

1 Use of fluorinated polybrominated diphenyl ethers and simplified cleanup for the analysis of
2 polybrominated diphenyl ethers in house dust

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11 **Abstract**

12 A simple, cost-effective method is described for the analysis of polybrominated diphenyl ethers
13 (PBDEs) in house dust using pressurized fluid extraction, cleanup with modified silica solid
14 phase extraction tubes, and fluorinated internal standards. There are 14 PBDE congeners
15 included in the method, some typically contained in the commercial mixtures used as flame
16 retardants, and some which are not routinely reported in the peer-reviewed literature. A gas
17 chromatographic-mass spectrometry instrumental method provides baseline separation in < 20
18 minutes, detection limits <20 ng/g, and quantitation limits <60 ng/g for most congeners. Method
19 blanks contained an average concentration < 9 ng/g for all congeners except BDE209 which had
20 an average around 40 ng/g. Spiked samples showed good accuracy with relative percent
21 difference (RPD) <7%, and good precision with relative standard deviation <22% for all
22 congeners except BDE209. The method was applied to the analysis of a standard dust (NIST
23 Standard Reference Material 2585) and showed good accuracy with RPD <25% except for
24 BDE154. Overall, this method exhibited good performance characteristics in all categories
25 including simplicity, cost, sensitivity, selectivity, accuracy, and precision.

26 **Highlights**

- 27 • Using F-PBDEs saves over \$400 per standard compared to ¹³C-labeled.
- 28 • Surrogate recovery was over 80% in performance standards.
- 29 • Relative percent difference (RPD) was generally <7% in spiked samples.
- 30 • Concentrations found in SRM were generally within 20% of reported value.

31

32 **Keywords:** Flame retardants; polybrominated diphenyl ethers; house dust; fluorinated PBDEs;
33 solid phase extraction

34

35 **1. Introduction**

36 Polybrominated diphenyl ethers (PBDEs) are among a class of brominated flame retardants
37 (BFRs) that have widely been used in consumer products. In a typical home, PBDEs can be
38 found in electronic products, textiles such as mattresses and carpets, and furniture. PBDEs are
39 typically additive flame retardants, meaning that they are physically bound to the substrate. Since

40 they are not chemically bound, PBDEs tend to migrate from the product into the indoor
41 environment [1], particularly to dust which is a substantial source of exposure [2-5]. PBDEs are
42 of concern because of potential health impacts including disruption of thyroid hormones [6],
43 neurodevelopmental consequences [7-9] and endocrine disruption [10,11]. While the PBDEs
44 have been or will be removed from U.S. products due to growing concerns about potential health
45 risks[12,13], products containing these chemicals will remain in household use for the
46 foreseeable future. Thus it is important to continue adding and improving the methods for
47 assessing the presence of these chemicals.

48 There are several important issues that complicate the study of PBDEs in house dust (e.g., matrix
49 complexity, range of physical chemical properties). To facilitate the analysis of PBDEs, sample
50 extracts are cleaned up to isolate analytes from the other components of the dust. These
51 procedures tend to be time consuming and use complex cleanup columns. Therefore, this work
52 developed simple yet effective methods for the cleanup using commercial SPE tubes modified to
53 improve their performance.

54 Alternatives to the ¹³C-labeled PBDE internal standards are needed because the labeled PBDEs
55 are relatively expensive and have analytical challenges, particularly ion selection issues and
56 breakdown during analysis [14]. Gas chromatographic mass spectrometry (GC/MS) with
57 negative chemical ionization (NCI) is the method of choice for the analysis of PBDEs in
58 environmental samples. However, labeled PBDEs cannot be used in the NCI analysis because for
59 most congeners, the 79/81 bromide ions are used for quantitation for sensitivity reasons and no
60 differentiation can be made between the labeled and unlabeled congeners. If different analytical
61 instrumentation, ionization modes, or NCI ion selections were made to allow the use of
62 isotopically labeled standards, it is very likely the sensitivity would be diminished and/or the cost
63 of analysis would increase. In the case of higher brominated congeners where alternate ion
64 selection can differentiate between labeled and unlabeled molecules, we have found that there is
65 not a significant methodological advantage over less expensive alternative internal standards.
66 The use of fluorinated PBDEs reduces or eliminates degradation of highly brominated internal
67 standards to lower brominated BDEs generated during analysis. In this work we demonstrate the
68 analysis of 14 PBDEs using three fluorinated PBDEs (F-PBDEs) as internal standards.
69 Fluorinated PBDEs have different retention time than the parent PBDE from which it was
70 derived, and are less costly than labeled PBDEs.

71

72 **2. Materials and methods**

73 A variety of methods have been recently published for the extraction (Soxhlet [15], PFE [16],
74 sonication [17]), cleanup (manually packed columns [18], SPE [19], on-line [20], in-cell [21]),
75 and surrogate/internal standards (¹³C-BDE(s) [22], native BDEs [23], F-BDEs [24], non-BFR
76 compounds [25]), and instrumental analysis (GC-EI [18], GC-ECNI [26], GC-ECD [27], GC-
77 MS/MS [28], LC-MS/MS[29]) for different combinations of PBDE congeners in dust. The
78 method presented here is similar to that described by Stapleton, et al. [30] in that both use
79 pressurized fluid extraction (PFE; ASE 200; Dionex Corp., Sunnyvale, CA), commercially
80 available solid phase extraction (SPE) cartridges, analysis by GC/MS with NCI, and MCDE 86L
81 as a surrogate (recovery) standard. In contrast, this method decreases the volume of
82 dichloromethane, pressure and temperature used for extraction; uses modified SPE cleanup
83 cartridges, a thinner film in the GC column with Guard column and heated injection; adds both

84 surrogate recovery and internal standards; and quantitates different congeners and monitors
85 different ions in the mass spectrometer. The exact conditions for this method are described below
86 and are displayed in Figure 1A-C.

87

88 **2.1 Extraction**

89 Figure 1B shows the assembly of the PFE cells to extract house dust mixed with Ottawa sand
90 (Fisher Scientific, Fair Lawn, NJ). 200 ng of MCDE 86L (Wellington Laboratories, Ontario,
91 Canada) and PBDE 181 (Cambridge Isotope Laboratories, Andover, MA) surrogate recovery
92 standards (SRS) were added to the dust prior to extraction. Table 1 details which SRS was used
93 for each measured congener. Each sample was extracted twice and collected in separate 60 mL
94 vials that were later combined. Additional extraction details are included in Figure 1A.

95

96 **2.2 Cleanup**

97 Sample cleanup was accomplished using two modified 3-mL, SPE cartridges (Sigma-Aldrich, St.
98 Louis, MO) in tandem. The bottom SPE cartridge was modified by adding 500 μ L 95-98%
99 sulfuric acid:water (1:1) and was used without drying. The top SPE cartridge was modified as
100 shown in Figure 1C. The SPE cartridges were flushed three times with 2 mL of
101 hexane:dichloromethane (4:1). The concentrated sample extract was loaded onto the top
102 cartridge, and eluted as shown in Figure 1A. After concentration to 1 mL, 500 ng of F-BDEs 69
103 and 160 and 1000 ng of F-BDE 208 internal standards (Cambridge Isotope Laboratories or
104 Chiron AS, Trondheim, Norway) were added to the extract. Table 1 details which internal
105 standard was used for each measured congener.

106

107 **2.3 GC/MS analysis**

108 GC/MS analyses were performed on an Agilent Technologies (Santa Clara, CA) 6890N GC
109 equipped with a model 5973 inert MS. The GC column specifications (Agilent Technologies and
110 Restek, Bellefonte, PA), GC temperature program and select MS conditions are shown in Figure
111 1A. Helium carrier gas was used at a constant flow rate of 3.2 mL/minute. Injections were made
112 in the splitless mode with the inlet temperature set at 260 °C. The ion source and quadrupole
113 temperatures were 150 °C and methane reagent gas was used. Table 1 shows the retention time
114 and ions monitored for each congener.

115

116 **2.4 Detection/Quantitation limit determination and calibration**

117 To determine the detection and quantitation limits for the target congeners, eight PFE cells
118 containing 0.5g each of diatomaceous earth were spiked with 0 (blank), 1, 5, 10, 25, 50, 250, and
119 500 ng of each PBDE, respectively. These were prepared in triplicate and taken through the
120 complete analytical procedure.

121 The procedure recommended by The International Conference on Harmonization of Technical
122 Requirements for Registration of Pharmaceuticals for Human Use [31] was followed to
123 determine the detection limit (DL) and quantitation limit (QL) for the method. Using this
124 procedure, $DL = 3.3 * \sigma / S$, where σ is the standard deviation of the y-intercept and S is slope of

125 the calibration curve. The $QL = 10 \cdot \sigma / S$ or $3 \cdot DL$. A plot was prepared of the spiked amount
126 versus the average concentration determined for each congener. The four lowest spiked amounts
127 that produced a response (above the blank) were used to determine S and σ were determined. The
128 DL and QL found for each congener is shown in Table 1.

129 Calibration of the mass spectrometer was conducted at least weekly using the instrument's
130 autotune functionality. The instrument response was calibrated for a low (25-500 ng/mL in
131 hexane) and high (50-1000 ng/mL in hexane) range for 13 PBDEs, and at double the
132 concentration for BDE209. Linear regression, with inverse concentration weighting and without
133 origin forcing, was performed for both the low and high ranges for each sample set. Samples
134 above the high calibration range were diluted and reanalyzed. Concentrations were converted to
135 a mass/mass basis using the quantity of dust extracted.

136

137 **2.5 Quality assurance**

138 A sample run using this method consisted of field collected samples, a method blank, method
139 spike, and 100% recovery spike. For the method spike, a specific concentration of the SRS and
140 PBDE solution are added to 0.5 g diatomaceous earth (DE; used as blank matrix; Sigma Aldrich)
141 prior to the extraction and cleanup process. For the 100% recovery spike, the SRS and the PBDE
142 solution are added to the extract resulting from the extraction and cleanup of the blank DE
143 immediately prior to GC/MS analysis. Standard Reference Material (SRM) 2585 (NIST,
144 Gaithersburg, MD) was also analyzed with the run.

145 Regression lines were considered acceptable with $R^2 \geq 0.98$. Calibration check standards were
146 run at least once per sample set, inserted every 5-6 samples. If check standards were outside the
147 $\pm 25\%$ nominal concentration criteria, the instrument was retuned and recalibrated. For
148 quantification, an experienced analyst considered the retention time, ion presence, and ion ratios
149 to confirm the identity of PBDE analytes of interest; strict performance criteria for these values
150 were not used due to the complexity of the dust sample matrix and chromatograms.

151

152 **3. Results and discussion**

153 This exact combination of cleanup conditions had not previously been reported, and produced
154 good recovery of target analytes and produced purified extracts. Because PBDEs are stable under
155 acidic conditions, acidified silica is used to remove hydrophobic interferences from samples [32].
156 Alumina further eliminates polar interferences, and has the added benefit of separating PBDEs
157 from newer brominated flame retardants such as TBB which can coelute with BDE99 [33].

158 Diatomaceous earth proved to be a good dust surrogate since it is difficult or impossible to find
159 house dust that does not contain PBDEs [25]. A 0.5 g sample of DE is used rather than the 1g of
160 dust because it has a lower density and therefore fills a larger volume. This mass of DE also
161 proved to be sufficient to absorb spikes.

162 Good baseline separation was achieved for all congeners using the guard and 15 meter DB-5MS
163 columns (Figure 2A), and showed minimal interferences for a SRM 2585 housedust sample
164 (Figure 2B). As previously reported, the instrument responses decreased with the increasing
165 PBDE masses, in part due to thermal breakdown in the injector and column [14]. It is likely that
166 the response of the higher PBDEs could be improved with the use of a programmable

167 temperature vaporization inlet, on column injection, and shorter or narrow bore columns [14].

168

169 Table 2 shows method performance data for method blanks (N = 32), spikes (N = 17) and 100%
170 recovery (N = 17) samples. All concentrations were adjusted based on the recoveries of the two
171 surrogate recovery standards. MCDE 86L was used for the correction of BDEs 47 through 183
172 while BDE181 was used for the correction of BDEs 190 through 209. Percent recovery of SRS
173 in these 83 quality control samples (including 17 SRMs in Table 3) was $87.8 \pm 23.4\%$ for
174 MCDE-86L (mean \pm standard deviation) and $81.9 \pm 21.8\%$ for BDE181.

175 The blank samples contained a maximum of 28.4 ng/g for individual PBDE congeners, excepting
176 BDE 209 (Table 2). For BDE209, the concentration in blanks averaged 39.2 ± 35.6 ng/g and was
177 found in all but four samples, which may reflect the ubiquitous nature of this compound. For
178 BDEs 190 and 206, no detectable concentrations were found in any blank sample. The 100%
179 recovery and method spikes were amended with 300 ng/g of each congener except BDE209 for
180 which the concentration was doubled. The 100% recovery spikes had mean concentrations
181 within 10% of the actual concentrations as shown by the relative percent difference (RPD) in
182 Table 2, absolute standard deviation (SD) less than 65 ng/g for most congeners and relative
183 standard deviations (RSD) $< 34\%$. The method spikes had mean concentrations within 13% of
184 the actual concentrations, absolute SD less than 55 ng/g for most congeners and RSD $< 20\%$ for
185 most congeners. In all cases, the RPD and SD were largest for BDE209. The decreased accuracy
186 and precision for BDE209 are likely due to previously mentioned analytical difficulties with this
187 compound, including the likelihood of inadvertent contamination, photo and thermal degradation
188 [14]. The performance of this method was considered very good for all aspects of laboratory
189 created samples.

190 Table 3 shows the concentrations measured from the analysis of SRM 2585 (House Dust) used to
191 evaluate this method. The seventeen SRM samples were analyzed over a year period during the
192 analysis of environmental dust samples (to be reported separately). The precision of the method
193 can be assessed by the RSD of the repeat measurements. The RSD ranges from 16 to 43% which
194 is reasonable for an environmentally derived sample. The results were in close agreement with
195 the SRM certified values [34] with the relative percent difference (RPD) for individual
196 congeners ranging from 2.5 to 35%, with six of the nine congeners below 15%, showing good
197 accuracy of the method.

198 Table 3 also shows SRM 2585 results using other methods reported in the literature, grouped by
199 instrumental method. The RSD for this method was usually higher than other methods reported
200 by a few percent (0-15%), however, for this study N = 17 over a year period, while in others N =
201 3-6. Using this method, the RPD for all congeners except BDE154 is in the same range as other
202 methods. Two congeners do not have a NIST reported reference value (BDEs 197 and 207). For
203 BDE 197 our mean concentration (28.5 ng/g) was within the range of other reported
204 concentrations (10-33 ng/g) and for BDE 207 our mean concentration was quite similar to the
205 one other reported value (124 vs 120 ng/g).

206

207 **4. Conclusion**

208 A simple and cost effective method is presented for the analysis of PBDEs in house dust
209 samples. This method makes use of F-PBDEs as internal standards, which we have demonstrated

210 to be just as effective as isotopically labeled standards. For 1 mL of a 50 µg/mL solution of ¹³C₁₂
211 BDE209, the cost is \$725; fluorinated BDE standards in the same volume and concentration cost
212 \$295, a savings of over \$400 per standard. Also, the method uses a dust cleanup procedure using
213 modified commercial SPE cartridges that is simpler, more consistent, and requires less time than
214 hand-packed, and/or multiple column procedures. Only BDEs 99, 118 and 209 had average blank
215 concentrations above the detection limit (none were above the quantitation limit). Spiked
216 laboratory samples had good accuracy and precision. The method was also used to measure the
217 PBDE concentration of a dust standard reference material. The results show the method
218 produced results in the expected range of values for SRM and compared well with other reported
219 values. Results from the full study of house dust present in Southern California homes will be
220 presented in future papers.

221

222 Disclaimer

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226

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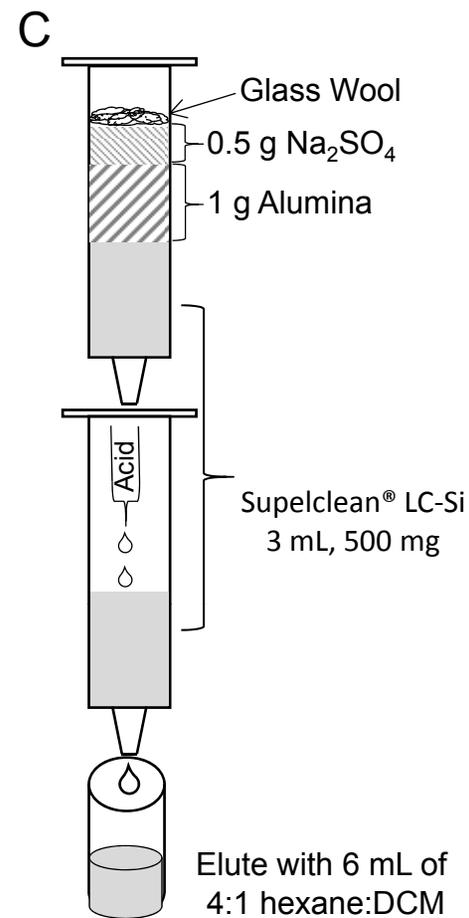
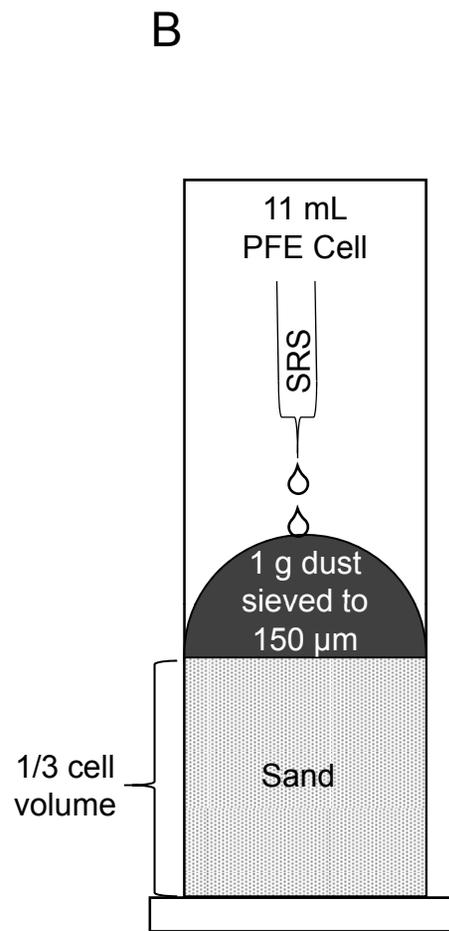
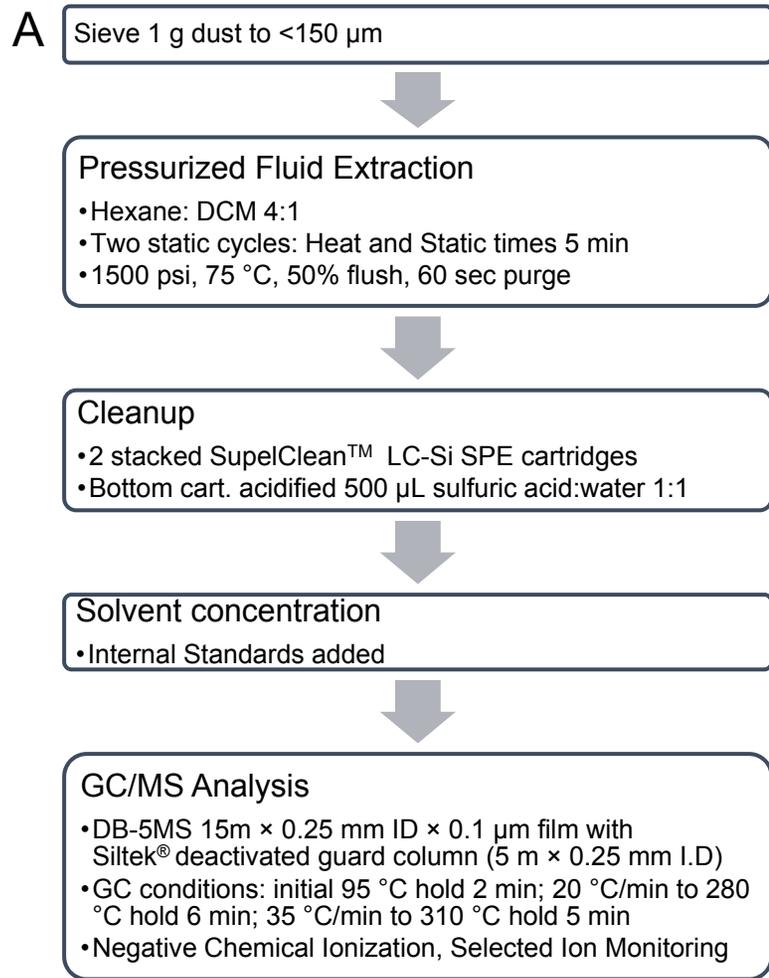
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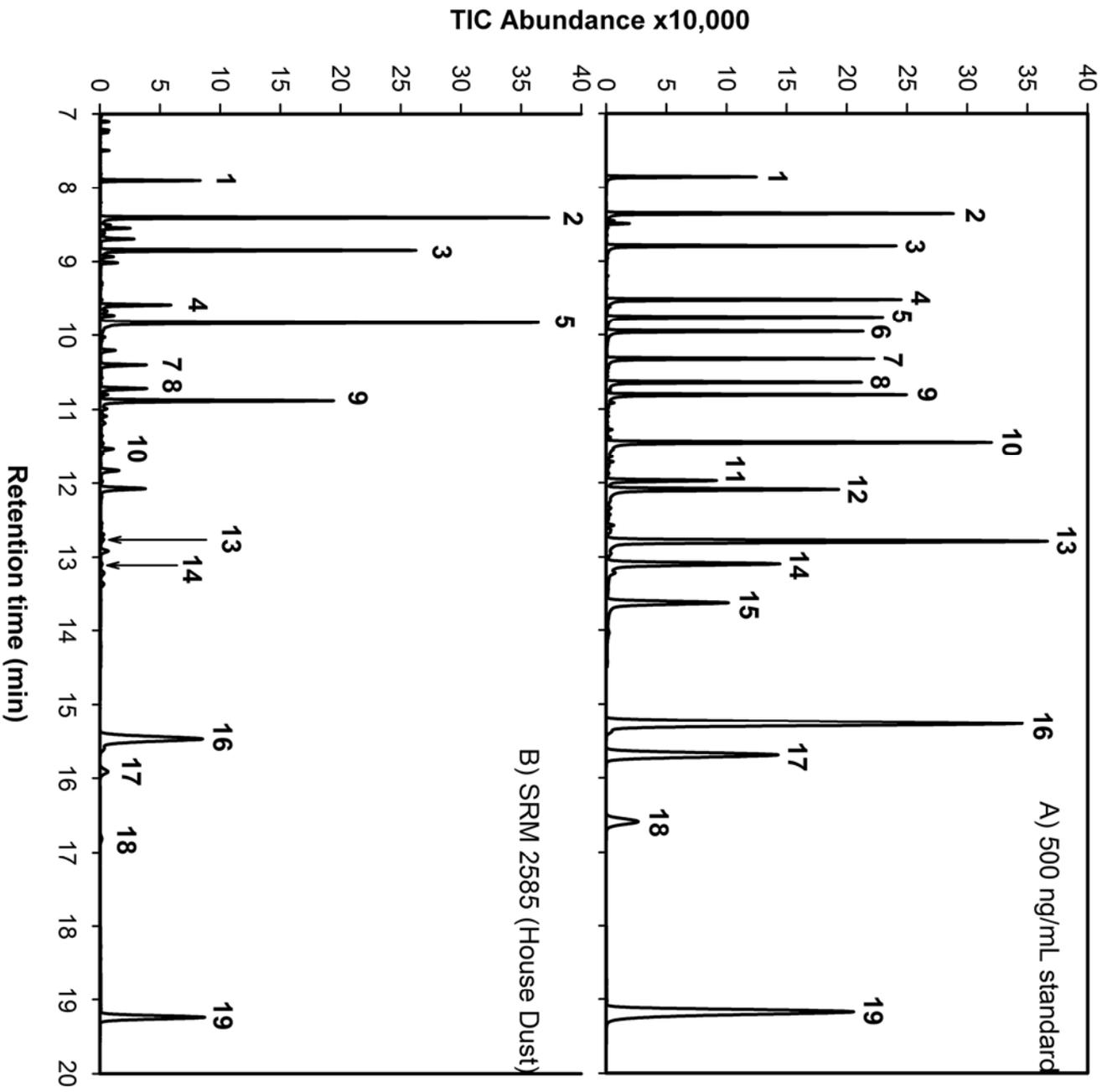
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293 Figure 1. Analytical procedure flow diagram (A) and drawings of extraction (B) and cleanup (C)
294 steps.

295 Figure 2. Total ion chromatogram (using SIM conditions noted in Table 1) of a 500 ng/mL
296 standard (A) and SRM 2585 House Dust extract (B) with internal and surrogate standards.

297 Compound identification: (1) MCDE 86L, (2) F-BDE69, (3) BDE47, (4) BDE100, (5) BDE99,
298 (6) BDE118, (7) BDE154, (8) BDE153, (9) F-BDE160, (10) BDE183, (11) BDE181, (12)
299 BDE190, (13) BDE197, (14) BDE203, (15) BDE205, (16) F-BDE208, (17) BDE207, (18)
300 BDE206, (19) BDE209





Use of fluorinated polybrominated diphenyl ethers and simplified cleanup for the analysis of polybrominated diphenyl ethers in house dust

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Table 1. Retention times, ions monitored, detection (DL) and Quantitation Limits (QL) for PBDE analysis.

Congener	Ret. time (min)	Target ion (<i>m/z</i>)	Qualifier ion #1 (<i>m/z</i>)	Qualifier Ion #2 (<i>m/z</i>)	DL ng/g	QL ng/g
<i>Target analytes</i>						
BDE47 ^{b, d}	8.8	81.0	79.0	161.0	20	59
BDE100 ^{b, d}	9.6	81.0	79.0	161.0	7	20
BDE99 ^{b, d}	9.8	81.0	79.0	161.0	32	96
BDE118 ^{b, d}	10.0	81.0	79.0	161.0	2	6
BDE154 ^{b, d}	10.4	81.0	79.0	562.0	6	17
BDE153 ^{b, d}	10.7	81.0	79.0	562.0	6	19
BDE183 ^{b, e}	11.5	81.0	79.0	562.0	3	9
BDE190 ^{c, e}	12.1	81.0	79.0	562.0	10	31
BDE197 ^{c, e}	12.9	409.7	79.0	81.0	7	21
BDE203 ^{c, e}	13.2	81.0	79.0	561.7	6	18
BDE205 ^{c, e}	13.8	81.0	79.0	561.7	7	20
BDE207 ^{c, f}	15.9	486.6	488.6	— ^a	8	24
BDE206 ^{c, f}	16.8	486.6	488.6	641.6	125	378
BDE209 ^{c, f}	19.3	488.6	486.6	—	26	78
<i>Internal standards and surrogates</i>						
F-BDE69	8.4	81.0	79.0	161.0		
F-BDE160	10.9	81.0	79.0	502.0		
F-BDE208	15.4	486.6	488.6	—		
MCDE 86L	7.9	318.0	354.0	—		
BDE181	12.0	81.0	79.0	562.0		

^a No additional qualifier ion used

^b Compounds corrected by MCDE 86L surrogate standard

^c Compounds corrected by BDE181 surrogate standard

^d Compounds quantitated by F-BDE69 internal standard

^e Compounds quantitated by F-BDE160 internal standard

^f Compounds quantitated by F-BDE208 internal standard

Table 2. Method performance for blank (N = 32), 100% recovery (N = 17), and spiked (N = 17) samples (300 ng/g except BDE209 at 600 ng/g) including mean \pm standard deviation (SD; ng/g), range (ng/g), relative standard deviation (RSD) and percent difference (RPD) from the nominal spike concentration.

Congener	Method Blank	100% Recovery Spike		Method Spike	
	Mean \pm SD (min-max) ng/g	Mean \pm SD (min-max) ng/g	RPD RSD %	Mean \pm SD (min-max) ng/g	RPD RSD %
BDE47	5.45 \pm 4.16	287 \pm 48.3	4.26	286 \pm 37.7	4.69
	0.52-16.6	153-392	16.8	205-383	13.2
BDE99	7.80 \pm 7.06	284 \pm 51.5	5.31	293 \pm 39.2	2.44
	0-28.2	157-382	18.1	211-376	13.4
BDE100	3.44 \pm 4.28	281 \pm 56.3	6.48	280 \pm 53.9	6.80
	0-19.6	157-389	20.1	170-393	19.3
BDE118	5.28 \pm 5.28	290 \pm 48.2	3.18	296 \pm 39.4	1.47
	0-24.6	155-382	16.6	207-374	13.3
BDE154	2.07 \pm 3.70	292 \pm 47.7	2.73	292 \pm 41.5	2.82
	0-19.5	161-386	16.3	217-391	14.2
BDE153	2.96 \pm 4.79	290 \pm 48.4	3.48	286 \pm 40.5	4.83
	0-21.2	158-378	16.7	211-380	14.2
BDE183	1.76 \pm 2.67	298 \pm 55.0	0.78	286 \pm 39.5	4.79
	0-13.6	144-402	18.5	192-360	13.8
BDE190	0 \pm 0	291 \pm 46.0	2.90	288 \pm 35.7	3.88
	0	140-349	15.8	187-329	12.4
BDE197	1.89 \pm 5.77	283 \pm 52.2	5.60	284 \pm 44.4	5.22
	0-28.4	132-392	18.4	179-376	15.6
BDE203	0.79 \pm 2.20	287 \pm 54.4	4.47	282 \pm 44.9	6.02
	0-9.05	121-378	19.0	160-350	15.9
BDE205	0.31 \pm 1.73	284 \pm 59.4	5.46	275 \pm 49.5	8.34
	0-9.79	111-365	20.9	146-329	18.0
BDE206	0 \pm 0	295 \pm 65.3	1.52	289 \pm 53.8	3.60
	0	136-418	22.1	201-424	18.6
BDE207	1.47 \pm 4.51	298 \pm 56.6	0.63	292 \pm 49.2	2.60
	0-23.4	143-378	19.0	193-392	16.8
BDE209	39.2 \pm 35.6	538 \pm 181	10.3	524 \pm 202	12.7
	0-172	151-917	33.6	200-969	38.5

Table 3. BDE congener mean \pm standard deviation or range of means (ng/g; first line) with Relative Standard Deviation (second line) and Relative Percent Difference (third line) for various methods vs. NIST certified value for SRM 2585.

BDE	This method ^a	GC-ECNI-MS [1-5]	GC-EI-MS [6,7]	GC-MS/MS [8-10]	GC-ECD [11,12]	LC-NI-APPI-MS/MS [13]	NIST [14]
47	484 \pm 77.9	390-520	445 \pm 29	486-526	537 \pm 26	561 \pm 36	497 \pm 46 ^d
RSD ^b	16.1	3.2-12	6.5	2.1-9.5	4.8	6.4	
RPD ^c	2.55	0.20-22	10	0.60-5.8	8.0	13	
99	858 \pm 133	680-1110	838 \pm 67	803-952	890 \pm 36	883 \pm 26	892 \pm 53 ^d
	15.5	5.6-13	8.0	5.6-6.8	4.0	2.9	
	3.85	2.1-24	6.1	1.9-10	0.22	1.0	
100	149 \pm 34.0	110-158	140 \pm 13	135-147	160.7 \pm 5.6	143 \pm 5	145 \pm 11 ^d
	22.8	3.3-13	9.3	3.4-8.9	3.5	3	
	2.72	3.7-24	3.4	0.69-6.9	11	1.4	
154	113 \pm 22.1	70-95	87.8	77-81	96 \pm 12	77.4 \pm 4.2	83.5 \pm 2.0 ^e
	19.5	7.9-12	8.9	6.2-13	13	5.4	
	35.3	5.6-16	5.1	3.0-7.8	15	7.3	
153	131 \pm 24.6	90-124	118	118-133	127.1 \pm 8.8	132 \pm 6	119 \pm 1 ^d
	18.7	1.6-13	8.5	5.7-14	6.9	5	
	10.3	0-24	0.84	0.84-12	6.8	11	
183	40.0 \pm 11.4	25-62		44 \pm 4	42.5 \pm 5.2	43.3 \pm 4.7	43 \pm 3.5 ^d
	28.6	3.2-17	NR ^f	9.1	12	11	
	6.83	1.6-44		2.3	1.2	0.70	
197	28.5 \pm 9.22	10-33					NR ^f
	32.3	20-61	NR	NR	NR	NR	
	NA ^g	NA					
203	29.2 \pm 8.01	10-20.6			41.3 \pm 4.8		36 \pm 6.4 ^e
	27.4	0-12	NR	NR	12	NR	
	19.0	43-72			15		
206	204 \pm 59.0	157-159	188				271 \pm 42 ^e
	28.9	19-21	7.4	NR	NR	NR	
	24.6	41-42	31				
207	124 \pm 23.4	120 \pm 40					NR
	18.9	33.3	NR	NR	NR	NR	
	NA	NA					
209	2670 \pm 1140	2050-2800	2703	2971 \pm 333	2091-2706	2580 \pm 140	2510 \pm 190 ^e
	42.8	8.5-25	5.2	11	2.8-7.2	5.4	
	6.34	1.2-18	7.7	18	7.8-17	2.8	

^a N = 17

^b Relative standard deviation

^c Relative percent difference from the NIST reported value.

^d Certified values are weighted means of the results from four analytical methods. The uncertainty listed with each value is an expanded uncertainty about the mean, with coverage factor 2 (approximately 95% confidence) except for PBDE 153 with a coverage factor of 10, calculated by combining a between-source variance incorporating inter-method bias with a pooled within-source variance.

^e Certified values are unweighted means of the results from two or three analytical methods. The uncertainty listed with each value is an expanded uncertainty about the mean, with coverage factor 2, calculated by combining a between-method variance with a pooled, within-method variance.

^f Not reported

^g Not applicable- RPD cannot be calculated when a reference value is not reported.

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