1 Title:

2 A biomonitoring framework to support exposure and risk assessments

4 Authors:

5 Jon R. Sobus•, Yu-Mei Tan, Joachim D. Pleil, and Linda S. Sheldon

- 7 Address and affiliation of all authors:
- 8 ational Exposure Research Laboratory, Office of Research and Development, U.S.
- 9 Environmental Protection Agency, 109 T.W. Alexander Drive, Research Triangle Park, NC

10 27711, USA

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12 *Corresponding author:

13 Jon R. Sobus, U.S. Environmental Protection Agency, 109 T.W. Alexander Drive, Research 14 Triangle Park, NC 27711, USA, Mail Code: E205-04, email: sobus.jon@epa.gov,phone:(919) 15 541-2232, fax: (919) 541-0905

47 Abstract

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Background: Biomonitoring is used in exposure and risk assessments to reduce uncertainties along the source-to-outcome continuum. Specifically, biomarkers can help identify exposure sources, routes, and distributions, and reflect kinetic and dynamic processes following exposure events. A variety of computational models now utilize biomarkers to better understand exposures at the population, individual, and sub-individual (target) levels. However, guidance is needed to clarify biomonitoring use given available measurements and models.

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56 **Objective:** This article presents a biomonitoring research framework designed to improve 57 biomarker use and interpretation in support of exposure and risk assessments.

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Discussion: The biomonitoring research framework is based on a modified source-to-outcome continuum. Five tiers of biomonitoring analyses are included in the framework, beginning with simple cross-sectional and longitudinal analyses, and ending with complex analyses using various empirical and mechanistic models. Measurements and model requirements of each tier are given, as well as considerations to enhance analyses. Simple theoretical examples are also given to demonstrate applications of the framework for observational exposure studies.

66 **Conclusion:** This biomonitoring framework can be used as a guide for interpreting existing 67 biomarker data, designing new studies to answer specific exposure- and risk-based questions, 68 and integrating knowledge across scientific disciplines to better address human health risks.

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74 Key Words:

biomonitoring; biomarkers; exposure science; exposure assessment; risk assessment

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77 Abbreviations:

78 BR biomarker, biologically-relevant biomarker; BR dose, biologically-relevant dose

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93 **1.0 Introduction**

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95 The U.S. Environmental Protection Agency (USEPA) and other organizations use risk
96 assessments to determine whether actions should be taken to protect public health (USEPA,
97 2009). Risk assessments are based on the concept that:

$Risk = Hazard \times Exposure$

For a given chemical, toxicity testing is used to identify a hazard and to establish a quantitative relationship between administered dose and the incidence of health effects. This dose-response relationship is then used to develop an acceptable human exposure level (*e.g.*, a reference dose [RfD]). Next, exposure assessments identify the source(s), route(s), and magnitude of human exposure. The risk of an adverse outcome is then determined by comparing observed or estimated exposures to the acceptable level. Finally, information on sources and routes of exposure are used to identify effective mitigation strategies.

In the past, blunt tools have generally been used for risk assessments – that is, high-dose animal toxicity tests and screening-level exposure assessments. With these tools, the links between exposure and health outcome are highly uncertain. Biomonitoring, because it is close to the center of the source-to-outcome continuum, should better inform these linkages and reduce the associated uncertainties. However, applications of biomonitoring in exposure and risk assessments are limited by a lack of guidance on data use and interpretation.

This article presents a modified source-to-outcome continuum that provides a framework for biomonitoring to support exposure and risk assessments. The framework includes five tiers that can be used to answer specific exposure- and risk-based questions. The overall intent is to provide guidance for interpreting existing biomarker data, designing new biomonitoring studies

116	to efficiently answer targeted research questions, and synthesizing relevant information across		
117	scientific disciplines to address human health risks.		
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119	2.0 Discussion		
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121	2.1 A modified source-to-outcome continuum		
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123	Figure 1 illustrates a modified source-to-outcome continuum that highlights traditional		
124	components of exposure science and contemporary components of health effects science.		
125	Biomarker measurements are central to the continuum, and therefore link the exposure and		
126	health effects components.		
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128	Definitions of components within the modified source-to-outcome continuum (shown in Fig. 1)		
129	include the following:		
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131	• Environmental measurements are observed concentrations of stressors in environmental		
132	media. While stressors can be biological, physical, or even psychosocial, this article		
133	focuses on chemical stressors.		
134	• Exposure models mathematically combine environmental measurements with human		
135	activities and other exposure factors to generate exposure estimates.		
136	• Exposure estimates are route specific (e.g., inhalation, ingestion, or dermal exposure)		
137	and quantify the mass of a chemical that comes into contact with a human over time.		

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Kinetic models mathematically describe the movement of a chemical through the body;
 that is, its movement into the body (*absorption*) following exposure, *distribution* to
 various tissues, *metabolism* by various processes, and ultimate *elimination* from the body
 (these kinetic processes are collectively called "ADME").

- *Dose estimates* are based on exposure estimates and kinetic processes, and quantify the
 integrated (over time) mass of a chemical inside the body.
- *Biomarker measurements* are observations of chemicals, chemical metabolites, and 145 target molecules (*e.g.*, chemical adducts) in biological media (*e.g.*, blood, breath, and 146 urine) that, for the purposes of this paper, reflect exposure events and dose.
- Biologically-relevant (BR) dose estimates are based on dose estimates and kinetic
 processes, and quantify the amount of the dose at a specific target (inside a human) that is
 associated with key events in a disease process (e.g., the amount of a neurotoxin in the
 brain, or the amount of a genotoxic metabolite that interacts with genetic material). (This
 definition follows from that of "biologically relevant exposures" given by Birnbaum
 (2010)).
- *Dynamic models* mathematically describe the impacts of the BR dose on biological
 systems (*e.g.*, enzyme inhibition from a neurotoxin, or DNA damage from a genotoxic
 metabolite) and are used to predict BR biomarker levels.
- *BR biomarker measurements* are observations of chemicals/molecules in biological
 media that reflect (directly or indirectly) the BR dose (*e.g.*, blood enzyme levels to reflect
 the BR dose of a neurotoxin).
- *Statistical models* compare observations of random variables for hypothesis testing. For example, statistical models can evaluate associations between environmental and

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biomarker measurements, and biomarker and BR biomarker measurements, as well as the effects of covariates (*e.g.*, age, gender, and human activities) on these associations.

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164 Figure 1 shows that components of the modified source-to-outcome continuum align along two 165 planes: (1) measured values (*i.e.*, environmental, biomarker, and BR biomarker measurements) 166 shown with blue boxes; and (2) estimated values (*i.e.*, exposure, dose, and BR dose estimates) 167 shown with red triangles. While measured values are subject to uncertainty based on assay 168 precision, estimated values depend on measurements, observations, and model parameters, and 169 are therefore subject to greater uncertainty. Biomarker measurements, which are at the center of 170 the continuum, can help reduce uncertainties in estimated values as described in the following 171 sections.

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173 Our proposed biomonitoring framework has five tiers that describe different uses for 174 biomarkers with different levels of complexity. Tier 1 considers only biomarker measurements, 175 and subsequent tiers consider additional measurements, models, and estimated values. Simple 176 theoretical examples are given for each tier to demonstrate how biomarker data can be used to 177 answer important exposure- and risk-based questions. Theoretical examples are given, rather 178 than results from published studies, to allow continuity from one tier to the next, and to simplify 179 the interpretation and discussion. Example biomarkers for each tier are assumed to be 180 measurable using reliable sampling and analytical methods, and to reflect exposure to 181 environmental chemicals. The specific criteria for biomarker selection and use are not the focus 182 of this article and can be found elsewhere (Metcalf and Orloff, 2004; NRC, 1987; NRC, 1991; 183 NRC, 2006; Sobus et al., 2010a).

184	2.2	Biomonitoring framework tiers
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186	2.2.1	Tier 1: Biomonitoring for exposure surveillance
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188	Tier 1	analyses of biomarker data aim to answer one or more of the following questions for
189	exposu	re surveillance:
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191	•	Who is exposed?
192	•	What are the exposure trends?
193	•	Which chemicals should be prioritized for higher-tier analyses?
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195	As sho	wn in Figure 2, biomarker measurements are the only requirement for a Tier 1 analysis.
196	Specifi	cally, cross-sectional biomarker measurements are used to evaluate exposures across
197	popula	tions, and longitudinal biomarker measurements are used to evaluate exposure trends
198	within	a population. To demonstrate these uses, two theoretical examples are given in Figure 2.
199		In Example 1, the two distributions represent biomarker measurements that have been
200	separat	ed into groups for hypothesis testing; example groups could include those separated by
201	gender	, geographical area, age (<i>e.g.</i> , < 18 years old vs. ≥ 18 years old), or product use. All other
202	things	being equal, observed differences between grouped measurements indicate an effect of the
203	groupi	ng variable on biomarker levels, and suggest exposure differences between the groups.
204		In Example 2, longitudinal biomarker measurements for a population decrease over time,
205	sugges	ting a decrease in exposure levels. Trends over time can indicate a change in the exposure
206	source	(e.g., de-registration of a consumer product), or a change in human activities through

207	which contact occurs (e.g., product use patterns). However, higher-tier analyses of the
208	biomarker data are generally needed to pinpoint the cause of a trend. For examples of Tier 1
209	analyses, see Barr et al. (2010), Naeher et al. (2010), Calafat et al. (2010), Pirkle et al. (2006),
210	and Sobus et al. (2009b; 2009c).
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212	2.2.2 Tier 2: Biomonitoring to support exposure assessment
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214	Tier 2 analyses of biomarker data can answer the following questions to support exposure
215	assessments:
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217	• What are the likely exposure sources?
218	• What are the likely exposure routes?
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220	As shown in Figure 3, Tier 2 analyses consider paired environmental and biomarker
221	measurements at the subject level, and focus on statistical comparisons of these data. A graph in
222	Figure 3 shows a regression of "spot" biomarker measurements (one observation per subject) on
223	corresponding environmental measurements. A positive linear trend is shown in this example
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with a R^2 value of 0.3. This indicates that biomarker levels increased with increasing environmental levels, and that 30% of the biomarker measurement variance was explained by corresponding environmental measurements.

If, for example, environmental measurements in this example were concentrations of a chemical in food, and biomarker measurements were corresponding blood levels of the same chemical, then the results of the regression analysis would point to dietary ingestion as a likely exposure route. Potentially, the results could point to food or a specific food item as an exposuresource.

232 Considerable unexplained variance in the biomarker data (*i.e.*, 70%), however, would 233 suggest additional exposure routes, and/or considerable covariate effects (e.g., timing of 234 sampling events, gender, age, and ethnicity) on biomarker levels. Therefore, additional data 235 would be necessary to better explain the observed biomarker variance and to further support the 236 exposure assessment. These data could be part of a more complex Tier 2 analysis (e.g., 237 including environmental measurements of different media to identify additional exposure 238 routes), or of a higher-tier analysis. For examples of Tier 2 analyses, see Egeghy et al. (2005), 239 Thomas et al. (2010a; 2010b), and Sobus et al. (2009a)

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241 **2.2.3** Tier 3: Biomonitoring to support risk assessment

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Tier 3 analyses of biomarker data can be used to support risk assessments since they can answerthe following questions:

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• What are the important exposure factors?

• What are the likely exposure levels?

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The requirements of a Tier 3 analysis of biomarker data are shown in Figure 4, and build on the Tier 2 parameters by adding considerations for exposure. Here, exposure is directly linked to environmental measurements via exposure models, and indirectly linked to biomarker measurements via statistical models (*e.g.*, multiple regression models) that consider environmental measurements, human activities, and other exposure factors. Statistical models are used to identify important predictors of exposure and can therefore inform exposure calculations. Since exposure estimates are comparable to acceptable levels based on animal studies, Tier 3 analyses can place biomarkers into a risk context.

257 In our Tier 2 regression example (Figure 3), we showed how measurements of a chemical 258 in food explained 30% of the observed biomarker variance. This result suggests that exposure 259 estimates based on dietary ingestion would be appropriate (but not necessarily accurate) and 260 comparable to acceptable levels for risk evaluation. However, given added information (such as 261 human activities data), it would be possible to explain more biomarker variance, thus increasing 262 the accuracy of the exposure estimates. The graph in Figure 4 shows a regression of biomarker 263 levels on covariate-adjusted environmental levels. Here, the adjusted environmental levels 264 reflect for each individual the combined effects of food concentration, food consumption (total mass), and other covariates. A regression R^2 value of 0.6 in this example suggests that the 265 266 combined effects of food concentration and covariates could explain 30% more biomarker 267 variance than food concentration alone. Thus, the significant covariates identified through 268 statistical analyses could be considered, along with the food measurements, to improve exposure 269 estimates. For examples of Tier 3 analyses, see Morgan et al. (2007; 2005; 2008), Wilson et al. 270 (2007), and Tulve et al. (2010).

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272 **2.2.4** Special considerations for biomarker variance components

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The frequency and duration of exposure, timing of sampling, and rates of kinetic processes (*i.e.*, ADME) can impact biomarker interpretation with respect to exposure and dose. Regression slopes and R² values in Tier 2 and Tier 3 analyses can reflect the magnitude of these impacts. However, it is difficult to quantify these effects without measuring biomarker concentrations over time. Longitudinal studies (with repeated measurements) can partition biomarker variance into that which is observed between subjects and that which is observed for a given subject over time. The respective magnitudes of these variance components can inform the importance of timed events and kinetic processes in biomonitoring studies, as demonstrated below.

282 Figures 5A and B show repeated biomarker measurements of individuals from two 283 theoretical groups. Both figures show 10 consecutive measurements of 50 subjects, with the first 284 measurements made at 6:00 am and the final measurements at midnight (12:00 am) on the same 285 In Figure 5A, biomarker measurements vary slightly over time (small within-person day. 286 variance) and are distinguishable between individuals (large between-person variance). These 287 observations suggest that individuals have different exposure/dose levels and that kinetic 288 processes occur slowly. Figure 5B shows large within-person variance in biomarker levels and 289 considerable overlap across individuals (small between-person variance), suggesting similar 290 exposure/dose levels between individuals and rapid chemical uptake and elimination.

Three example regressions of dose estimates on biomarker levels are given in both Figures 5A and B; here dose is approximated for each individual as their average biomarker level across all 10 measurements. In Example 1, dose is regressed on randomly selected spot biomarker levels; this simulates studies where one random biomarker measurement is made for each subject. Example 2 shows a regression of dose on end-of-day biomarker levels; this simulates studies where one biomarker measurement is made for each subject at a specific time point. In Example 3, dose is regressed on the average of three randomly selected measurements; this simulates studies where repeated measurements are made for each subject, and the measurements (or the biological samples themselves) are pooled (averaged) prior to analysis.

All three examples in Figure 5A show very similar slopes (ranging from 0.92 to 1.0) and R^2 values (ranging from 0.93 to 0.99). These results indicate that sampling strategy has little impact on biomarker interpretation when the between-person variance is large compared to the within-person variance. Specifically, these results suggest that the biomarker measurements from each of the three sampling examples could be used to accurately and precisely estimate dose levels (given the approximation of "dose" using average biomarker levels).

306 In contrast, the examples in Figure 5B indicate increased impacts of sampling events on 307 biomarker interpretation when the within-person variance is large compared to the betweenperson variance (slope range: 0.23 - 0.82; R² range: 0.22 - 0.76). Example 3, using the average 308 309 of three random measurements, shows the strongest association suggesting that multiple 310 measurements are needed when longitudinal data are highly varied. Results also show that spot 311 biomarker measurements, collected randomly (Example 1) or at a fixed time (Example 2), can 312 severely underestimate dose levels (slopes $\ll 1$) in these instances. (The importance of slope 313 attenuation with measurement error, defined as "attenuation bias", is further described by Lin et 314 al. (2005), Rappaport and Kupper (2008), and Sobus et al. (2010b)). Taken together, the 315 examples in Figures 5A and B demonstrate the need to (1) understand biomarker variance 316 components via repeated observations, and (2) sample more frequently when the within-person 317 variance is large compared to the between-person variance; this allows better estimation of the 318 "true" average biomarker level as an indicator of exposure/dose.

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320 **2.2.5** Tier 4: Biomonitoring for exposure and risk assessments

322	Figure 6 shows that Tier 4 analyses of biomarker data include the components for tier 3 analyses,
323	as well as kinetic models to link (1) exposure and dose, (2) dose and biomarker levels, and (3)
324	dose and BR dose. Linking the external environment to internal dose and biomarker levels is a
325	primary goal of exposure science (Sheldon and Cohen Hubal, 2009). Therefore, Tier 4 analyses
326	represent a general endpoint for exposure science and a starting point for health effects science.
327	Furthermore, since the BR dose estimate is the final output from a Tier 4 analysis, it can be
328	considered as the final output from exposure research, as well as a useful input for health effects
329	research (Pleil and Sheldon, 2010).
330	Using environmental measurements, exposure factors, and exposure and kinetic models,
331	Tier 4 analyses can answer the following questions for exposure and risk assessments:
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333	• What is the importance of each exposure route?
334	• What are the best estimates of exposure and dose?
335	• What are the likely BR dose levels?
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336 337	In Tier 3, exposure estimates and biomarker measurements were not directly linked. Rather,
336 337 338	In Tier 3, exposure estimates and biomarker measurements were not directly linked. Rather, results from statistical models were used as support for exposure estimates. Risk-based decisions
336337338339	In Tier 3, exposure estimates and biomarker measurements were not directly linked. Rather, results from statistical models were used as support for exposure estimates. Risk-based decisions can be supported by statistical associations, but can be further refined with an understanding of
 336 337 338 339 340 	In Tier 3, exposure estimates and biomarker measurements were not directly linked. Rather, results from statistical models were used as support for exposure estimates. Risk-based decisions can be supported by statistical associations, but can be further refined with an understanding of mass transfer from exposure to dose to biomarker levels; kinetic models are used to describe
 336 337 338 339 340 341 	In Tier 3, exposure estimates and biomarker measurements were not directly linked. Rather, results from statistical models were used as support for exposure estimates. Risk-based decisions can be supported by statistical associations, but can be further refined with an understanding of mass transfer from exposure to dose to biomarker levels; kinetic models are used to describe these mass transfer processes. More specifically, they are used to predict biological levels of

343 Example 1 in Figure 6 shows a theoretical comparison of observed and predicted 344 biomarker levels over time. Here the predicted values are estimated blood levels of a chemical 345 with a short biological half-life following three dietary exposure events (*e.g.*, breakfast, lunch, 346 and dinner). Assuming a well-parameterized and calibrated model, good agreement between 347 predicted and observed values support that diet is the primary exposure source and help validate 348 exposure estimates. Overestimation of the observed values would suggest incorrect exposure 349 estimates, whereas underestimation could suggest additional exposure routes or endogenous 350 sources of the biomarker. In these situations, exposure estimates could be reconstructed to be 351 consistent with observed values. (For methods and examples of exposure reconstruction, see 352 Kim et al. (2007), Tan et al. (2006), and Clewell et al. (2008)).

353 Given the appropriate model structure and parameters, the same kinetic models used to 354 predict biomarker levels may be used to predict the BR dose. Example 2 in Figure 6 extends 355 Example 1 and shows predicted levels at a target over time. In this theoretical example, the 356 parent chemical is neurotoxic, the target is the brain, and the predicted values are concentrations 357 of the parent chemical in the brain. Here, the health risks of the predicted values could be 358 evaluated using results of health effects studies. Specifically, the area under the target-level 359 curve (AUC_{target}, which is the time-integrated BR dose), or the maximum concentration at the 360 target, could be interpreted given some knowledge of the BR dose-response relationship. For an 361 example of a Tier 4 analysis, see Hore et al. (2006).

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363 **2.2.6** Tier 5: Biomonitoring to advance exposure and risk assessments

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As shown in Figure 7, Tier 5 analyses of biomarker data include all components of the source-tooutcome continuum, and predict both biomarker and BR biomarker levels for comparison to measured values. These comparisons allow Tier 5 analyses to answer the following research questions:

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- What are the best estimates of BR dose?
- What are the likely impacts of exposure on health risks?
- What other factors may affect health risks?
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In Tier 4 analyses, BR dose is estimated using kinetic models, and interpreted using knowledge of the BR dose-response relationship. Since BR dose estimates are not confirmed with measured values, there is uncertainty in model predictions. Tier 5 analyses further utilize kinetic/dynamic models to predict BR biomarker levels based on the BR dose estimates. Comparison of the predicted and observed BR biomarker levels can then reduce uncertainties in the BR dose estimates.

Example 1 in Figure 7 shows predicted versus observed levels of a BR biomarker. This extends the examples in Figure 6 where the brain was a target tissue, and the stressor was a chemical neurotoxin from food. In this example, blood enzymes are the BR biomarkers and act as surrogates for brain enzymes (*e.g.*, cholinesterase). Combined kinetic and dynamic models were used to predict blood enzyme levels following three theoretical dietary exposure events. Predicted and observed levels were then compared to evaluate the BR dose estimate.

In Example 1, good agreement between predicted and observed values indicates an
 accurate estimation of BR dose. Thus, the BR dose estimate could be used to inform health risks

from exposure. However, poor agreement between predicted and observed BR biomarker levels would suggest an incomplete understanding of kinetic/dynamic processes *in vivo*. Overestimation of BR biomarker levels could suggest the omission of important recovery processes, whereas underestimation could suggest additional exogenous or endogenous sources. In these instances, clarification would be necessary before placing BR dose estimates into a risk context.

394 Statistical comparisons of biomarker and BR biomarker measurements are also used in 395 Tier 5 analyses to elucidate health risks from exposure. For example, *in vivo* dose-response 396 associations can be informed using regressions of BR biomarker levels (representing response) 397 on biomarker levels (representing dose). Modifiers of the *in vivo* dose-response relationships can 398 also be observed by studying covariates such as age, gender, personal and family health history, 399 and genetic information.

Example 2 in Figure 7 shows a regression of BR biomarker levels on covariate-adjusted biomarker levels. Continuing from the previous example, this plot suggests that blood enzyme activities decreased with increasing adjusted biomarker levels. In other words, biological function was suppressed given elevated dose levels. This observation (specifically, the slope of the regression line), as well as the model results for covariates (coefficients and p-values), could help inform both exposure and susceptibility effects on BR dose, and by extension, health risks. For an example of a Tier 5 analysis, see Garabrant et al. (2009).

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408 **3.0** Conclusions

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410 Biomonitoring data can be used in many different ways; applications can be as simple as 411 documenting a population-based change in exposure, or as complex as linking the source-to-412 outcome progression through empirical data and sophisticated models. In this article, we created 413 a biomonitoring framework to demonstrate how measurements, models, and model estimates are 414 used together to answer specific exposure- and risk-based questions. In doing so, we presented a 415 tiered approach that categorizes the uses of biomarker data in the presence or absence of 416 corresponding information. (A summary of the uses and requirements of the five biomonitoring 417 tiers is given in Table 1). This tiered approach does not imply a hierarchy of biomonitoring 418 research based on value; that is, it does not rank analyses from "least useful" to "most useful". 419 Rather, it poses a logical structure to what is often a complex web of information. This structure 420 will help researchers (1) conceptualize sampling and analysis approaches when designing 421 targeted studies, and (2) weigh the costs (personnel, instrumentation, and time) and benefits 422 (ability to answer specific questions) of proposed studies, with the goal of maximizing public 423 health benefits.

424 Simple theoretical examples were given throughout the text to articulate the tiered 425 While we recommend using these examples as a guide for biomonitoring approaches. 426 interpreting existing data sets and for designing new studies, we caution that these examples 427 were simplified for demonstration purposes. That is, they did not address some common 428 challenges of biomonitoring studies, such as the needs to resolve non-specific biomarkers, 429 overcome analytical detection limitations, and interpret urinary biomarkers that are affected by 430 host hydration level (Albertini et al., 2006; Angerer et al., 2006; Barr and Angerer, 2006; Clewell 431 et al., 2008; Needham et al., 2007; NRC, 2006; Sobus et al., 2010a). Also, the examples were 432 generally geared towards evaluating short-term biomarkers of non-persistent chemicals,

433 particularly those given for Tier 4 and Tier 5 analyses; different approaches can be considered 434 when evaluating intermediate- and long-term biomarkers (Clewell et al., 2008; Rappaport and 435 Kupper, 2008). Finally, the examples assumed that analytical methods, mechanistic models 436 (*e.g.*, kinetic and dynamic models), and acceptable exposure levels exist for a given chemical of 437 interest. Complications arising from any one of these issues can hinder a biomonitoring analysis 438 and limit the use of individual biomarkers in research studies.

439 Given the potential limitations of biomonitoring studies, this framework lays a foundation 440 for identifying the key data and modeling gaps, and prioritizing research needs. For example, 441 this framework can help prioritize the needs for (1) empirical evidence to inform kinetic 442 parameters, (2) well-vetted exposure and kinetic models, (3) improved methods to measure 443 environmental concentrations, biomarkers, and BR biomarkers, and (4) robust datasets with 444 which to estimate biomarker variance components and to perform statistical analyses. In ensuing 445 articles, we will discuss current and proposed efforts to address these needs while keeping within the structure of this framework. The ultimate goal of these efforts is to provide cohesive 446 447 guidance that informs future biomonitoring studies, and catalyzes biomarker use in exposure and 448 human health research.

449

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- 457

458 **Figure captions:**

459 460

0 Figure 1. A source-to-outcome continuum.

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Figure 2. Requirements and examples of Tier 1 analyses of biomarker data. Grey objects are
unavailable in a Tier 1 analysis. Example 1 demonstrates a cross-sectional analysis where
cumulative percentile distributions of biomarker levels are compared across two groups.
Example 2 demonstrates a longitudinal analysis where biomarker levels for one group are
examined over time.

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- 468 *Figure 3. Requirements and an example of a Tier 2 analysis of biomarker data.* Grey objects 469 are unavailable in a Tier 2 analysis. The example graph shows a regression of spot biomarker 470 measurements on corresponding environmental measurements with an R^2 of 0.3. 471
- 472 *Figure 4. Requirements and an example of a Tier 3 analysis of biomarker data.* Grey objects 473 are unavailable in a Tier 3 analysis. Exposure estimates can be compared to acceptable levels 474 determined from animal studies. The example graph shows a regression of spot biomarker 475 measurements on corresponding covariate-adjusted environmental measurements with an R^2 of 476 0.6.
- 477

478 Figure 5. Impacts of sampling events on biomarker interpretation when between-person 479 variance is large and within-person variance is small (A), and when within-person variance is 480 large and between-person variance is small (B). In both (A) and (B), Example 1 shows dose 481 regressed on randomly selected spot biomarker measurements, Example 2 shows dose regressed 482 on end-of-day biomarker levels, and Example 3 shows dose regressed on the average of three 483 randomly selected measurements. Here dose is approximated for each individual as their 484 average biomarker level across all 10 measurements.

- 485
- *Figure 6. Requirements and examples of Tier 4 analyses of biomarker data.* Grey objects are
 unavailable in a Tier 4 analysis. Example 1 compares predicted and observed levels of a
 chemical biomarker over time. Example 2 shows predicted target levels over time of the same
 chemical from example 1. Here, the area under the target-level curve (AUC_{target}) is the BR dose.
- 490
- 491 *Figure 7. Requirements and examples of Tier 5 analyses of biomarker data.* Example 1
 492 compares predicted and observed BR biomarker levels over time. Example 2 shows a regression
 493 of BR biomarker measurements on covariate-adjusted biomarker levels.
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499	References
500	
501 502 503	Albertini R, Bird M, Doerrer N, Needham L, Robison S, Sheldon L, et al. The use of biomonitoring data in exposure and human health risk assessments. Environ Health Perspect 2006; 114: 1755-62.
504 505	Angerer J, Bird MG, Burke TA, Doerrer NG, Needham L, Robison SH, et al. Strategic biomonitoring initiatives: moving the science forward. Toxicol Sci 2006; 93: 3-10.
506 507 508	Barr DB, Angerer J. Potential uses of biomonitoring data: a case study using the organophosphorus pesticides chlorpyrifos and malathion. Environ Health Perspect 2006; 114: 1762.0
508 509 510	Barr DB, Olsson AO, Wong LY, Udunka S, Baker SE, Whitehead RD, et al. Urinary concentrations of metabolites of pyrethroid insecticides in the general U.S. population:
511 512	National Health and Nutrition Examination Survey 1999-2002. Environ Health Perspect 2010; 118: 742-8.
513 514	Birnbaum LS. Applying research to public health questions: biologically relevant exposures. Environ Health Perspect 2010; 118: A152.
515 516	Calafat AM, Ye X, Wong LY, Bishop AM, Needham LL. Urinary concentrations of four parabens in the U.S. population: NHANES 2005-2006. Environ Health Perspect 2010;
517 518	Clewell HJ, Tan YM, Campbell JL, Andersen ME. Quantitative interpretation of human
519 520 521	Egeghy PP, Quackenboss JJ, Catlin S, Ryan PB. Determinants of temporal variability in NHEXAS Maryland environmental concentrations, exposures, and biomarkers. J Expo
521 522 523	Anal Environ Epidemiol 2005; 15: 388-97.
525 524 525	inhibition in chlorpyrifos workers: Characterization of biomarkers of exposure and response in relation to urinary TCPy. J Expo Sci Environ Epidemiol 2009; 19: 634-42.
526 527 528	Hore P, Zartarian V, Xue J, Ozkaynak H, Wang SW, Yang YC, et al. Children's residential exposure to chlorpyrifos: application of CPPAES field measurements of chlorpyrifos and TCPy within MENTOP/SHEDS. Posticides model. Sci Total Environ 2006; 366: 525-37
528 529 530	Kim D, Andersen ME, Chao YC, Egeghy PP, Rappaport SM, Nylander-French LA. PBTK modeling demonstrates contribution of dermal and inhalation exposure components to
530 531 532	end-exhaled breath concentrations of naphthalene. Environ Health Perspect 2007; 115: 894-901.
533 534	Lin YS, Kupper LL, Rappaport SM. Air samples versus biomarkers for epidemiology. Occup Environ Med 2005; 62: 750-60.
535 536	Metcalf SW, Orloff KG. Biomarkers of exposure in community settings. J Toxicol Environ Health A 2004; 67: 715-26.
537 538 539	Morgan MK, Sheldon LS, Croghan CW, Jones PA, Chuang JC, Wilson NK. An observational study of 127 preschool children at their homes and daycare centers in Ohio: environmental pathways to cis- and trans-permethrin exposure. Environ Res 2007; 104:
540 541 542	266-74. Morgan MK, Sheldon LS, Croghan CW, Jones PA, Robertson GL, Chuang JC, et al. Exposures of preschool children to chlorpyrifos and its degradation product 3,5,6-trichloro-2-

- 543pyridinol in their everyday environments. J Expo Anal Environ Epidemiol 2005; 15: 297-544309.
- Morgan MK, Sheldon LS, Thomas KW, Egeghy PP, Croghan CW, Jones PA, et al. Adult and
 children's exposure to 2,4-D from multiple sources and pathways. J Expo Sci Environ
 Epidemiol 2008; 18: 486-94.
- Naeher LP, Tulve NS, Egeghy PP, Barr DB, Adetona O, Fortmann RC, et al. Organophosphorus
 and pyrethroid insecticide urinary metabolite concentrations in young children living in a
 southeastern United States city. Sci Total Environ 2010; 408: 1145-53.
- Needham LL, Calafat AM, Barr DB. Uses and issues of biomonitoring. Int J Hyg Environ Health
 2007; 210: 229-38.
- 553 NRC. Biological markers in environmental health research. Committee on Biological Markers of
 554 the National Research Council. Environ Health Perspect 1987; 74: 3-9.
- 555 NRC. Monitoring human tissues for toxic substances. Washington, DC: The National Academies
 556 Press, 1991.
- 557 NRC. Human biomonitoring for environmental chemicals. Washington, DC: The National
 558 Academies Press, 2006.
- 559 Pirkle JL, Bernert JT, Caudill SP, Sosnoff CS, Pechacek TF. Trends in the exposure of
 560 nonsmokers in the U.S. population to secondhand smoke: 1988-2002. Environ Health
 561 Perspect 2006; 114: 853-8.
- 562 Pleil JD, Sheldon LS. Adapting concepts from systems biology to develop systems exposure
 563 event networks for exposure science research. Biomarkers 2010; 16: 99-105.
- Rappaport SM, Kupper LL. Quantitative exposure assessment. El Cerrito, CA: Stephen M.
 Rappaport, 2008.
- 566 Sheldon LS, Cohen Hubal EA. Exposure as part of a systems approach for assessing risk.
 567 Environ Health Perspect 2009; 117: 119-1194.
- Sobus JR, McClean MD, Herrick RF, Waidyanatha S, Nylander-French LA, Kupper LL, et al.
 Comparing urinary biomarkers of airborne and dermal exposure to polycyclic aromatic
 compounds in asphalt-exposed workers. Ann Occup Hyg 2009a; 53: 561-71.
- Sobus JR, McClean MD, Herrick RF, Waidyanatha S, Onyemauwa F, Kupper LL, et al.
 Investigation of PAH biomarkers in the urine of workers exposed to hot asphalt. Ann
 Occup Hyg 2009b; 53: 551-60.
- Sobus JR, Morgan MK, Pleil JD, Barr DB. Biomonitoring: uses and considerations for assessing
 nonoccupational human exposure to pesticides. In: Krieger R, editor. Hayes' Handbook
 of Pesticide Toxicology. 1. Elsevier Inc., 2010a, pp. 1021-1036.
- Sobus JR, Pleil JD, McClean MD, Herrick RF, Rappaport SM. Biomarker variance component
 estimation for exposure surrogate selection and toxicokinetic inference. Toxicol Lett
 2010b; 199: 247-53.
- Sobus JR, Waidyanatha S, McClean MD, Herrick RF, Smith TJ, Garshick E, et al. Urinary
 naphthalene and phenanthrene as biomarkers of occupational exposure to polycyclic
 aromatic hydrocarbons. Occup Environ Med 2009c; 66: 99-104.
- Tan YM, Liao KH, Conolly RB, Blount BC, Mason AM, Clewell HJ. Use of a physiologically
 based pharmacokinetic model to identify exposures consistent with human biomonitoring
 data for chloroform. J Toxicol Environ Health A 2006; 69: 1727-56.
- Thomas KW, Dosemeci M, Coble JB, Hoppin JA, Sheldon LS, Chapa G, et al. Assessment of a
 pesticide exposure intensity algorithm in the agricultural health study. J Expo Sci Environ
 Epidemiol 2010a; 20: 559-69.

- Thomas KW, Dosemeci M, Hoppin JA, Sheldon LS, Croghan CW, Gordon SM, et al. Urinary
 biomarker, dermal, and air measurement results for 2,4-D and chlorpyrifos farm
 applicators in the Agricultural Health Study. J Expo Sci Environ Epidemiol 2010b; 20:
 119-34.
- Tulve NS, Egeghy PP, Fortmann RC, Xue J, Evans J, Whitaker DA, et al. Methodologies for
 estimating cumulative human exposures to current-use pyrethroid pesticides. J Expo Sci
 Environ Epidemiol 2010.
- USEPA. A conceptual framework for U.S. EPA's National Exposure Research Laboratory.
 2009. Available: <u>http://www.epa.gov/nerl/features/exposure_framework.html</u> [accessed
 28 April 2011].
- 599 Wilson NK, Chuang JC, Morgan MK, Lordo RA, Sheldon LS. An observational study of the 600 potential exposures of preschool children to pentachlorophenol, bisphenol-A, and 601 normalisher and damage Environ Res 2007; 102: 0, 20
- nonylphenol at home and daycare. Environ Res 2007; 103: 9-20.

Tier	Primary uses	Measurements needed	Models needed	Estimated values
1	Exposure surveillance: Who is exposed? What are the exposure trends? Which chemicals should be prioritized for higher-tier analyses?	1) Biomarker	none	none
2	Supporting exposure assessment: What are the likely exposure sources? What are the likely exposure routes?	 Environmental Biomarker 	1) Statistical	none
3	Supporting risk assessment: What are the important exposure factors? What are the likely exposure levels?	 Environmental Biomarker 	 1) Statistical 2) Exposure 	1) Exposure
4	Exposure and risk assessment: What is the importance of each exposure route? What are the best estimates of exposure and dose? What are the likely BR dose levels?	1) Environmental 2) Biomarker	 Statistical Exposure Kinetic 	 1) Exposure 2) Dose 3) BR dose
5	Advancing exposure and risk assessment: What are the best estimates of BR dose? What are the likely impacts of exposure on health risks? What other factors may affect health risks?	 Environmental Biomarker BR biomarker 	 Statistical Exposure Kinetic Dynamic 	 1) Exposure 2) Dose 3) BR dose

Table 1. Uses and requirements of the five biomonitoring tiers.



Components of the source-to-outcome continuum

Symbol	Key	Parameter	Definition
\bigtriangleup	Estimated Value	 1) Exposure estimate 2) Dose estimate 3) BR dose estimate 	 Estimated mass of a chemical that comes into contact with a human over time Estimated mass of a chemical inside a human over time Estimate amount of the dose at a specific target inside a human
	Measured value	 Environmental measurement Biomarker measurement BR biomarker measurement 	 Observation of a stressor in environmental media that reflects a source Observation of a stressor in biological media that reflects an exposure/dose Observation of a stressor in biological media that reflects a BR dose
	Empirical model Mechanistic model	 Statistical model (blue) Exposure model (red) Kinetic model (red) Dynamic model (red) 	 Model that evaluates observed variables for hypothesis testing Model that estimates exposure using environmental measurements and exposure factors Model that describes how a stressor enters and is removed from a human Model that describes the effect of a stressor on the human body

Figure2 Click here to download Figure: Figure2.pptx





Figure4 Click here to download Figure: Faceptable levels



Figure5a Click here to download Figure: Figure5a.pptx



Figure5b Click here to download Figure: Figure5b.pptx





Figure7 Click here to download Figure: Figure7.pptx



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