

Available online at www.sciencedirect.com





Toxicology in Vitro 21 (2007) 1373-1381

# Nanoparticle effects on rat alveolar epithelial cell monolayer barrier properties

Nazanin R. Yacobi <sup>a,c,\*</sup>, Harish C. Phuleria <sup>d</sup>, Lucas Demaio <sup>a,b</sup>, Chi H. Liang <sup>c</sup>, Ching-An Peng <sup>c</sup>, Constantinos Sioutas <sup>d</sup>, Zea Borok <sup>a,b</sup>, Kwang-Jin Kim <sup>a,b</sup>, Edward D. Crandall <sup>a,b,c</sup>

<sup>a</sup> Will Rogers Institute Pulmonary Research Center, University of Southern California, Keck School of Medicine, 2011 Zonal Avenue, HMR 914, Los Angeles, CA 90033, USA

<sup>b</sup> Department of Medicine, University of Southern California, Los Angeles, CA 90033, USA <sup>c</sup> Mork Family Department of Chemical Engineering and Materials Science, University of Southern California, Los Angeles, CA 90033, USA <sup>d</sup> Department of Civil and Environmental Engineering, University of Southern California, Los Angeles, CA 90033, USA

> Received 22 February 2007; accepted 13 April 2007 Available online 27 April 2007

#### Abstract

Inhaled nanoparticles have been reported to contribute to deleterious effects on human health. In this study, we investigated the effects of ultrafine ambient particulate suspensions (UAPS), polystyrene nanoparticles (PNP; positively and negatively charged; 20, 100, 120 nm), quantum dots (QD; positively and negatively charged; 30 nm) and single-wall carbon nanotubes (SWCNT) on alveolar epithelial cell barrier properties. Transmonolayer resistance ( $R_t$ ) and equivalent short-circuit current ( $I_{eq}$ ) of primary rat alveolar epithelial monolayers were measured in the presence and absence of varying concentrations of apical nanoparticles. In some experiments, apical-to-basolateral fluxes of radiolabeled mannitol or inulin were determined with or without apical UAPS exposure and lactate dehydrogenase (LDH) release was analyzed after UAPS or SWCNT exposure. Results revealed that exposure to UAPS decreased  $R_t$  and  $I_{eq}$ significantly over 24 h, although neither mannitol nor inulin fluxes changed. Positively charged QD decreased  $R_t$  significantly (with subsequent recovery), while negatively charged QD did not.  $R_t$  decreased significantly after SWCNT exposure (with subsequent recovery). On the other hand, PNP exposure had no effects on  $R_t$  or  $I_{eq}$ . No significant increases in LDH release were observed after UAPS or SWCNT exposure. These data indicate that disruption of alveolar epithelial barrier properties due to apical nanoparticle exposure likely involves alteration of cellular transport pathways and is dependent on specific nanoparticle composition, shape and/or surface charge. © 2007 Elsevier Ltd. All rights reserved.

Keywords: Lung injury; Ultrafine particles; Pulmonary toxicity; Epithelial transport; Primary culture

## 1. Introduction

Nanoparticles are commonly defined as particles having at least one dimension of <100 nm. Investigators in environmental health usually refer to particles smaller than 100 nm diameter in ambient air as ultrafine particles. These ultrafine ambient particulates and engineered nanoparticles possess nanostructure-dependent properties due to their small size, chemical composition, surface charge, solubility and/or shape (Oberdorster et al., 2005b; Xia et al., 2006).

Particulates in ambient air and engineered nanoparticles have increasingly been found to be associated with adverse cardiovascular and pulmonary effects, with suggestions of increased morbidity and mortality in susceptible populations (Oberdorster et al., 1995; Oberdorster et al., 2005b; Peters et al., 2001; Sun et al., 2005; Wichmann et al.,

<sup>&</sup>lt;sup>\*</sup> Corresponding author. Address: Will Rogers Institute Pulmonary Research Center, University of Southern California, Keck School of Medicine, 2011 Zonal Avenue, HMR 914, Los Angeles, CA 90033, USA. Tel.: +1 323 442 1217; fax: +1 323 442 2611.

E-mail address: nyaghoob@usc.edu (N.R. Yacobi).

 $<sup>0887\</sup>text{-}2333/\$$  - see front matter @ 2007 Elsevier Ltd. All rights reserved. doi:10.1016/j.tiv.2007.04.003

2000). Since inhaled ambient ultrafine particles can be found in heart, bone marrow, blood vessels and other organs (Nemmar et al., 2002; Nemmar et al., 2001; Oberdorster, 2001), their most likely route of entry into the circulation is across the epithelia of the lung, especially the alveolar epithelium with its very large surface area and thin barrier thickness. Further knowledge about the mechanisms by which particles injure, interact with and/or are transported across the alveolar epithelium is thus of considerable importance for understanding health effects related to inhalation of ultrafine particles in ambient air.

Determination of the characteristics of ambient particulates and engineered nanoparticles that might cause injury. and the mechanisms by which they do so, requires further study (Calcabrini et al., 2004; Ghio and Devlin, 2001; Oberdorster et al., 2005b; Xia et al., 2006). Size, shape, charge and/or composition may be important factors that influence how particles affect human health (Alfaro-Moreno et al., 2002; Calcabrini et al., 2004; Gutierrez-Castillo et al., 2006; Oberdorster et al., 2005a; Topinka et al., 2000; Vedal, 1997; Xia et al., 2006). Particles smaller than 250 nm are known to reach the distal lung and likely interact with alveolar epithelium. Because of their increased number and surface area as well as their high pulmonary deposition efficiency, ambient ultrafine particles are likely to be important in environmental health (Cassee et al., 2002; Donaldson et al., 2001; Oberdorster et al., 2005b), although some reports have suggested that coarse particles  $(250 \text{ nm} < \text{aerodynamic diameter} < 10 \,\mu\text{m})$  may be more toxic than fine (100 nm  $\leq$  aerodynamic diameter  $\leq$  250 nm) and ultrafine particles (Monn and Becker, 1999; Osornio-Vargas et al., 2003). Different reports about the consequences of exposure to engineered nanoparticles are inconsistent, with some studies indicating little effect (Gevs et al., 2006; Muldoon et al., 2005; Zhang et al., 2006) and others suggesting significant toxicity using both in vivo and in vitro models (Gurr et al., 2005; Magrez et al., 2006; Sayes et al., 2006; Shvedova et al., 2005).

Studies using in vitro models have permitted more detailed understanding of important biological properties of the lung in vivo, such as the presence of functional epithelial tight junctions and pathways responsible for active and passive ion transport in alveolar epithelium. Greater than 95% of lung surface area is lined by alveolar epithelial type I (AT1) cells. Alveolar epithelial type II (AT2) cells in primary culture have been demonstrated to undergo morphologic (Cheek et al., 1989a) and phenotypic (Danto et al., 1992) transdifferentiation into AT1-like cells (Adamson and Bowden, 1975; Kim et al., 2001a). AT1 cell-like monolayers represent a reliable model for the study of alveolar epithelial transport biology/physiology, since many of the transport processes and other characteristics demonstrated in these primary cultures appear representative of those in the respiratory epithelium lining the distal region of the intact lung (Elbert et al., 1999; Kim et al., 2001a). In this study, we utilized primary rat alveolar epithelial cell monolayers (RAECM) exhibiting AT1 cell-like phenotype

(Cheek et al., 1989a; Danto et al., 1992) to investigate potential toxicity of ultrafine ambient particle suspensions (UAPS) and several different engineered nanoparticles.

#### 2. Materials and methods

## 2.1. Engineered nanoparticles

Polystyrene nanoparticles (PNP) were purchased from Molecular Probes (Eugene, OR). Carboxylate-modified PNP of 20 and 100 nm diameter (-304.3 and  $-320 \mu Eq$  surface charge/g, respectively) are negatively charged. Amidine-modified PNP of 20 and 120 nm diameter (80.2 and  $39.7 \mu Eq$  surface charge/g, respectively) are positively charged.

Hipco<sup>®</sup> single-wall carbon nanotubes (SWCNT) were purchased from Carbon Nanotechnologies (Houston, TX). SWCNT were produced by a high pressure CO conversion synthesis method (Bronikowski et al., 2001). Individual SWCNT diameter is between 0.8 and 1.2 nm and length is between 100 and 1000 nm.

Chitosan coated (positively charged) and alginate coated (negatively charged) quantum dots (QD, 30 nm) were manufactured in our laboratories. To synthesize CdSe/ZnS QD, 25.68 mg CdO (Sigma, St. Louis, MO) as precursor was used (Huang et al., 2004). CdO was dissolved in a coordinating solvent mixture of 3.88 mg tri-noctylphosphine oxide (TOPO, Sigma) and 2.41 mg hexadecylamine (HDA, Sigma). The entire process was carried out in a dry nitrogen atmosphere. Selenium powder (31.58 mg) dissolved in 5 mL tributylphosphine (TBP, Sigma) was injected rapidly into the TOPO-CdO-HDA solution with vigorous stirring. CdSe nanocrystals were grown for 5 min at 300 °C after mixing. To form a ZnS shell on the CdSe core, temperature was reduced to 160 °C and a ZnS shell solution (379.4 mg zinc stearate and 12.8 mg sulfur powder dissolved in 5 mL TBP) was added into the core solution under thorough stirring over a period of 15 min. After the addition of shell solution was completed, the resulting core/shell solution was cooled to 120 °C and left stirring to anneal for 2 h. The solution was further cooled to 70 °C and anhydrous methanol was added to precipitate the nanocrystals, which were collected by centrifugation and dispersed in anhydrous toluene or chloroform. To modify surface charge of QD, chitosan (Sigma) or alginate (Sigma) after grafting with hydrophobic alkyl moieties were used to encapsulate QD with average diameter of 5 nm. The size of QD coated with amphiphilic alginate or chitosan is  $\sim 30$  nm by dynamic light scattering measurements (Wyatt Technology, Santa Barbara, CA).

## 2.2. Ultrafine ambient particulate suspensions (UAPS)

Ultrafine particle samples were collected in Los Angeles, CA in the summer. The collection site was located downwind of two major freeways (15–20 m and 150 m from the Santa Ana and Pomona Freeways, respectively). Particle samples were collected over a period of 7-10 days for 5-6 h/day. At the end of each day's collection, samples were frozen in Teflon-lid glass jars. After one set of complete collections, multiple samples were combined and re-frozen prior to subsequent utilization in the experiments described below.

Ambient ultrafine particles were collected using the Versatile Aerosol Concentration Enrichment System (VACES) (Kim et al., 2001b,c). Theory and operation of VACES is described in detail elsewhere (Kim et al., 2001b; Li et al., 2003). Briefly, ambient ultrafine particles were concentrated using 0.15 µm cut-point preimpaction to remove larger particles. Air samples are drawn through a saturationcondensation system that grows particles to 2-3 µm droplets, which are subsequently concentrated by virtual impaction. Highly concentrated particle suspensions were obtained by connecting the VACES output to a sterilized liquid impinger (BioSampler, SKC West, Fullerton, CA). Aerosols were collected using ultrapure deionized water as the collection medium. Total amount of particulate loading in the collection medium was determined by multiplying the ambient concentration of each particulate population by the total air sample volume collected by each VACES line. The particle concentration in the aqueous medium was then calculated by dividing the particle loading by the total volume collected in that time period. The concentration enrichment process does not alter the physical, chemical or morphological properties of the particles (Kim et al., 2001b,c; Li et al., 2003). Table 1 shows the chemical composition of UAPS utilized in this study, which contain a high percentage of organic components (mostly hydrophobic) and inorganic hydrophilic compounds such as sulfates and nitrates and a portion of trace elements and metals.

## 2.3. Primary culture of RAECM

The detailed procedure for routine generation of primary RAECM has appeared elsewhere (Borok et al., 1994; Borok et al., 1995). Briefly, fresh AT2 cells were isolated from adult, male, specific pathogen-free Sprague-Dawley rats (125–150 g) using elastase digestion and purification by IgG panning. Purified AT2 cells were plated onto tissue culture-treated polycarbonate filters (Transwell, 0.4 µm pore, 12 mm diameter, Corning-Costar, Cam-

Table I			
Chamiaal	composition	of IIA	DC

**T** 1 1 1

Chemical species	% By mass	
Nitrate	8.76	
Sulfate	8.84	
Elemental carbon	5.73	
Organic carbon	47.80	
Trace elements and metals <sup>a</sup>	15.88	
Unknown	13.01	

<sup>a</sup> Predominant components of trace elements and metals were silicon, aluminum, iron, calcium and zinc.

bridge, MA) at  $1.2 \times 10^6$  cells/cm<sup>2</sup>. Cells were maintained at 37 °C in a humidified atmosphere of 5% CO<sub>2</sub>/95% air. Culture medium (MDS) consisted of 10% newborn bovine serum in minimally defined serum-free medium (MDSF). MDSF is a 1:1 mixture of DME/F-12 (Sigma, St. Louis, MO) supplemented with 1% nonessential amino acids (Sigma), 0.2% primocin (InvivoGen, San Diego, CA), 10 mM *N*-(2-hydroxyethyl)piperazine-*N'*-(2-ethanesulfonic acid) hemisodium salt (Sigma), 1.25 mg/mL bovine serum albumin (BD Bioscience, San Jose, CA) and 2 mM L-glutamine (Sigma). Cells were fed every other day, starting on day 3 in culture, when they form confluent monolayers.

#### 2.4. Measurement of RAECM bioelectric properties

Transmonolayer resistance ( $R_t$ , K $\Omega$  cm<sup>2</sup>) and potential difference (PD, mV, apical side as reference) in the presence or absence of varying concentrations of apical nanoparticles were measured using a rapid screening device (Millicell-ERS, Millipore, Bedford, MA) equipped with a pair of silver/silver chloride (Ag/AgCl) electrodes (Cheek et al., 1989b). Short-circuit current ( $I_{eq}$ ,  $\mu$ A/cm<sup>2</sup>) was calculated according to Ohm's law as described previously. Immediately before and at different time points after apical exposure to nanoparticles,  $R_t$  and PD were measured and  $I_{eq}$  calculated.

## 2.5. Apical exposure of RAECM to nanoparticles

We studied the effects of apical exposure to nanoparticles by replacing monolayer apical fluid on days 4, 5 or 6 in culture with isotonic solutions of UAPS, PNP, OD or SWCNT. UAPS in water were adjusted to isosmolarity using NaCl. PNP, QD and SWCNT were suspended in MDS. These working stocks were sonicated briefly and appropriate volume was used to replace apical fluid. Apical fluid of control RAECM was replaced at the same time as experimental RAECM using MDS or isosmolar NaCl solution without nanoparticles. Effects of UAPS, PNP, QD and SWCNT concentrations of up to 36, 706, 176 and 88 µg/ mL, respectively, were studied. In some experiments, monolayers were apically exposed for 120 min to UAPS (9 µg/ mL), followed by replacement of apical fluid with fresh culture medium.  $R_t$  and  $I_{eq}$  were usually assessed at 15, 30, 60, 120, 240 and 1440 min. Furthermore, permeabilities of <sup>14</sup>Cmannitol (180 Da) and <sup>14</sup>C-inulin (~5000 Da) were estimated from their steady state fluxes across RAECM in the apical-to-basolateral direction in the presence or absence of UAPS (36 µg/mL). Five microliter of radiotracer amounts of either mannitol or inulin were added immediately after apical UAPS exposure. Fifty microliter samples of basolateral fluid were taken at 30, 60, 120, 240 and 1440 min after radiotracer instillation into apical fluid. Ten microliters samples were taken from apical fluid at 60 and 1440 min for determination of upstream radioactivity. Radioactive samples were mixed with 15 mL Ecoscint (National Diagnostics, Atlanta, GA) and assayed

using a beta counter (Beckman Instruments, Fullerton, CA).

## 2.6. Measurement of lactate dehydrogenase (LDH) release

Extracellular LDH (in apical and basolateral fluid) at 1 and 2 h after exposure to UAPS ( $18 \mu g/mL$ ) or SWCNT ( $88 \mu g/mL$ ) was measured using a colorimetric cytotoxicity detection kit (Roche, Indianapolis, IN) following the manufacturer's instructions. As control, release of LDH was obtained for unexposed cells (low control) and maximum release of LDH was obtained by lysis of cells with 0.2% TX-100 (high control). Cytotoxicity is defined by:

(release of LDH of exposed cells – release of LDH (low control))/(release of LDH (high control) – release of LDH (low control))

#### 2.7. Statistical analyses

Data are presented as mean  $\pm$  standard error. For comparisons of multiple group means, one-way or two-way analyses of variance (ANOVA) followed by post-hoc procedures based on modified Newman-Keuls-Student tests were performed using GB-STAT v9.0 software (Dynamic Microsystems, Silver Spring, MD). p < 0.05 was considered to be statistically significant.

## 3. Results

Time courses of changes in  $R_t$  following apical exposure of RAECM to UAPS are shown in Fig. 1. Average  $R_t$  of RAECM before UAPS exposure was  $1.41 \pm 0.05 \text{ K}\Omega \text{ cm}^2$ (n = 45). At the highest UAPS concentration (36 µg/mL)studied,  $R_t$  declined ~60% with half time of ~30 min. From 2 to 24 h, no further significant changes in  $R_t$  were observed, despite continued presence of UAPS in apical fluid. No significant changes in cytotoxicity compared to controls were observed after 1 and 2 h of monolayer exposure to 18 µg/mL UAPS (n = 5, data not shown).

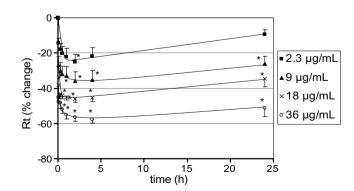


Fig. 1. Effects of apical exposure to UAPS on  $R_t$  of RAECM (n = 4-9 for each concentration).  $R_t$  of all monolayers prior to apical UAPS exposure (at t = 0) was  $1.41 \pm 0.05 \text{ K}\Omega \text{ cm}^2$  (n = 45). At the maximum concentration of UAPS (36 µg/mL) studied,  $R_t$  declined significantly by ~60% after 2 h of exposure and did not change further over 24 h. \* = significantly different (p < 0.05) from control (monolayers not exposed to UAPS).

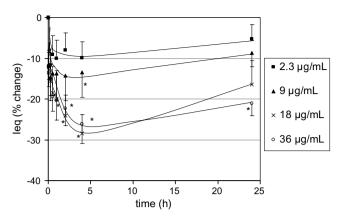


Fig. 2. Effects of apical exposure to UAPS on  $I_{eq}$  of RAECM (n = 4-9 for each concentration).  $I_{eq}$  of all monolayers prior to apical UAPS exposure of (at t = 0) was  $4.72 \pm 0.10 \,\mu\text{A/cm}^2$  (n = 45). At the maximum concentration of UAPS (36  $\mu$ g/mL) studied,  $I_{eq}$  declined significantly by ~25% after 30 min of exposure and did not change further for up to 24 h. \* = significantly different (p < 0.05) from control (monolayers not exposed to UAPS).

Fig. 2 shows the effects of apical UAPS on  $I_{eq}$ .  $I_{eq}$  prior to UAPS exposure was  $4.72 \pm 0.10 \,\mu\text{A/cm}^2$  (n = 45). At the highest apical UAPS concentration ( $36 \,\mu\text{g/mL}$ ) studied,  $I_{eq}$  decreased by ~25% at 2 h. From 2 to 24 h, no further significant changes in  $I_{eq}$  were observed, despite the continued presence of UAPS in apical fluid.

Fig. 3 shows time courses of changes in  $R_t$  of RAECM exposed apically to 9 µg/mL UAPS for 2 h, followed by replacement of apical fluid with fresh culture medium.  $R_t$  prior to UAPS exposure was  $2.33 \pm 0.68 \text{ K}\Omega \text{ cm}^2$  (n = 3).  $R_t$  recovered toward control after replacing apical fluid with fresh culture medium. Washout caused an increase in  $I_{eq}$ , followed by a gradual return toward its initial value (data not shown).  $I_{eq}$  prior to UAPS exposure was  $5.29 \pm 1.55 \mu \text{A/cm}^2$  (n = 3).

Apparent permeabilities  $(P_{app})$  of <sup>14</sup>C-mannitol and <sup>14</sup>Cinulin measured in the presence and absence of apical exposure to UAPS (36 µg/mL) are summarized in Table 2.  $P_{app}$ 

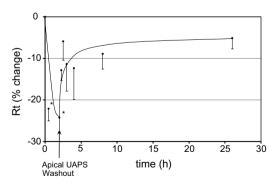


Fig. 3. Effects of removal of apical UAPS (9 µg/mL) on  $R_t$  of RAECM after 2 h of exposure.  $R_t$  prior to apical exposure to UAPS (t = 0) was  $2.33 \pm 0.68 \text{ K}\Omega \text{ cm}^2$  (n = 3). After 1 and 2 h of UAPS exposure,  $R_t$  decreased significantly.  $R_t$  after replacement of UAPS with fresh culture medium recovered toward control values. \* = significantly different (p < 0.05) from control (monolayers not exposed to UAPS).

Table 2 Apparent permeability  $(P_{app})$  of mannitol and inulin

	Control	UAPS
<sup>14</sup> C-Mannitol	$7.4 \pm 4.2$	$8.8\pm8.1$
<sup>14</sup> C-Inulin	$3.0\pm2.0$	$7.0\pm3.4$

Effects of apical exposure to UAPS (36 µg/mL) on apparent permeability coefficient ( $P_{app} \times 10^7$  cm/s) of RAECM to <sup>14</sup>C-mannitol and <sup>14</sup>C-inulin.  $P_{app}$  are shown as mean  $\pm$  SEM (n = 3). No significant differences due to UAPS exposure were seen.

of neither <sup>14</sup>C-mannitol nor <sup>14</sup>C-inulin measured in the apical-to-basolateral direction was significantly altered by UAPS exposure.

Time courses of changes in  $R_t$  following apical exposure of RAECM to positively or negatively charged QD (176 µg/mL) are shown in Fig. 4. Average  $R_t$  before QD exposure was  $2.68 \pm 0.14 \text{ K}\Omega \text{ cm}^2$  (n = 10). Positively charged QD decreased  $R_t$  by ~60% after 2 h of exposure, while negatively charged QD did not decrease  $R_t$  significantly. Effects of these two types of QD on  $I_{eq}$  were not significantly different from control over 24 h (data not shown). Lower concentrations (44–88 µg/mL) of QD (positively or negatively charged) did not decrease  $R_t$  significantly over 24 h of exposure (data not shown).

Time courses of changes in  $R_t$  following apical exposure of RAECM to SWCNT (up to 88 µg/mL) are shown in Fig. 5. Average  $R_t$  before SWCNT exposure was  $3.24 \pm 0.07 \text{ K}\Omega \text{ cm}^2$  (n = 33).  $R_t$  decreased significantly compared to control monolayers by ~40% after 1 h of exposure, with recovery to initial value by 4–24 h. The decrease in  $R_t$  after exposure to SWCNT was not dose dependent. Effects of SWCNT (up to 88 µg/mL) on  $I_{eq}$ were not significantly different from control for up to 24 h of exposure (data not shown). No significant changes in cytotoxicity were observed after 1 and 2 h of monolayer exposure to 88 µg/mL SWCNT (n = 5, data not shown).

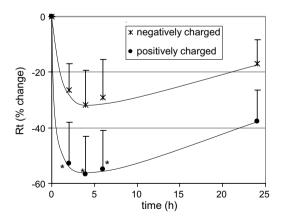


Fig. 4. Effects of apical exposure to QD (176 µg/mL) on  $R_t$  of RAECM (n = 6 for each concentration).  $R_t$  of all monolayers prior to apical exposure to QD (at t = 0) was  $2.68 \pm 0.14 \text{ K}\Omega \text{ cm}^2$  (n = 19). Exposure to positively charged QD decreased  $R_t$  significantly after 1 h, with recovery to control level by 24 h. \* = significantly different (p < 0.05) from control (monolayers not exposed to QD).

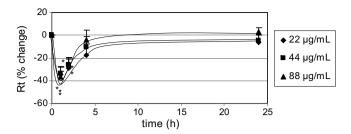


Fig. 5. Effects of apical exposure to SWCNT on  $R_t$  of RAECM (n = 9 for each concentration).  $R_t$  of all monolayers used prior to apical exposure to SWCNT (at t = 0) was  $3.24 \pm 0.07 \text{ K}\Omega \text{ cm}^2$  (n = 33). Exposure to SWCNT up to 88 µg/mL decreased  $R_t$  significantly after 1 h.  $R_t$  recovered to control level by 4–24 h. \* = significantly different (p < 0.05) from control (monolayers not exposed to SWCNT).

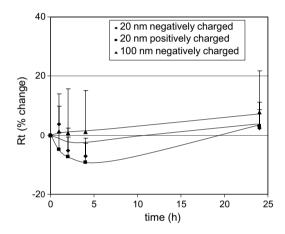


Fig. 6. Effects of apical exposure to PNP (176 µg/mL, 20 nm positively and negatively charged and 100 nm negatively charged) on  $R_t$  of RAECM (n = 10-13 for each concentration).  $R_t$  of all monolayers prior to apical exposure to PNP (at t = 0) was  $3.58 \pm 0.11 \text{ K}\Omega \text{ cm}^2$  (n = 47). These three types of PNP, as well as positively charged 120 nm PNP (data not shown), did not cause significant changes in  $R_t$  or  $I_{eq}$  (data not shown).

Fig. 6 shows the time courses of  $R_t$  observed for PNP (176 µg/mL) exposure. RAECM prior to PNP exposure show average  $R_t = 3.58 \pm 0.11 \text{ K}\Omega \text{ cm}^2$  and average  $I_{eq} = 4.48 \pm 0.16 \text{ µA/cm}^2$  (n = 47). No significant changes were observed for 20 nm positively and negatively charged PNP or for 100 nm negatively and 120 nm positively (data not shown) charged PNP. Up to 706 µg/mL of PNP in apical fluid led to no significant changes in  $R_t$  or  $I_{eq}$  for up to 24 h (data not shown), and 176 µg/mL PNP in apical fluid led to no significant changes in  $R_t$  or  $I_{eq}$  for up to 6 days (data not shown).

Fig. 7 summarizes changes in  $R_t$  after 2, 4 and 24 h of exposure to 36 µg/mL of UAPS, 176 µg/mL QD (positively and negatively charged), 88 µg/mL SWCNT and PNP (20 and 100 nm negatively charged and 20 nm positively charged). Average  $R_t$  of RAECM before nanoparticle exposure was  $2.13 \pm 0.27 \text{ K}\Omega \text{ cm}^2$  (n = 66). Exposure of RAECM to different nanoparticles affects  $R_t$  differently due to different nanoparticle composition, charge and/or concentration used.

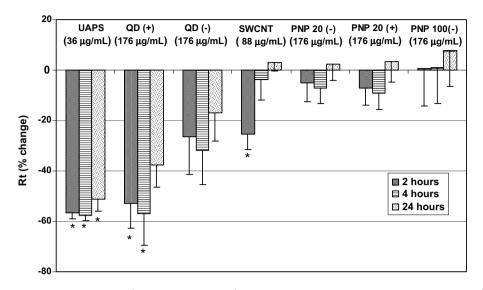


Fig. 7. Effects of apical exposure to UAPS (36 µg/mL), SWCNT (88 µg/mL), QD (positively and negatively charged, 176 µg/mL), and PNP (20 nm (negatively and positively charged) and 100 nm (negatively charged), 176 µg/mL) on  $R_t$  of RAECM after 2, 4 and 24 h of exposure (n = 5-9 for each condition).  $R_t$  of all monolayers prior to apical exposure to nanoparticles (at t = 0) was  $2.13 \pm 0.27$  K $\Omega$  cm<sup>2</sup> (n = 66). \* = significantly different (p < 0.05) from control (unexposed monolayers).

#### 4. Discussion

In this study, we demonstrate that effects of ambient ultrafine particles and engineered nanoparticles on barrier properties of RAECM are strongly dependent on nanoparticle composition, charge and/or concentration. Exposure to UAPS (up to 36 µg/mL) decreased  $R_t$  and  $I_{eq}$  over time significantly, while PNP (negatively and positively charged, up to 706 µg/mL) did not affect  $R_t$  or  $I_{eq}$ . After 1 h of RAECM exposure to SWCNT (up to 88 µg/mL),  $R_t$ decreased significantly but transiently, recovering to initial values after 4–24 h. Exposure to positively charged QD (176 µg/mL) decreased  $R_t$  significantly, but negatively charged QD (up to 176 µg/mL) and lower concentrations of positively charged QD (44 and 88 µg/mL) had no effect on  $R_t$  and  $I_{eq}$ .

Analysis of UAPS composition (Table 1) revealed a high percentage of organic components (mostly hydrophobic) and inorganic hydrophilic compounds (sulfates, nitrates and trace elements and metals). Despite these data on composition, properties of UAPS remain complex and illdefined. Due to this complexity, the mechanisms of UAPS toxicity remain obscure. Disruption of lung alveolar epithelial barrier properties may be caused by multiple factors associated with UAPS, including trace elements and metals, elemental and/or organic carbon content and large overall particle surface area (Oberdorster et al., 2005b). Ultrafine particles may also act as a catalyst for endogenous effects on lung epithelium, leading to increased toxicity and inflammation (Oberdorster, 2001). The larger surface afforded by ultrafine particles compared to larger particles may increase particle surface-dependent reactions (e.g., generation of reactive oxidant species) (Ibald-Mulli et al., 2002; Wichmann et al., 2000). Lack of effects of UAPS on paracellular movement of hydrophilic solutes suggests that the observed decreases in  $R_t$  and  $I_{eq}$  in response to apical UAPS exposure of monolayers reflect primarily changes in transcellular transport properties (e.g., ion channel and pump activities), consistent with the absence of cytotoxicity (no increase in LDH release) after UAPS exposure.

In related studies, ultrafine carbon black particles were reported to cause detectable proinflammatory effects in rat lung, including modest neutrophil influx, protein leak, and modulation of glutathione levels (Donaldson et al., 2001). Ultrafine particles made of low-solubility and lowtoxicity materials were found to be inflammatory in rat lung (Donaldson et al., 2002). Aerosolized ultrafine TiO<sub>2</sub> particles were shown to cause severe bronchoalveolar inflammation (Ferin et al., 1992). Instillation of ultrafine carbon black particles (~14 nm) intratracheally into rat lungs led to a marked increase in lactate dehydrogenase (LDH) levels in bronchoalveolar lavage (BAL) fluid (Li et al., 1999). Intratracheal instillation of diesel exhaust particles ( $\sim$ 30 nm) and an amorphous silicon dioxide (commercially available as Carbosil,  $\sim 7 \text{ nm}$ ) into lungs of male Sprague-Dawley rats increased air-blood barrier permeability, inflammation and edema formation, leading to a strong correlation of plasma viscosity with an index of type I cell injury (reflected as rat type I cell marker found in lung lavage fluid) (Evans et al., 2006). By contrast, healthy mice did not appear to respond to ultrafine carbon and platinum particles (15 and 25 nm, respectively) when inhaled for up to 6 h at  $110 \,\mu\text{g/mL}$  (Oberdorster, 2001). These effects may be dependent on many factors, including physicochemical characteristics of particles. Health effects manifested by ambient particulates are currently thought to be especially important in specific risk groups of persons

who are predisposed to injury by genetic susceptibility, age and/or disease (Kreyling et al., 2006).

OD toxicity appears to be dependent on particle size. charge, concentration and surface coating bioactivity (Derfus et al., 2003; Hardman, 2006; Hoshino et al., 2004; Zhang et al., 2006). Our results show that positively charged QD are more toxic to RAECM compared to negatively charged QD. Cationic amino acids and peptides are known to increase leakage of solutes into alveolar fluid by disruption of tight junctional pathways (Kim and Malik, 2003), suggesting that positively charged nanoparticles may cause injury to air-blood barriers of the lung by related mechanisms. Uncoated QD made of core/shell CdSe/ZnS have been reported to be toxic to cells due to surface oxidation which, through a variety of pathways, leads to release Cd<sup>2+</sup> ions into the cellular environment (Derfus et al., 2003; Zhang et al., 2006). Studies of toxicity of QD with different coatings (neutral, negatively or positively charged) in primary human epidermal keratinocytes showed that negatively and positively charged QD were both toxic up to 20 nM, while neutral QD were not, after 48 h of exposure (Ryman-Rasmussen et al., 2007). These varying results suggest that QD cytotoxicity depends at least in part on specific cellular interactions and particle physicochemical properties.

Our results suggest that SWCNT acutely and transiently affect passive barrier properties, without appreciably affecting active ion transport properties of the alveolar epithelial barrier. Manufactured SWCNT may contain significant amounts of metallic impurities that can serve as catalysts for oxidative stress. The specific physicochemical properties and metallic content (especially iron) of SWCNT may have contributed to the observed effects. After intratracheal instillation in mice (Lam et al., 2004) and rats (Warheit et al., 2004), unpurified and/or purified SWCNT caused epithelial granulomatous reactions. In another study, it was shown that purified SWCNT induce acute inflammatory reactions and decreased bacterial clearance in rats (Shvedova et al., 2005). Unpurified SWCNT exposure led to generation of reactive oxygen species and increased oxidant stress and cytotoxicity in human bronchial epithelial cells (Shvedova et al., 2003). Other studies showed purified and/or unpurified SWCNT causes very low toxicity to the rat alveolar cell line NR8383 (Pulskamp et al., 2007) and human alveolar epithelial cell line A549 (Davoren et al., 2007; Pulskamp et al., 2007). It has been reported that iron content of enriched SWCNT has a significantly more pronounced toxic effect on cultured cell lines (human keratinocytes) than SWCNT treated with iron chelators (Shvedova et al., 2003), consistent with the possibility that metal impurities in SWCNT preparations may be more damaging to cells than SWCNT themselves (Shvedova et al., 2003).

Our studies demonstrate that apical exposure of RAECM to concentrations up to 706  $\mu$ g/mL of positively and negatively charged PNP (20, 100 and 120 nm) cause no significant changes in active and passive transport prop-

erties over 24 h. These findings are in agreement with a recent preliminary report using 46 nm PNP (positive or negative surface charge) in isolated primary rat AT2 cells (Geys et al., 2006). On the other hand, studies with a phagocytic cell line (RAW 264.7) showed that positively charged PNP (60 nm) produces reactive oxygen species, mitochondrial damage and cellular toxicity (although negatively charged PNP have little effect (Xia et al., 2006)), while studies in rat lungs *in vivo* showed that instillation of 64 nm PNP (1 mg) caused significant cell death as measured by BAL LDH levels (Brown et al., 2001). These data suggest that cellular toxicity of PNP may be related at least in part to the specific cell types and experimental milieu in which exposure occurs.

In summary, we have shown that UAPS, QD and SWCNT (but not PNP) can alter the active and/or passive ion transport properties of primary RAECM. We conclude that physicochemical properties of nanoparticles, a result of their varying sources, formation mechanisms, composition or surface charge, are important factors in determining their effects on lung alveolar epithelial barrier properties. Further mechanistic studies with defined nanoparticles may provide additional insights into how nanoparticles interact with alveolar epithelial cells and lead to modulation of barrier properties.

#### Acknowledgements

The authors appreciate the technical assistance of Juan Raymond Alvarez. E.D. Crandall is Hastings Professor and Kenneth T. Norris Jr. Chair of Medicine. This work was supported in part by the Hastings Foundation, research grants HL 38578, HL 38621, HL 38658, HL 62569 and HL 64365 from the National Institutes of Health, and the Southern California Particle Center and Supersite (SCPCS) funded by the Environmental Protection Agency (EPA) under the STAR Program. Research described herein has not been subjected to the EPA required peer and policy review and therefore does not necessarily reflect the views of EPA, and no official endorsement should be inferred. Mention of trade names or commercial products does not constitute an endorsement or recommendation for use.

#### References

- Adamson, I.Y., Bowden, D.H., 1975. Derivation of type 1 epithelium from type 2 cells in the developing rat lung. Laboratory Investigation 32, 736–745.
- Alfaro-Moreno, E., Martinez, L., Garcia-Cuellar, C., Bonner, J.C., Murray, J.C., Rosas, I., Rosales, S.P., Osornio-Vargas, A.R., 2002. Biologic effects induced in vitro by PM10 from three different zones of Mexico City. Environmental Health Perspectives 110, 715–720.
- Borok, Z., Danto, S.I., Zabski, S.M., Crandall, E.D., 1994. Defined medium for primary culture de novo of adult rat alveolar epithelial cells. In Vitro Cellular & Developmental Biology. Animal 30A, 99– 104.
- Borok, Z., Hami, A., Danto, S.I., Zabski, S.M., Crandall, E.D., 1995. Rat serum inhibits progression of alveolar epithelial cells toward the type I

cell phenotype in vitro. American Journal of Respiratory Cell and Molecular Biology 12, 50–55.

- Bronikowski, M.J., Willis, P.A., Colbert, D.T., Smith, K.A., Smalley, R.E., 2001. Gas-phase production of carbon single-walled nanotubes from carbonmonoxide via the HiPco process: A parametric study. Journal of Vacuum Science and Technology 19, 1800–1805.
- Brown, D.M., Wilson, M.R., MacNee, W., Stone, V., Donaldson, K., 2001. Size-dependent proinflammatory effects of ultrafine polystyrene particles: a role for surface area and oxidative stress in the enhanced activity of ultrafines. Toxicology and Applied Pharmacology 175, 191– 199.
- Calcabrini, A., Meschini, S., Marra, M., Falzano, L., Colone, M., De Berardis, B., Paoletti, L., Arancia, G., Fiorentini, C., 2004. Fine environmental particulate engenders alterations in human lung epithelial A549 cells. Environmental Research 95, 82–91.
- Cassee, F.R., Muijser, H., Duistermaat, E., Freijer, J.J., Geerse, K.B., Marijnissen, J.C., Arts, J.H., 2002. Particle size-dependent total mass deposition in lungs determines inhalation toxicity of cadmium chloride aerosols in rats. Application of a multiple path dosimetry model. Archives of Toxicology 76, 277–286.
- Cheek, J.M., Evans, M.J., Crandall, E.D., 1989a. Type I cell-like morphology in tight alveolar epithelial monolayers. Experimental Cell Research 184, 375–387.
- Cheek, J.M., Kim, K.J., Crandall, E.D., 1989b. Tight monolayers of rat alveolar epithelial cells: bioelectric properties and active sodium transport. American Journal of Physiology 256, C688–C693.
- Danto, S.I., Zabski, S.M., Crandall, E.D., 1992. Reactivity of alveolar epithelial cells in primary culture with type I cell monoclonal antibodies. American Journal of Respiratory Cell and Molecular Biology 6, 296–306.
- Davoren, M., Herzog, E., Casey, A., Cottineau, B., Chambers, G., Byrne, H.J., Lyng, F.M., 2007. In vitro toxicity evaluation of single walled carbon nanotubes on human A549 lung cells. Toxicology In Vitro 21, 438–448.
- Derfus, A.M., Chan, W.C.W., Bhatia, S.N., 2003. Probing the cytotoxicity of semiconductor quantum dots. Nano Letters 4, 11–18.
- Donaldson, K., Stone, V., Clouter, A., Renwick, L., MacNee, W., 2001. Ultrafine particles. Occupational and Environmental Medicine 58, 211–216, 199.
- Donaldson, K., Brown, D., Clouter, A., Duffin, R., MacNee, W., Renwick, L., Tran, L., Stone, V., 2002. The pulmonary toxicology of ultrafine particles. Journal of Aerosol Medicine 15, 213–220.
- Elbert, K.J., Schafer, U.F., Schafers, H.J., Kim, K.J., Lee, V.H., Lehr, C.M., 1999. Monolayers of human alveolar epithelial cells in primary culture for pulmonary absorption and transport studies. Pharmaceutical Research 16, 601–608.
- Evans, S.A., Al-Mosawi, A., Adams, R.A., Berube, K.A., 2006. Inflammation, edema, and peripheral blood changes in lung-compromised rats after instillation with combustion-derived and manufactured nanoparticles. Experimental Lung Research 32, 363–378.
- Ferin, J., Oberdorster, G., Penney, D.P., 1992. Pulmonary retention of ultrafine and fine particles in rats. American Journal of Respiratory Cell and Molecular Biology 6, 535–542.
- Geys, J., Coenegrachts, L., Vercammen, J., Engelborghs, Y., Nemmar, A., Nemery, B., Hoet, P.H., 2006. In vitro study of the pulmonary translocation of nanoparticles: a preliminary study. Toxicology Letters 160, 218–226.
- Ghio, A.J., Devlin, R.B., 2001. Inflammatory lung injury after bronchial instillation of air pollution particles. American Journal of Respiratory and Critical Care Medicine 164, 704–708.
- Gurr, J.R., Wang, A.S., Chen, C.H., Jan, K.Y., 2005. Ultrafine titanium dioxide particles in the absence of photoactivation can induce oxidative damage to human bronchial epithelial cells. Toxicology 213, 66–73.
- Gutierrez-Castillo, M.E., Roubicek, D.A., Cebrian-Garcia, M.E., De Vizcaya-Ruiz, A., Sordo-Cedeno, M., Ostrosky-Wegman, P., 2006. Effect of chemical composition on the induction of DNA damage by

urban airborne particulate matter. Environmental and Molecular Mutagenesis 47, 199–211.

- Hardman, R., 2006. A toxicologic review of quantum dots: toxicity depends on physicochemical and environmental factors. Environmental Health Perspectives 114, 165–172.
- Hoshino, A., Fujioka, K., Oku, T., Nakamura, S., Suga, M., Yamaguchi, Y., Suzuki, K., Yasuhara, M., Yamamoto, K., 2004. Quantum dots targeted to the assigned organelle in living cells. Microbiology and Immunology 48, 985–994.
- Huang, G.W., Chen, C.Y., Wu, K.C., Ahmed, M.O., Chou, P.T., 2004. One-pot synthesis and characterization of high-quality CdSe/ZnX (X = S, Se) nanocrystals via the CdO precursor. Journal of Crystal Growth 265, 250–259.
- Ibald-Mulli, A., Wichmann, H.E., Kreyling, W., Peters, A., 2002. Epidemiological evidence on health effects of ultrafine particles. Journal of Aerosol Medicine 15, 189–201.
- Kim, K.J., Malik, A.B., 2003. Protein transport across the lung epithelial barrier. American Journal of Physiology 284, L247–L259.
- Kim, K.J., Borok, Z., Crandall, E.D., 2001a. A useful in vitro model for transport studies of alveolar epithelial barrier. Pharmaceutical Research 18, 253–255.
- Kim, S., Jaques, P.A., Chang, M., Froines, J.R., Sioutas, C., 2001b. Versatile aerosol concentration enrichment system (VACES) for simultaneous in vivo and in vitro evaluation of toxic effects of ultrafine, fine and coarse ambient particles. Part I: Development and laboratory characterization. Journal of Aerosol Science 32, 1281–1297.
- Kim, S., Jaques, P.A., Chang, M., Froines, J.R., Sioutas, C., 2001c. Versatile aerosol concentration enrichment system (VACES) for simultaneous in vivo and in vitro evaluation of toxic effects of ultrafine, fine and coarse ambient particles. Part II: Field evaluation. Journal of Aerosol Science 32, 1299–1314.
- Kreyling, W.G., Semmler-Behnke, M., Moller, W., 2006. Ultrafine particle–lung interactions: does size matter? Journal of Aerosol Medicine 19, 74–83.
- Lam, C., James, J.T., McCluskey, R., Hunter, R.L., 2004. Pulmonary toxicity of single-wall carbon nanotubes in mice 7 and 90 days after intratracheal instillation. Respiratory Toxicology 77, 126–134.
- Li, X.Y., Brown, D., Smith, S., MacNee, W., Donaldson, K., 1999. Shortterm inflammatory responses following intratracheal instillation of fine and ultrafine carbon black in rats. Inhalation Toxicology 11, 709–731.
- Li, N., Sioutas, C., Cho, A., Schmitz, D., Misra, C., Sempf, J., Wang, M., Oberley, T., Froines, J., Nel, A., 2003. Ultrafine particulate pollutants induce oxidative stress and mitochondrial damage. Environmental Health Perspectives 111, 455–460.
- Magrez, A., Kasas, S., Salicio, V., Pasquier, N., Seo, J.W., Celio, M., Catsicas, S., Schwaller, B., Forro, L., 2006. Cellular toxicity of carbonbased nanomaterials. Nano Letters 6, 1121–1125.
- Monn, C., Becker, S., 1999. Cytotoxicity and induction of proinflammatory cytokines from human monocytes exposed to fine (PM2.5) and coarse particles (PM10-2.5) in outdoor and indoor air. Toxicology and Applied Pharmacology 155, 245–252.
- Muldoon, L.L., Sandor, M., Pinkston, K.E., Neuwelt, E.A., 2005. Imaging, distribution, and toxicity of superparamagnetic iron oxide magnetic resonance nanoparticles in the rat brain and intracerebral tumor. Neurosurgery 57, 785–796.
- Nemmar, A., Vanbilloen, H., Hoylaerts, M.F., Hoet, P.H., Verbruggen, A., Nemery, B., 2001. Passage of intratracheally instilled ultrafine particles from the lung into the systemic circulation in hamster. American Journal of Respiratory and Critical Care Medicine 164, 1665–1668.
- Nemmar, A., Hoet, P.H., Vanquickenborne, B., Dinsdale, D., Thomeer, M., Hoylaerts, M.F., Vanbilloen, H., Mortelmans, L., Nemery, B., 2002. Passage of inhaled particles into the blood circulation in humans. Circulation 105, 411–414.
- Oberdorster, G., 2001. Pulmonary effects of inhaled ultrafine particles. International Archives of Occupational and Environmental Health 74, 1–8.

- Oberdorster, G., Gelein, R.M., Ferin, J., Weiss, B., 1995. Association of particulate air pollution and acute mortality: involvement of ultrafine particles? Inhalation Toxicology 7, 111–124.
- Oberdorster, G., Maynard, A., Donaldson, K., Castranova, V., Fitzpatrick, J., Ausman, K., Carter, J., Karn, B., Kreyling, W., Lai, D., Olin, S., Monteiro-Riviere, N., Warheit, D., Yang, H., 2005a. Principles for characterizing the potential human health effects from exposure to nanomaterials: elements of a screening strategy. Particle and Fibre Toxicology 2, 8.
- Oberdorster, G., Oberdorster, E., Oberdorster, J., 2005b. Nanotoxicology: an emerging discipline evolving from studies of ultrafine particles. Environmental Health Perspectives 113, 823–839.
- Osornio-Vargas, A.R., Bonner, J.C., Alfaro-Moreno, E., Martinez, L., Garcia-Cuellar, C., Ponce-de-Leon Rosales, S., Miranda, J., Rosas, I., 2003. Proinflammatory and cytotoxic effects of Mexico City air pollution particulate matter in vitro are dependent on particle size and composition. Environmental Health Perspectives 111, 1289–1293.
- Peters, A., Dockery, D.W., Muller, J.E., Mittleman, M.A., 2001. Increased particulate air pollution and the triggering of myocardial infarction. Circulation 103, 2810–2815.
- Pulskamp, K., Diabate, S., Krug, H.F., 2007. Carbon nanotubes show no sign of acute toxicity but induce intracellular reactive oxygen species in dependence on contaminants. Toxicology Letters 168, 58–74.
- Ryman-Rasmussen, J.P., Riviere, J.E., Monteiro-Riviere, N.A., 2007. Surface coatings determine cytotoxicity and irritation potential of quantum dot nanoparticles in epidermal keratinocytes. Journal of Investigative Dermatology 127, 143–153.
- Sayes, C.M., Liang, F., Hudson, J.L., Mendez, J., Guo, W., Beach, J.M., Moore, V.C., Doyle, C.D., West, J.L., Billups, W.E., Ausman, K.D., Colvin, V.L., 2006. Functionalization density dependence of singlewalled carbon nanotubes cytotoxicity in vitro. Toxicology Letters 161, 135–142.
- Shvedova, A.A., Castranova, V., Kisin, E.R., Schwegler-Berry, D., Murray, A.R., Gandelsman, V.Z., Maynard, A., Baron, P., 2003. Exposure to carbon nanotube material: assessment of nanotube cytotoxicity using human keratinocyte cells. Journal of Toxicology and Environmental Health. Part A 66, 1909–1926.

- Shvedova, A.A., Kisin, E.R., Mercer, R., Murray, A.R., Johnson, V.J., Potapovich, A.I., Tyurina, Y.Y., Gorelik, O., Arepalli, S., Schwegler-Berry, D., Hubbs, A.F., Antonini, J., Evans, D.E., Ku, B.K., Ramsey, D., Maynard, A., Kagan, V.E., Castranova, V., Baron, P., 2005. Unusual inflammatory and fibrogenic pulmonary responses to singlewalled carbon nanotubes in mice. American Journal of Physiology 289, L698–L708.
- Sun, Q., Wang, A., Jin, X., Natanzon, A., Duquaine, D., Brook, R.D., Aguinaldo, J.G., Fayad, Z.A., Fuster, V., Lippmann, M., Chen, L.C., Rajagopalan, S., 2005. Long-term air pollution exposure and acceleration of atherosclerosis and vascular inflammation in an animal model. JAMA: Journal of the American Medical Association 294, 3003–3010.
- Topinka, J., Schwarz, L.R., Wiebel, F.J., Cerna, M., Wolff, T., 2000. Genotoxicity of urban air pollutants in the Czech Republic. Part II. DNA adduct formation in mammalian cells by extractable organic matter. Mutation Research 469, 83–93.
- Vedal, S., 1997. Ambient particles and health: lines that divide. Journal of the Air & Waste Management Association 47, 551–581.
- Warheit, D.B., Laurence, B.R., Reed, K.L., Roach, D.H., Reynolds, G.A., Webb, T.R., 2004. Comparative pulmonary toxicity assessment of single-wall carbon nanotubes in rats. Toxicological Sciences 77, 117–125.
- Wichmann, H.E., Spix, C., Tuch, T., Wolke, G., Peters, A., Heinrich, J., Kreyling, W.G., Heyder, J., 2000. Daily mortality and fine and ultrafine particles in Erfurt, Germany part I: role of particle number and particle mass. Research Report (Health Effects Institute), 5–87, discussion 87–94.
- Xia, T., Kovochich, M., Brant, J., Hotze, M., Sempf, J., Oberley, T., Sioutas, C., Yeh, J.I., Wiesner, M.R., Nel, A.E., 2006. Comparison of the abilities of ambient and manufactured nanoparticles to induce cellular toxicity according to an oxidative stress paradigm. Nano Letters 6, 1794–1807.
- Zhang, T., Stilwell, J.L., Gerion, D., Ding, L., Elboudwarej, O., Cooke, P.A., Gray, J.W., Alivisatos, A.P., Chen, F.F., 2006. Cellular effect of high doses of silica-coated quantum dot profiled with high throughput gene expression analysis and high content cellomics measurements. Nano Letters 6, 800–808.