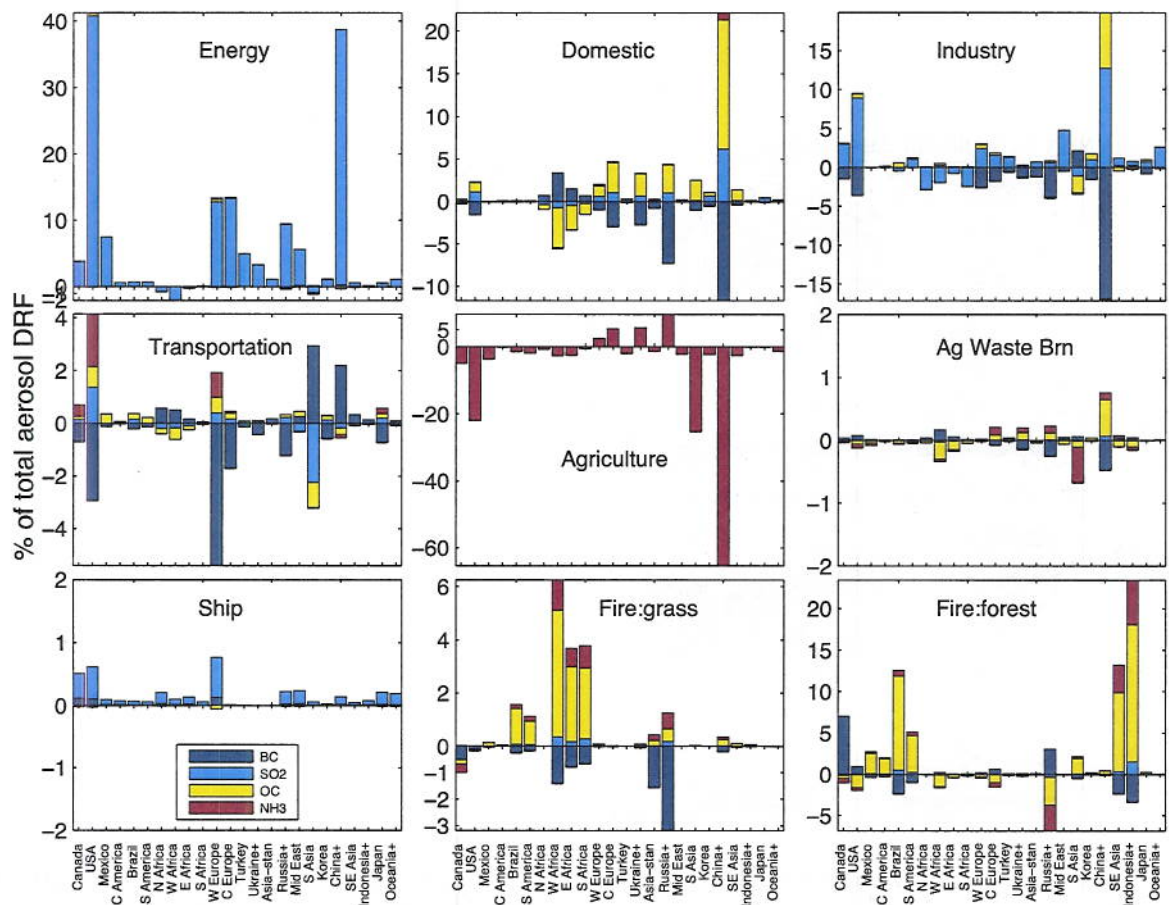


Figure 4: Aerosol DRF of RCP 4.5 in 2050 relative to 2000, for each emission sector and species, lumped by region. Percentages show the contribution of changes to emissions in each region to the total DRF of 0.05 W m^{-2} .

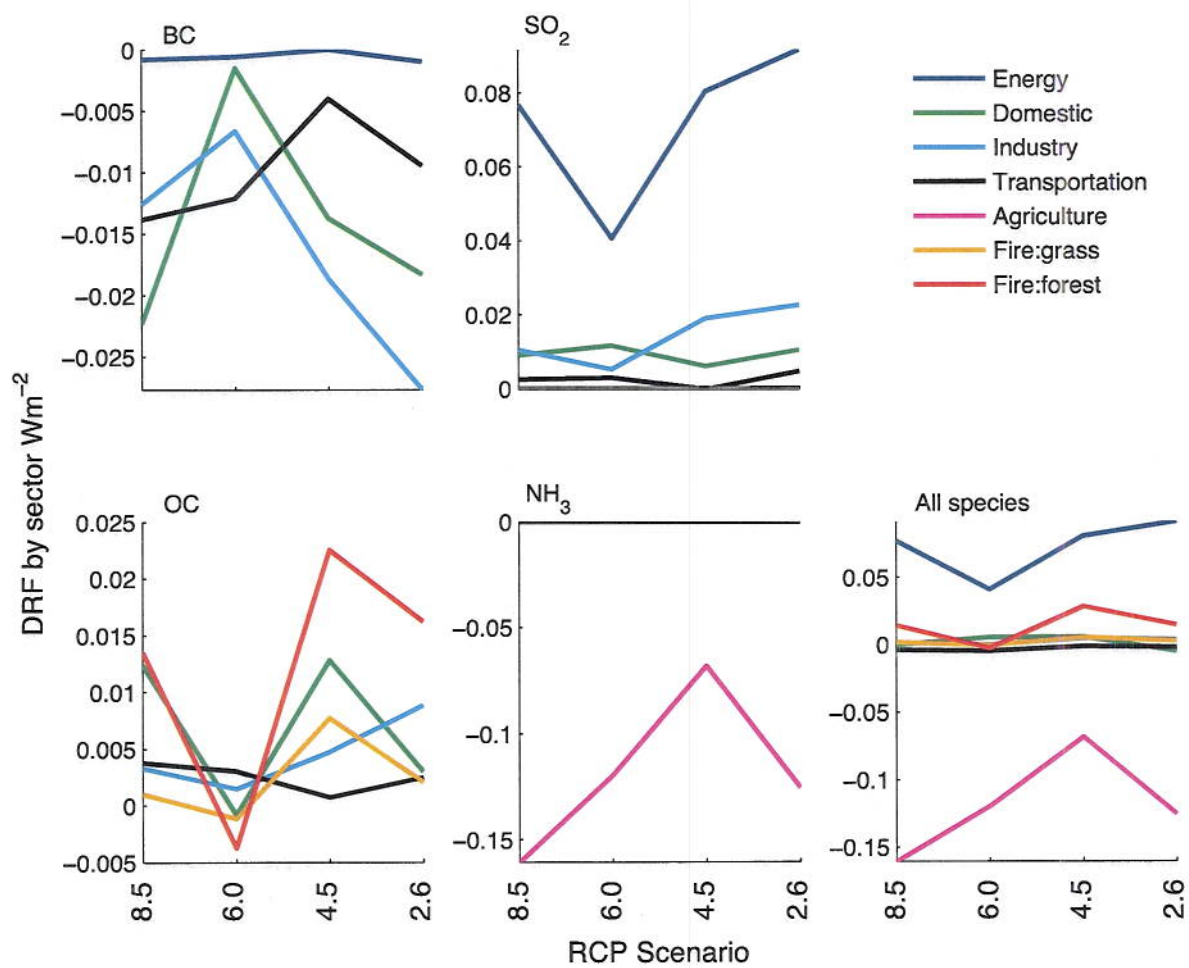


Figure 5: DRF from each sector plotted by species for RCP 8.5 through RCP 2.6, separated by species, and the total of all species.

