Sensitivity of the Community Multiscale Air Quality (CMAQ) Model v4.7 Results for the Eastern United States to MM5 and WRF Meteorological Drivers

K. Wyat Appel * , Shawn J. Roselle, Robert C. Gilliam and Jonathan E. Pleim

Atmospheric Modeling and Analysis Division, National Exposure Research Laboratory, Office of Research and Development, U.S. Environmental Protection Agency, RTP, NC 27711

Corresponding author address: K. Wyat Appel, US EPA, 109 TW Alexander Dr., Research Triangle Park, NC 27711; Tel.: +19195410757; fax: +19195411379; E-mail address: appel.wyat@epa.gov

Abstract

This paper presents a comparison of the operational performances of two Community Multiscale Air Quality (CMAQ) model v4.7 simulations that utilize input data from the 5th-generation Mesoscale Model (MM5) and the Weather Research and Forecasting (WRF) meteorological models. Two sets of CMAQ model simulations were performed for January and August 2006. One set utilized MM5 meteorology (MM5-CMAQ) and the other utilized WRF meteorology (WRF-CMAQ), while all other model inputs and options were kept the same. For January, predicted ozone (O₃) mixing ratios were higher in the Southeast and lower Mid-west regions in the WRF-CMAQ simulation, resulting in slightly higher bias and error as compared to the MM5-CMAQ simulations. The higher predicted O₃ mixing ratios are attributed to less dry deposition of O₃ in the WRF-CMAQ simulation due to differences in the calculation of the vegetation fraction between the MM5 and WRF models. The WRF-CMAQ results showed better performance for particulate sulfate (SO₄²-), similar performance for nitrate (NO₃-), and slightly worse performance for nitric acid (HNO₃), total carbon (TC) and total fine particulate (PM_{2.5}) mass than the corresponding MM5-CMAQ results. For August, predictions of O3 were notably higher in the WRF-CMAQ simulation, particularly in the southern United States, resulting in increased model bias. Concentrations of predicted particulate SO₄² were lower in the region surrounding the Ohio Valley and higher along the Gulf of Mexico in the WRF-CMAQ simulation, contributing to poorer model performance. The primary causes of the differences in the MM5-CMAQ and WRF-CMAQ simulations appear to be due to differences in the calculation of wind speed, planetary boundary layer height, cloud cover and the friction velocity (u*) in the MM5 and WRF model simulations, while differences in the calculation of vegetation fraction and several other parameters result in smaller differences in the predicted CMAQ model concentrations. The performance for SO₄², NO₃ and NH₄ wet deposition was similar for both simulations for January and August.

1. Introduction

Air quality models, such as the Community Multiscale Air Quality (CMAQ) modeling system (Byun and Schere, 2006) and the Comprehensive Air Quality Model with extensions (CAMx) (ENVIRON, 2008), require gridded, high resolution (both temporally and spatially) meteorological data in order to accurately predict the transformation, transport and fate of pollutants in the atmosphere. Gridded Eulerian meteorological models, such as the 5th Generation Mesoscale Model (MM5; Grell et al., 1994) and the Weather Research and Forecasting model (WRF; Skamarock et al., 2008), are used to provide the meteorological data required by air quality models.

For the past 15 years, MM5 has been used to provide meteorological data for air quality simulations. The modular design of MM5 allows users to choose among various physics options such as: land-surface models (LSM), planetary boundary layer (PBL), radiation, microphysics and cloud schemes in order to optimize the model for a specific application. However, releases of new versions of MM5 by the community have ceased since the WRF model has taken its place. The WRF model incorporates the same capabilities as the MM5 model, but includes various improvements in the underlying dynamics of the model (e.g. mass conservation) along with updated physics, including new versions of the LSM, PBL, radiation and cloud microphysics schemes (add reference).

Although the WRF model has been available for several years and is being used operationally by the National Centers for Environmental Prediction (NCEP) and many other research groups, the model has seen limited use for retrospective air quality modeling applications. Until recently, operational performance of retrospective WRF model simulations has lagged that of MM5

simulations, due mostly to a lack of a comparable analysis nudging scheme. Analysis nudging is widely used by the air quality community to improve the performance of the meteorological simulations used in retrospective air quality simulations. A recently released version of an objective analysis utility for WRF (Obsgrid; Deng et al., 2008) improves the operational performance of retrospective WRF model simulations, making the performance comparable to MM5 (Gilliam and Pleim, 2010).

While other studies have compared the performance of air quality model predictions using different meteorological models (e.g. Smyth et al., 2006; de Meij et al., 2009), no studies have specifically compared the performance of MM5 and WRF driven CMAQ model simulations. This study has two main objectives. One is to test the WRF-CMAQ modeling system to insure that no major model performance issues exist (e.g. issues that would prohibit using the WRF model with the CMAQ model), as using WRF model data to drive the CMAQ model is a relatively new option and performance issues may exist. The second objective is to inform the CMAQ model user and development communities of the differences in CMAQ model performance they might encounter when transitioning from MM5 to the WRF model as the meteorological driver for CMAQ. This study specifically examines the operational performance of two sets of January and August 2006 CMAQ simulations, with one set using meteorological data provided by MM5 (MM5-CMAQ) and the other using data provided by the WRF model (WRF-CMAQ). The configurations of the meteorological models used are typical of those that would be used for regulatory air quality simulations and no modifications have been made to the underlying codes of either models to make them more consistent with each other (e.g. "out-ofthe-box" configurations are used). The performance results for each simulation are presented and where possible the likely reasons for large differences in performance are discussed.

2. Methodology

2.1. MM5 and WRF Model Simulations

MM5 and WRF model simulations were performed for the eastern United States for January and August 2006 (with a 10 day spin-up period in the previous month) that utilize a horizontal grid with 12-km by 12-km grid cells and 34 vertical layers extending up to 100 hPa. Boundary conditions for both the MM5 and WRF simulations were provided directly by the 12-km North American Model (NAM) simulation for the same time period. The details provided here regarding the MM5 and WRF model simulations are based on Gilliam and Pleim (2010), which compares the performance of similarly configured MM5 and WRF simulations as used in this study.

The MM5 simulation utilized version 3.7.4 of the model, with the Asymmetric Convective Model 2 (ACM2; Pleim, 2007a,b) PBL model, Pleim-Xiu (PX; Xiu and Pleim, 2001; Pleim and Xiu, 1995) LSM, Dudhia shortwave radiation scheme (Dudhia, 1989), RRTM longwave radiation scheme (Mlawer et al., 1997), Kain-Fritsch 2 (KF2; Kain, 2004) sub-grid convective scheme and the Reisner-2 (Reisner et al., 1998) explicit microphysics scheme. The PX LSM included indirect soil moisture and temperature nudging (Pleim and Xiu, 2003; Gilliam and Pleim, 2010). The similarly configured WRF model simulation utilized the Advanced Research WRF (ARW) core version 3.0 (Skamarock et al., 2008), with the ACM2 PBL model, PX LSM, Dudhia shortwave and RRTM longwave radiation schemes, KF2 sub-grid convective scheme and the Thompson (Thompson et al., 2004) microphysics scheme. A summary of the configuration options for the MM5 and WRF model simulations is shown in Table 1. These options were chosen in order to obtain consistent performance for the two simulations, and are those options typically used by the MM5 and WRF communities, especially for retrospective air quality simulations.

Several distinct nudging strategies are used in both the MM5 and WRF simulations that employ the PX LSM. These include four dimensional data assimilation (FDDA) and indirect soil moisture and temperature (T) nudging. In both the MM5 and WRF model simulations, FDDA is essentially the same in terms of the analyses used and the nudging configuration, which follow after Stauffer et al. (1991) and Otte (2008a) and are described in detail for these particular simulations by Gilliam and Pleim (2010). The reanalysis fields considered are the wind components in the east-west and north-south directions, T and water vapor mixing ratio (w). The winds are nudged at all model levels, while T and w are only nudged above the model simulated PBL in the WRF and MM5 simulations. Indirect soil moisture and soil T nudging were used by both the MM5 and WRF models. It should be noted that for the MM5 simulation the RAWINS tool was used to create the 2-meter (m) T and 2-m w analyses, while for the WRF simulation the Obsgrid tool, a relative of RAWINS, was used. Gilliam and Pleim (2010) demonstrated that while these tools are very similar and ingest the same base analysis and surface observations, Obsgrid produces a reanalysis with an overall better comparison to the surface observations.

2.2. CMAQ Model Simulations

The CMAQ model simulations were performed using CMAQv4.7 (Foley et at., 2009) for the eastern United States for January and August 2006 using a three day spin-up period in the previous month on the same grid as the meteorology models except that its horizontal dimensions were reduced by 5 grid cells on each of the 4 lateral boundaries to avoid spurious boundary artifacts in the meteorology simulations. CMAQ was configured using the AERO5 aerosol module and the CB05 chemical mechanism with chlorine chemistry extensions (Yarwood et al., 2005) and the ACM2 PBL scheme. The vertical layers for the CMAQ simulations match those of the meteorological simulations. Version 3.4.1 of the Meteorology-Chemistry Interface Processor (MCIP; Otte et al., 2005) was used to process the MM5 and WRF meteorology for use with

CMAQ. The simulations used a 2005 base year emissions inventory which was updated with year specific mobile emissions and Continuous Emissions Monitoring System (CEMS) data for point emissions for 2006. The latest version of the CMAQ model includes the option to calculate biogenic and plume rise emissions in-line during the simulation, an option that was used for this study.

2.3. Model Assessment Techniques

The evaluation of the MM5, WRF and CMAQ model simulations was done primarily using the Atmospheric Model Evaluation Tool (AMET) (Appel and Gilliam, 2008). Meteorological predictions of 2-m T, 2-m w and 10-m wind speed (WS) are paired in space and time with observations from the Meteorological Assimilation Data Ingest System (MADIS; http://madis.noaa.gov) database. The performance of the predictions is then assessed using available analyses in the AMET. Additionally, predicted monthly precipitation is compared against observations from the National Precipitation Analysis (NPA), which is a blend of radar estimated precipitation and rain gauge data (Fulton, 1998; Seo 1998a; Seo 1998b).

The CMAQ model predictions are paired in space and time with observations from the Environmental Protections Agency's (EPA) Air Quality System (AQS) for O₃, the Interagency Monitoring of PROtected Visual Environments (IMPROVE) network, the Chemical Speciation Network (CSN; previously called the Speciation Trends Network(STN)) and the Clean Air Status and Trends Network (CASTNet) for fine particulate matter, and the National Atmospheric Deposition Program (NADP) network for wet deposition species. Observations from the AQS (353 sites in January; 861 sites in August) are hourly; observations from the IMPROVE network (90 sites) and the CSN (174 sites in January; 157 sites in August) are daily average concentrations

available every third day; CASTNet (67 sites) observations are weekly average concentrations, while the NADP network (202 sites) observations are weekly accumulated values.

Several statistical quantities are provided that assess the model bias and error. Root Mean Square Error (RMSE), Normalized Mean Error (NME), Normalized Median Error (NMdnE), Mean Error (ME) and Median Error (MdnE) are used to assess model error. Normalized Mean Bias (NMB), Normalized Median Bias (NMdnB), Mean Bias (MB) and Median Bias (MdnB) are used to assess model bias. The MdnB, MdnE, NMdnB and NMdnE are defined below as:

$$MdnB = median(C_M - C_O)_N \tag{1}$$

$$MdnE = median \left| C_M - C_O \right|_{N} \tag{2}$$

$$NMdnB = \frac{median(C_M - C_O)_N}{median(C_O)_N} *100\%$$
(3)

$$NMdnE = \frac{median|C_M - C_O|_N}{median(C_O)_N} *100\%$$
(4)

where C_M and C_O are modeled and observed concentrations, respectively, and N is the total number of model/observation pairs. In equations (2) and (4) which calculate error, the absolute value of the difference between the modeled and observed concentration is used, denoted in the equations by the vertical bars. Median is preferred here over mean since median gives a better representation of the central tendency of the data than the mean when analyzing data with non-normal distributions, which the observed PM species data often are. The metrics are normalized by the median of observed data to avoid instances of extremely large biases and errors that can occur when normalizing by the individual observations for observed concentrations that are very low, which is possible with the species being examined here. Additional details regarding these

statistics and how the observations from the various observing networks are paired with CMAQ predictions and are used in the AMET can be found in Appel et al. (2007, 2008).

3. MM5 and WRF Model Performance Assessment

Since the objective of this study is to examine the differences between the MM5-CMAQ and WRF-CMAQ predictions, it is important to determine what, if any, significant differences exist between the MM5 and WRF model simulations from an operational performance perspective. This section provides limited comparison of the MM5 and WRF model performance, since a more detailed assessment of the MM5 and WRF model performance can be found in Gilliam and Pleim (2010).

3.1. January

Fig. 1a presents a comparison of the daily RMSE for 2-m T, 2-m w and 10-m WS for January for the MM5 and WRF model simulations. The RMSE for all three variables is very similar for January, although there are some periods where the RMSE for 2-m T is notably higher for the MM5 simulation. Fig. 1b presents a comparison of the diurnal (hourly) bias for the same three variables for January. The WRF model simulation has much lower bias for 2-m T during the nighttime hours (8PM – 8AM EST; 01-13 UTC) than the MM5 simulation, while the daytime (8AM – 8PM EST; 13-01 UTC) bias is similar for the two simulations. The w bias is slightly lower in the WRF model simulation throughout most of the day, while the bias in 10-m WS is generally lower in the MM5 simulation. These analyses suggest that the WRF model is generally performing as well as the MM5 for these key meteorological variables for January.

A comparison of the observed accumulated monthly precipitation versus MM5 and WRF predicted precipitation for January is provided in Figs. 2a-c. The spatial pattern and amount of predicted precipitation from the MM5 (Fig. 2b) and WRF (Fig. 2c) model simulations are similar over land, and are generally comparable to the observed precipitation (Fig 2a). The largest difference in predicted precipitation between the two simulations occurs over the Gulf of Mexico and off the east coast of the United States, where the WRF model predicts much greater precipitation than MM5. It is not possible to determine which model is more correct, since the radar-based precipitation dataset is not available beyond the coast. However, the impact from the differences in the offshore precipitation on CMAQ predictions should be relatively small.

3.2. August

Fig. 1c shows a comparison of the daily RMSE for 2-m T, w and 10-m WS for August for the MM5 and WRF simulations. The RMSE values for all three variables track very close to each other for most of the month. The RMSE for w is higher in both simulations for the first third of the month as compared to the other two-thirds due to the higher moisture of the air-mass at the beginning of the month, after which a dryer air-mass dominated most of the eastern United States. The diurnal bias in 2-m T (Fig. 1d) is higher during the nighttime hours and lower during the daytime hours for the WRF simulation, while the w bias is significantly reduced in the WRF model simulation during most of the day. Although the bias in 10-m WS is similar throughout the afternoon, the MM5 simulation has slightly less bias during the overnight and early morning hours. See Gilliam and Pleim (2010) for additional details regarding the causes for the differences in performance.

Comparison of the monthly precipitation for August (Figs. 2d-f) shows greater variability compared to January, which is expected due to the convective nature of summertime

precipitation. The WRF model simulation (Fig. 2f) predicts greater precipitation over the southeast United States and offshore as compared to the MM5 simulation (Fig. 2e) and the observations (Fig. 2d), while the MM5 simulation has slightly higher predicted precipitation over the lower Midwest as compared to WRF model simulation. Both models overpredict precipitation in the lower Midwest and underpredict precipitation in the upper Midwest and western Great Lakes regions. Overall, the performance of the MM5 and WRF model simulations for January and August is similar, and generally compares well with the observations. This result is similar to the conclusions of Gilliam and Pleim (2010), in which they note similar performance for the MM5 and WRF model simulations for the two months.

4. CMAQ Model Performance Assessment

4.1. January

4.1.1. Ozone (O₃)

For January, O₃ predictions are generally higher in the WRF-CMAQ simulation, particularly across the southern portion of the model domain, where increases in monthly average O₃ mixing ratios of 2 ppb or more are present (Fig. 3a). The result is larger bias and error in the WRF-CMAQ simulation versus the MM5-CMAQ simulation (Table 2). Both simulations overpredict hourly O₃ on average, indicated by the positive NMdnB and MdnB for both simulations; however the NMdnB and MdnB are 3.7 % and 0.85 ppb higher for the WRF-CMAQ simulation, respectively. For maximum 8-hr average O₃, the NMdnB and MdnB are 3.7 % and 1.19 ppb higher for the WRF-CMAQ simulation. The error is similar between the two simulations for both measures of O₃.

Comparison of the O₃ dry deposition from the two simulations revealed that the higher predicted O₃ mixing ratios over the southern portion of the domain in WRF-CMAQ simulation are due to less O₃ dry deposition in the WRF-CMAQ simulation, which results in higher ambient O₃ mixing ratios. There are significant differences in the way the vegetation fraction and leaf area index (LAI) are parameterized in the PX LSM between the MM5 and WRF implementations. Both models use satellite-derived vegetation coverage to scale these vegetation parameters in areas dominated by crops. However, the parameterizations differ such that vegetation fraction and LAI are set to minimum values in the winter in all areas in WRF but maintain higher values in the southern-most areas (Gulf coast and Florida) in MM5. The result is less O₃ dry deposition in the WRF-CMAQ simulation due to less stomatal uptake (a result of the less vegetation and LAI) as compared to the MM5-CMAQ simulation, which in turn results in higher ambient O₃ mixing ratios. The WRF parameterization is being re-assessed and may be revised to be more like the MM5 parameterization in the future.

4.1.2. Fine Particulate Sulfate (SO₄²-)

Fig. 3b shows the predicted monthly average concentrations of particulate SO_4^{2-} for January between the two CMAQ simulations. Predictions of SO_4^{2-} are generally higher in the WRF-CMAQ simulation, with the exception of a small area off the coast of southern Florida. The largest differences occur over the northern portion of the domain, where areas of greater than 1 $\mu g/m^3$ difference in monthly average SO_4^{2-} exist. While both simulations underpredict SO_4^{2-} on average (Table 2), the underprediction is smaller in WRF-CMAQ simulation, with a NMdnB that is 4.6-10.0% lower and a MdnB that is 0.05-0.20 $\mu g/m^3$ lower than in the MM5-CMAQ simulation. The error is also smaller in the WRF-CMAQ simulation, with the NMdnE 1.1-5.0% lower and the MdnE 0.03-0.10 $\mu g/m^3$ lower than the MM5-CMAQ simulation.

The higher predicted concentrations of particulate SO_4^{2-} in the WRF-CMAQ simulation appear to be related primarily to a combination of greater predicted cloud fraction and less SO_4^{2-} wet deposition (Fig. 4b) than in the MM5-CMAQ simulation. A comparison of the resolved clouds between the MM5 and WRF model simulations reveal a large area over the upper Midwest and central Canada where the predicted cloud fraction in WRF is notably greater than MM5. The result is more in-cloud aqueous SO_4^{2-} production in that region, which results in the higher predicted SO_4^{2-} concentrations shown in Fig. 3b. The higher SO_4^{2-} concentrations along the east coast of the United States and in Louisiana are also related to differences in the predicted cloud fraction. In the Northeast and eastern Canada, less SO_4^{2-} wet deposition (Fig. 4b) in the WRF-CMAQ simulation results in higher particulate SO_4^{2-} concentrations in that region.

4.1.3. Fine Particulate Nitrate (NO₃) and Total Nitrate (TNO₃)

NO₃ tends to constitute the largest component of fine particulate mass in the eastern United States during the cold season. Fig. 3c shows the predicted monthly average NO₃ concentrations for the two simulations for January, along with the difference between the two model simulations. The WRF-CMAQ simulation predicts higher NO₃ concentrations on average; however the differences are generally small, with only a few localized areas where the differences reach 1 µg/m³ or greater. Since NO₃ is underpredicted in both simulations (Table 2), the higher predicted concentrations in the WRF-CMAQ simulation result in an improvement of both the bias and error. The NMdnB is more than 10 % lower at the CSN sites in the WRF-CMAQ simulation as compared to the MM5-CMAQ simulation, while the difference in NMdnB at IMPROVE network sites is less than a percent. However, the difference in the NMdnE is larger at the IMPROVE network sites (5.4 %) than at the CSN sites (1.7 %).

For TNO₃ (Fig. 3d), the differences between the two simulations are considerably larger and more widespread than for NO₃ alone, indicating significant differences in the HNO₃ predictions. The higher predicted HNO₃ concentrations in the WRF-CMAQ simulation result in an increase in the TNO₃ bias compared to the MM5-CMAQ simulation, as TNO₃ was already overpredicted in both simulations. The NMdnB and MdnB for TNO₃ at CASTNet sites are 9.1 % and 0.21 µg/m³ higher, respectively, in the WRF-CMAQ simulation. The NMdbE and MdnE are also higher in the WRF-CMAQ simulation, but to a slightly lesser degree. The difference in predicted TNO₃ concentrations may be the result of differences in predicted wind speeds, PBL heights and the overall stability between the two simulations. Comparisons of surface wind speeds showed that over land the wind speeds are on average lower in the WRF simulation in January (Fig. 1b), which results in less mixing and hence higher surface concentrations for the various pollutants. An examination of PBL heights between the two simulations showed that PBL heights over land are on average lower in the WRF simulation, which would tend to concentrate pollutants at the surface and lead to higher concentrations as well.

Additionally, a difference in the calculation of dry deposition velocity, which is very high for HNO₃ and limited primarily by the aerodynamic resistance, likely plays a smaller role in the difference in predicted HNO₃ between the two simulations in January. Aerodynamic resistance is strongly dependent on the friction velocity (u*) that is, on average, higher in MM5 than WRF in January (Fig. 4a). The higher u* values in MM5, particularly at night when wind speed is often very light, leads to greater dry deposition of HNO₃ in the MM5-CMAQ simulation. In the MM5 model, the minimum wind speed value in the u* calculation is set to 1.0 m/s, while in the WRF model the minimum is set to 0.1 m/s. It was suspected that the higher minimum value for wind speed in MM5 was responsible for the higher u* values. To test this, a simulation was performed for January in which the minimum wind speed value in MM5 was changed from 1.0 m/s to 0.1 m/s (the same as the WRF model). However, the change in the minimum wind speed threshold

resulted in little change in the calculated u_{*} values in the MM5 simulation. Additional analysis into the differences in u_{*} suggest that a combination of lower wind speeds (Fig. 4b) and smaller surface roughness lengths (Fig. 4c) in the WRF model simulation may be primarily responsible for the lower u_{*} values.

As an additional sensitivity to test the impact that u_{*} has on the HNO₃ dry deposition in the CMAQ model, the MM5-CMAQ simulation was re-run using the u_{*} and aerodynamic resistance values calculated by the WRF model in place of the values calculated by MM5 in the m3dry and aero_depv subroutines in the CMAQ model code. Results from the new MM5-CMAQ simulation showed that HNO₃ concentrations increased on average across the domain, particularly in Texas, the mid-Atlantic, the Northeast, and the Great Lakes regions (Fig. 4d). The average increase in HNO₃ was between 0.10 and 0.40 μg/m³, with the largest increase being 0.60 μg/m³. These increases are smaller than the majority of differences in TNO₃ between the MM5-CMAQ and WRF-CMAQ simulations, which generally range between 0.20 and 0.80 μg/m³. Overall, the higher concentrations of TNO₃ in January appear to be most likely due to the lower wind speeds and PBL heights in the WRF-CMAQ simulation, with the change in HNO₃ dry deposition due to differences in u_{*} being only a secondary contributor to the differences.

4.1.4. Total Carbon (TC)

Fig. 3e shows the predicted monthly average concentrations of TC for the two simulations for January. Differences are generally small and isolated; however, there are several areas where larger differences occur, specifically in the Northeast, along the Gulf of Mexico coast and in southern Florida. Although the differences in TC predictions are not very widespread, they do result in a larger bias at the CSN sites for the WRF-CMAQ simulation, with the NMdnB and MdnB 7.0 % and 0.15 μ g/m³ higher, respectively, than the MM5-CMAQ simulation (Table 2).

The error is also higher at the CSN sites in the WRF-CMAQ simulation. At the IMPROVE network sites, the bias and error are very similar between the two simulations. The larger bias at the CSN sites in the WRF-CMAQ simulation is due mainly to higher predicted TC concentrations in the Northeast, Great Lakes and Mid-Atlantic regions. Some of these differences are not apparent from Fig. 3e, as the average difference in TC between the two simulations is $0.15 \,\mu\text{g/m}^3$, which falls within the gray shading on the figure.

4.1.5. Total Fine Particulate Mass (PM_{2.5})

Fig. 3f shows the monthly average predictions in total $PM_{2.5}$ mass for January for the two CMAQ simulations. The differences in predicted total $PM_{2.5}$ mass between the two simulations are dominated by the differences in SO_4^{2-} , TNO₃ and TC predictions already noted. The MM5-CMAQ simulation has a slight bias in predicted total $PM_{2.5}$ mass (Table 2). The predicted total $PM_{2.5}$ mass is higher in the WRF-CMAQ simulation, which results in an increase in the NMdnB and MdnB of 5.0-8.6 % and 0.21-0.87 µg/m³, respectively.

Regarding the calculation of total PM_{2.5} mass from the raw CMAQ model output, PM_{2.5} concentrations are calculated as a weighted sum of 40 different chemical species tracked within the CMAQv4.7 aerosol module (Equation 5).

$$PM_{2.5} = SO 4_{i,j,k} + NO 3_{i,j,k} + NH4_{i,j,k} + Na_{i,j,k} + Cl_{i,j,k} + EC_{i,j} + 1.2 ORGPA_{i,j} + SOA_{j} + Unspec_{j,k} + Soil_{k}$$
(5)

The subscripts *i*, *j*, and *k* represent the Aitken, accumulation, and coarse modes of the particle size distribution, respectively; Na represents a sum of all sea-salt cations, including sodium, potassium, magnesium, and calcium; ORGPA represents the directly-emitted organic carbon; the multiplicative factor of 1.2 approximates the oxidation of ORGPA that occurs during atmospheric transport, a process that is not represented in CMAQ v4.7; SOA represents the sum of 19

secondary organic species described by Carlton et al. (2009); Unspec_j and Unspec_k are the model species A25J and ACORS, respectively, which represent directly-emitted PM that is not chemically speciated in the national emissions inventory. In Equation (5), each species with a subscript i is multiplied by a factor, PM25AT, to remove the portion of the Aitken mode mass distribution that exceeds 2.5 μ m in aerodynamic diameter. Likewise, all species with subscript j are multiplied by PM25AC and the species with subscript k are multiplied by PM25CO. These three scaling factors have values between 0 and 1, which are computed in each grid cell during each hour of the model simulation following the description by Jiang (2006), and written to the aerosol diagnostic output file.

4.1.6. Wet Deposition Species

Fig. 4 shows the predicted monthly precipitation and SO₄²⁻, NO₃ and NH₄⁺ wet deposition for January for the two CMAQ model simulations. The largest differences in precipitation (Fig. 4a) are generally limited to areas over the Atlantic Ocean and Gulf of Mexico, with smaller differences occurring over the eastern United States. Most of the significant differences in precipitation over land occur in the southern portion of the domain, where the WRF model generally predicts less precipitation than MM5. The bias and error for precipitation (Table 2) is similar for both simulations, with the WRF simulation having slightly lower bias and error than the MM5 simulation.

The SO₄²⁻, NO₃⁻ and NH₄⁺ wet deposition are all lower in the WRF-CMAQ simulation, particularly in the Northeast, where large differences in precipitation were not observed. The SO₄²⁻ wet deposition (Fig. 4b) shows the largest and most widespread decrease, which results in an unbiased NMdnB and MdnB for the WRF-CMAQ simulation, versus a NMdnB of 6.8 % and MdnB of 0.01 kg/ha for the MM5-CMAQ simulation (Table 2). The NO₃⁻ and NH₄⁺ wet

deposition (Figs. 4c and d) show smaller differences in bias between the two simulations. The error is generally comparable for the two simulations, with the WRF-CMAQ simulation having slightly lower error for SO₄²⁻ and NO₃⁻ wet deposition.

4.2. August

4.2.1. Ozone (O₃)

The predicted monthly average O₃ for August for the two CMAQ model simulations is shown in Fig. 5a, while the average observed and predicted diurnal concentrations for the entire domain are shown in Fig. 6. The predicted O₃ mixing ratios in the WRF-CMAQ simulation are higher throughout a large portion of the domain, particularly in the southern and western portions of the domain, while there are only a few isolated areas where O3 mixing ratios were lower in the WRF-CMAQ simulation. The largest differences in predicted O₃ mixing ratios occur along the Gulf of Mexico, where the difference in predicted monthly average O₃ is greater than 4 ppb over a widespread area, with some isolated areas of greater than 10 ppb higher O₃. Both simulations overpredict O₃ (Table 3), however the overprediction is much larger for the hourly O₃ than the maximum 8-hr average O₃ due to large overpredictions of O₃ during the nighttime hours (Fig. 6). As expected, the bias is larger in the WRF-CMAQ simulation, with a NMdnB 4.9 and 4.2 % higher and a MdnB 2.0 and 1.5 ppb higher than the MM5-CMAQ simulation for maximum 1-hr and 8-hr average O₃, respectively. The error is also slightly higher in the WRF-CMAQ simulation. Fig. 7a shows the difference in the mean bias of hourly O₃ (as compared to observations) at the AQS sites between the two simulations. The increase in mean bias is mainly limited to sites along the Gulf Coast, where the mean bias at some sites increases by as much as 16 ppb. For the rest of the domain, the change in mean bias is generally small. However, some slightly larger increases in mean bias are noted in the upper Great Lakes region.

The higher predicted O₃ mixing ratios in the WRF-CMAQ simulation appear to be due to several differences between the MM5 and WRF model predictions. First, an analysis of the predicted cloud fraction (CFRAC) from each simulation showed that the predicted CFRAC from the WRF-CMAQ simulation was on average less than that of the MM5-CMAQ simulation. The smaller CFRAC in the WRF-CMAQ simulation is favorable for greater O₃ production, as CFRAC is used in the calculation of the photolysis rate for O₃, and less CFRAC can result in increased O₃ photolysis. Although the CFRAC in the WRF-CMAQ simulation was on average less than the MM5-CMAQ simulation, it is difficult to quantify the exact impact the difference in CFRAC played in the differences in O₃ mixing ratios between the two simulations.

Second, a comparison of surface solar radiation (SR) at the CASTNet sites showed that while both simulations overpredicted SR, the hourly SR during the daytime (7am to 7pm LST) was on average 20 watts/m² higher in the WRF-CMAQ simulation than in the MM5-CMAQ simulation for August, suggesting less overall cloud cover in the WRF-CMAQ simulation. The greater surface SR results in higher surface temperatures in the WRF-CMAQ simulation, which results in significantly greater concentrations of biogenic Volatile Organic Compounds (VOCs), which are highly sensitive to surface temperature. The largest increase in VOCs in the WRF-CMAQ simulation (not shown) occurs along the Gulf of Mexico and through the upper Midwest, where the increase in the monthly average VOC mixing ratios is typically greater than 20 %, with the concentrations in some areas more than doubling. The areas with large increases (>20 %) in VOC mixing ratios in the WRF-CMAQ simulation correspond to those areas where O₃ concentrations were also much higher than in the MM5-CMAQ simulation.

A third difference between the MM5-CMAQ and WRF-CMAQ simulations that likely plays a role in the difference in the predicted O₃ mixing ratios (and other species as well) is the

differences in the calculation of the u_* in each of the models, which was described previously in section 4.1.3. The differences in the calculation of the u_* result in higher concentrations of NO and $NO_2(NO_x)$ in the WRF-CMAQ simulation, which is generally favorable for greater O_3 production. The combination of increased VOC and NO_x mixing ratios results in O_3 mixing ratios that are considerably higher across a large portion of the domain in the WRF-CMAQ simulation. The increase in O_3 may also be enhanced slightly along the Gulf of Mexico by a narrower and weaker sea-breeze front that was observed in the WRF model simulation, which results in less mixing along the coast. While other differences no doubt exist between the two simulations, these differences were identified as the most important factors contributing to the higher predicted O_3 mixing ratios in the WRF-CMAQ simulation.

As was done with January, an MM5 simulation was performed for August in which in the minimum value for u_{*} in the MM5 code was changed from 1.0 m/s to 0.1 m/s to match the WRF model code for the calculation for u_{*}. The new MM5 simulation showed virtually no difference in the calculated values of u_{*}, which was also the case with the January simulation. The wind speeds in the WRF simulation in August tend to be lower than MM5 during the nighttime hours, indicated in Fig. 1d by the larger negative bias in wind speed, which may contribute to the lower values of u_{*} in the WRF model simulation. To test the impact that u_{*} has on the CMAQ model predictions in August, an MM5-CMAQ simulation was performed for in which the u_{*} and aerodynamic resistance values calculated by the WRF model were used instead of those from MM5. As expected, replacing the MM5 calculated u_{*} values with those from the WRF simulation resulted in higher predicted mixing ratios of O₃, with increases in O₃ generally ranging from 1.0 to 2.0 ppb over a large area (Fig 9a). While differences in the u_{*} values in the MM5 and WRF model simulations may contribute to some of the differences in the CMAQ model predictions, other differences between the meteorological models (e.g. differences in predicted cloud cover affecting photolysis) likely play a larger role.

The predicted monthly average $SO_4^{2^2}$ for the two CMAQ simulations for August is shown in Fig. 5b. There are two well-defined areas with significant differences in the predicted $SO_4^{2^2}$ concentrations; one being the area surrounding the Ohio Valley, where $SO_4^{2^2}$ concentrations are lower in the WRF-CMAQ simulation and the other being the area along the Gulf of Mexico, where $SO_4^{2^2}$ concentrations are higher in the WRF-CMAQ simulation. The result of the differences in $SO_4^{2^2}$ predictions is higher bias and error in the WRF-CMAQ simulation, with the NMdnB 0.1 - 9.3 % higher and the MdnB 0.0 - 0.44 $\mu g/m^3$ higher than the MM5-CMAQ simulation (Table 3). Fig. 7b shows the spatial distribution of the difference in mean bias for $SO_4^{2^2}$ at the IMPROVE network, CSN and CASTNet sites. As expected, the largest increase in bias for the WRF-CMAQ simulation occurs in the Ohio Valley and adjacent regions, while there is a small improvement in the mean bias for sites along the Gulf of Mexico. The relatively dense collection of CASTNet sites in the Ohio Valley region results in the larger increase in bias and error for that network as compared to the CSN and IMPROVE networks (Table 3).

It was speculated that the lower predicted SO₄²⁻ concentrations in the Ohio Valley region in the WRF-CMAQ simulation were due to less aqueous-phase (in-cloud) production of SO₄²⁻, while the increase in SO₄²⁻ concentrations along the Gulf of Mexico were due to an increase in the gas-phase production of SO₄²⁻. To test this hypothesis, the sulfur tracking version of CMAQ, which provides the concentration of SO₄²⁻ from all the various sources (e.g. aqueous-phase, gas-phase, direct emissions, etc.) within the CMAQ model was implemented for August. The results from the sulfur tracking version of CMAQ confirmed that the lower SO₄²⁻ concentrations in the Ohio Valley region were due to less aqueous-phase SO₄²⁻ production, while the increase along the Gulf

of Mexico was due to greater gas-phase SO_4^{2-} production and higher OH concentrations in that region (not shown).

The reduced aqueous-phase production of SO₄²⁻ concentrations in the WRF-CMAQ simulation were due to the CMAQ sub-grid cloud model diagnosing fewer non-precipitating clouds than in the MM5-CMAQ simulation. A comparison of the precipitating and non-precipitating cloud fractions from CMAQ (available in the cloud diagnostic file) showed that the non-precipitating cloud fraction in the WRF-CMAQ model simulation was lower than that of the MM5-CMAQ simulation. Since non-precipitating clouds can be a significant source of SO₄²⁻ production in the atmosphere, it is likely that the lower SO₄²⁻ concentrations in the WRF-CMAQ simulation are due to this decrease in non-precipitating clouds. The increase in SO₄²⁻ along the Gulf of Mexico may be related to an increase in photolysis reactions in that area which results in higher OH concentrations and an increase in the gas-phase production of SO₄²⁻ (which is also indicated by the higher O₃ concentrations in that region).

4.2.3. Fine Particulate Nitrate (NO₃) and Total Nitrate (TNO₃)

Figs. 5c and d show the predicted monthly average NO₃ and TNO₃ for August for the two CMAQ model simulations. NO₃ and TNO₃ are both higher in the WRF-CMAQ simulation, with the largest increases occurring in the region surrounding the Great Lakes and along the Gulf of Mexico. The higher predicted NO₃ and TNO₃ concentrations in the WRF-CMAQ simulation are possibly due to less dry deposition of HNO₃ on average in the WRF-CMAQ simulation (a result of the difference in the calculation of u₂ between the two models). The higher concentrations of predicted NO₃ and TNO₃ in the WRF-CMAQ simulation result in a decrease in the bias in NO₃, which is largely underpredicted in both simulations, while the bias and error in TNO₃ predictions increase substantially compared to the MM5-CMAQ simulation (Table 3).

As with O₃, it was suspected that lower values of u* in the WRF model simulation were resulting in less dry deposition of NO_x and HNO₃, which results in higher concentrations of TNO₃ in the WRF-CMAQ simulation. Results from the MM5-CMAQ simulation in which the u* and aerodynamic resistance from the WRF simulation were used instead of those from MM5 show significantly higher concentrations of TNO₃ in some areas, which supports the hypothesis that differences in u* play at least some role in the differences in TNO₃ predictions in August.

Increases in the TNO₃ in the modified MM5-CMAQ simulation (as compared to the original MM5-CMAQ simulation) generally ranged from between 0.20 to 0.80 μg/m³ (Fig. 9b), which would contribute significantly to the differences in TNO₃ in some areas shown in Fig. 5d. While the differences in dry deposition of NO_x and HNO₃ contribute to differences in predicted TNO₃ concentrations in some areas, other differences in the meteorological predictions are obviously important as well.

4.2.4. Total Carbon (TC)

The largest differences in monthly average TC between the two simulations are generally limited to two regions, one along the Gulf of Mexico and the other in the upper Midwest (Fig. 5e). TC is largely underpredicted in both simulations (Table 3), and that underprediction is slightly less in the WRF-CMAQ simulation, with the NMdnB 5.7 - 7.4 % lower and the MdnB 0.09 - 0.23 $\mu g/m^3$ lower than the MM5-CMAQ simulation. Differences in the predicted TC concentrations between the two simulations are likely related to the same factors that result in the higher O_3 , SO_4^{2-} and TNO_3 concentrations.

4.2.5. Total Fine Particulate Mass (PM_{2.5})

Predictions of total PM_{2.5} mass are on average higher in the WRF-CMAQ simulation for August (Table 3), which results in a small improvement in the bias and error, as PM_{2.5} mass is underpredicted in both simulations. The NMdnB and MdnB decrease by 4.6-7.2% and 0.29-0.86 µg/m³, respectively, while the NMdnE and MdnE decrease by 3.4-4.1% and 0.26-0.41 µg/m³, respectively. The largest increase in PM_{2.5} mass in the WRF-CMAQ simulation occurs along the Gulf of Mexico, where the increases in SO₄²⁻, TNO₃ and TC in that same region result in widespread monthly average differences in total PM_{2.5} mass of more than 1 µg/m³, and in some areas differences exceeding 5 µg/m³ (Fig. 5f). There are some isolated areas in the Ohio Valley and surrounding regions where the PM_{2.5} mass decreases by 1-2 µg/m³ in the WRF-CMAQ simulation. Differences in PM_{2.5} mass are due to the differences in the PM_{2.5} constituent species already discussed, along with differences in the prediction of the unspeciated mass.

4.2.6. Wet Deposition Species

The predicted monthly accumulated precipitation and SO₄²⁻, NO₃ and NH₄⁺ wet deposition for August are shown in Fig. 8. There are widespread differences in the predicted precipitation between the two simulations (Fig. 8a). Much of the difference appears to be due to differences in the prediction of the convective precipitation from the two models. The WRF model tends to forecast more precipitation over the southeastern portion of the domain, including over the Atlantic Ocean, Gulf of Mexico and Florida and along the Gulf Coast states, while the MM5 model predicts greater precipitation over the Midwest. Both models overpredict precipitation on average, with the WRF model simulation having a smaller NMdnB and MdnB than the MM5 simulation (Table 3).

Differences in SO₄²⁻ wet deposition are widespread and mixed throughout much of the domain (Fig. 8b). Greater SO₄²⁻ wet deposition occurs over the Southeast and along the Gulf of Mexico

in the WRF-CMAQ simulation, which correlates to regions where greater precipitation was observed as well, while there are areas in the Midwest with less SO₄²⁻ wet deposition in the WRF-CMAQ simulation, which correlate to areas where less precipitation was also predicted. There are, however, also large differences in SO₄²⁻ wet deposition in the Northeast, a region where large differences in precipitation were not observed. It is not immediately apparent what the cause of these differences is, and requires further investigation. Overall, the performance for SO₄²⁻ wet deposition at the NADP network sites is slightly better for the WRF-CMAQ simulation, with slightly less bias and error as compared to the MM5-CMAQ simulation (Table 3). The NO₃⁻ and NH₄⁺ wet deposition (Figs. 8c and d) show similar patterns to the SO₄²⁻ wet deposition, with higher deposition of those species in the Southeast and along the Gulf of Mexico in the WRF-CMAQ simulation. These increases in wet deposition are likely due to the combination of increases in aerosol concentrations of those species as well as greater predicted precipitation in those regions in the WRF model simulation. The overall statistical performance for NO₃⁻ and NH₄⁺ wet deposition is slightly better for the WRF-CMAQ simulation (Table 3).

5. Summary

Two sets of CMAQv4.7 simulations were performed for January and August 2006, with one set using the MM5 meteorology and the other set using WRF model meteorology. Predictions from the CMAQ model simulations were compared against observations from various networks and the performance for each set of simulations was assessed and compared against the other set. For January, performance differences in the predicted O₃ mixing ratios from each simulation appear to be the result of differences in the calculation of the vegetation fraction between the two simulations, which ultimately affects the amount of O₃ dry deposition that takes place in each simulation. Higher predicted concentrations of SO₄²⁻ in January in the WRF-CMAQ simulation are likely related to a combination of more predicted cloud cover, which results in an increase in

the amount of aqueous-phase (in-cloud) SO₄²⁻ produced, and less SO₄²⁻ wet deposition as compared to the MM5-CMAQ simulation. Predictions of NO₃ and TNO₃ were also higher in the WRF-CMAQ simulation, likely due to lower predicted wind speeds and PBL heights in the WRF simulation, which result in less mixing and hence greater surface concentrations. A likely secondary contributor to the differences in TNO₃ concentrations in the January are differences in the calculated u* values in the MM5 and WRF model simulations, where smaller u* values calculated in the WRF model simulations result in less dry deposition of HNO₃ (and hence greater ambient concentrations) in the WRF-CMAQ simulation.

For August, the WRF-CMAQ simulation generally underperformed compared to the MM5-CMAQ simulation. The bias in O₃ mixing ratios was higher in the WRF-CMAQ simulation, with the largest increases in bias occurring in the southeast United States, particularly in Florida, along the Gulf of Mexico and in Texas. The increase in predicted O₃ mixing ratios in the WRF-CMAQ simulation appears to be most directly related to greater predicted surface SR (due to fewer predicted clouds) in the WRF-CMAQ simulation, which results in higher surface temperatures and an increase in the mixing ratios of surface biogenic VOCs. Additionally, the smaller predicted CFRAC in the WRF-CMAQ simulation results in an increase in the amount of O₃ photolysis taking place. Less dry deposition of NO_x due to lower u* values also appears to contribute to the higher O₃ mixing ratios in the WRF-CMAQ simulation.

Predicted concentrations of SO₄²⁻, which were already underpredicted in both simulations, were lower in the WRF-CMAQ simulation in the Ohio Valley region, but higher along the Gulf coast states. The decrease in predicted SO₄²⁻ concentrations in the WRF-CMAQ simulation is likely related to fewer predicted non-precipitating clouds in the WRF-CMAQ simulation, which results in less aqueous-phase production of SO₄²⁻ in the Midwest and Ohio Valley, while the increase along the Gulf of Mexico is due to greater gas-phase production of SO₄²⁻. Predicted

concentrations of NO₃ and TNO₃ were higher in the WRF-CMAQ simulation, which is thought to be a result of increased concentrations of NO_x and HNO₃, due in part to less dry deposition of those species in some areas (due to differences in the u- calculation) and to an increase in NO₃ replacement in response to lower predicted SO₄² concentrations. Other differences in the meteorological predictions, such as cloud cover, PBL heights and handling of the land-sea interface along the Gulf of Mexico, likely play a large role in the differences in NO₃ and TNO₃ as well.

The most significant differences in the meteorological predictions are related to the predictions of wind speed, PBL height, u_{*}, water vapor and the predicted cloud cover, all of which appear to contribute to differences in the CMAQ model predictions. However, it should be noted that the comparison presented here is limited to two months from a single year. Additional comparisons during other time periods would be useful for quantifying the robustness of the results presented here. This could be accomplished by extending the comparison to an annual or multi-annual simulation, which would capture differences under many different meteorological regimes. It should also be noted that while the comparison presented here uses configurations of the meteorological models that are typical of those used for air quality applications, many different configurations of the MM5 and WRF models are possible, and comparisons of CMAQ model predictions using different configurations of the meteorological models may lead to results that are different than those presented here. Finally, the results presented here are also limited geographically to the eastern United States. An analysis of the performance for the western United States would be beneficial, since the meteorology and air quality conditions in the western United States can be quite different from those of the eastern United States.

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TABLES

Table 1. Options used for the MM5 and WRF model simulations.

Model	MM5	WRF
Version	v3.7.4	ARW core v3.0
Grid Spacing	12km X 12km	12km X 12km
PBL Model	ACM2	ACM2
LSM	Pleim-Xiu	Pleim-Xiu
Sub-grid Convection Scheme	Kain-Fritsch 2	Kain-Fritsch 2
Shortwave Radiation Scheme	Dudhia	Dudhia
Longwave Radiation Scheme	RRTM	RRTM
Explicit Microphysics Scheme	Reisner 2	Thompson
Objective Analysis Approach	Rawins	OBSGRD

Table 2. Statistics of RMSE, NMdnB, NMdnE, MdnB and MdnE for fine particulate and wet deposition species for January 2006. MM5 indicates the MM5-CMAQ simulation; WRF indicates the WRF-CMAQ simulation. MdnB and MdnE values are in ppb for O_3 , $\mu g/m^3$ for aerosol species, mm for precipitation and kg/ha for wet deposition species.

Species	Notresale	# of Obs	NMdnB (%)		NMdnE (%)		MdnB		MdnE	
	Network		MM5	WRF	MM5	WRF	MM5	WRF	MM5	WRF
O ₃ (Hourly) O ₃ (8-h Max)	AQS	245129	8.2	11.9	25.4	26.5	1.88	2.73	5.85	6.10
	AQS	9925	1.5	5.2	12.9	12.8	0.5	1.69	4.19	4.16
SO ₄ ²⁻	IMPROVE	787	-6.7	-1.8	22.8	19.4	-0.08	-0.03	0.27	0.23
	CSN	1034	-13.2	-5.8	24.8	23.7	-0.29	-0.13	0.55	0.52
	CASTNet	247	-12.8	-2.8	17.3	12.4	-0.26	-0.06	0.35	0.25
	IMPROVE	787	-9.2	-8.2	73.9	71.1	-0.04	-0.03	0.29	0.28
	CSN	994	-17.2	-7.0	48.0	46.0	-0.24	-0.10	0.68	0.65
TNO ₃	CASTNet	247	8.3	17.4	19.1	21.6	0.19	0.40	0.43	0.49
TC IN	IMPROVE	820	-16.8	-15.4	41.7	41.7	-0.14	-0.13	0.36	0.36
	CSN	941	7.8	14.9	38.8	43.9	0.16	0.30	0.77	0.87
PM _{2.5}	IMPROVE	859	4.6	9.4	36.0	39.4	0.19	0.39	1.51	1.65
	CSN	883	1.6	10.2	28.6	31.1	0.16	1.02	2.86	3.11
Precipitation		711	6.0	3.2	45.0	40.4	0.53	0.28	4.00	3.59
WetD Sulf.	NADP	576	6.2	-0.1	49.2	44.6	0.01	0.00	0.06	0.06
WetD Amm.		576	-17.4	16.9	49.8	49.8	0.00	0.00	0.01	0.01
WetD Nitr.		576	2.3	-2.9	47.8	45.8	0.00	0.00	0.06	0.06

Table 3. Statistics of RMSE, NMdnB, NMdnE, MdnB and MdnE for fine particulate and wet deposition species for August 2006. MM5 indicates the MM5-CMAQ simulation; WRF indicates the WRF-CMAQ simulation. MdnB and MdnE values are in ppb for O_3 , $\mu g/m^3$ for aerosol species, mm for precipitation and kg/ha for wet deposition species.

Species Network	N-4 1	# of	NMdnB (%)		NMdnE (%)		MdnB		MdnE	
	Obs	MM5	WRF	MM5	WRF	MM5	WRF	MM5	WRF	
O ₃ (Hourly) O ₃ (8-h Max)	AQS	598583	14.1	19.0	28.9	31.1	4.36	5.88	8.95	9.65
		24413	1.2	5.4	13.3	14.2	0.57	2.62	6.47	6.89
SO ₄ ²⁻	IMPROVE	531	-8.5	-8.6	38.5	34.5	-0.12	-0.12	0.53	0.48
	CSN	932	-6.7	-8.0	25.0	23.1	-0.24	-0.28	0.89	0.82
	CASTNet	251	-11.8	-21.1	14.9	21.6	-0.57	-1.01	0.72	1.06
NO ₃	IMPROVE	531	-51.7	-44.9	73.0	70.8	-0.07	-0.06	0.10	0.10
	CSN	892	-45.2	-30.1	63.0	63.4	-0.18	-0.12	0.25	0.25
TNO ₃	CASTNet	251	10.9	28.4	32.0	40.1	0.18	0.46	0.52	0.65
TC	IMPROVE	701	-47.7	-42.0	53.5	47.3	-0.71	-0.62	0.80	0.70
	CSN	896	-44.7	-37.3	46.9	41.5	-1.40	-1.17	1.47	1.30
PM _{2.5} IMPROVE CSN	IMPROVE	693	-32.7	-28.1	38.2	34.1	-2.10	-1.81	2.45	2.19
	CSN	809	-22.1	-14.9	30.9	27.5	-2.65	-1.79	3.71	3.30
Precipitation	NADP	709	18.5	7.3	94.6	83.7	2.59	1.03	13.2	11.7
WetD Sulf.		634	4.4	2.3	70.2	64.6	0.01	0.00	0.16	0.14
WetD Amm.		634	-5.7	-1.2	70.0	68.1	0.00	0.00	0.03	0.03
WetD Nitr.		634	-44.8	-38.5	57.0	55.4	-0.09	-0.07	0.11	0.11

FIGURE CAPTIONS

Fig. 1. a) Daily RMSE and b) hourly bias of 2-m T (red; K), w (green; g/kg) and 10-m WS (blue; ms⁻¹) for the MM5 (dashed) and WRF (solid) model simulations for January 2006. c) As in (a), except for August 2006. d) As in (b), except for August 2006.

Fig 2. January 2006 monthly accumulated precipitation (cm) for a) NPA observed, b)

MM5 predicted and c) WRF predicted. August 2006 monthly accumulated precipitation
(cm) for d) NPA observed, e) MM5 predicted and f) WRF predicted.

Fig. 3. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) O_3 (ppb) b) SO_4^{2-} ($\mu g/m^3$) c) NO_3^{-} ($\mu g/m^3$) d) TNO_3 ($\mu g/m^3$) e) TC ($\mu g/m^3$) and f) total $PM_{2.5}$ mass ($\mu g/m^3$) for January 2006.

Fig. 4. Difference in monthly average (WRF –MM5) a) u* (m/s), b) layer one wind speed (m/s) and c) surface roughness length (m) for January 2006. (d) Difference in monthly average TNO₃ (μg/m³) between the MM5-CMAQ simulation using u* values calculated by WRF and the original MM5-CMAQ simulation for January 2006.

Fig. 5. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) precipitation (cm) b) SO₄²⁻ wet deposition (kg/ha) c) NO₃ wet deposition (kg/ha) and d) NH₄⁺ wet deposition (kg/ha) for January 2006.

Fig. 6. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) O_3 (ppb) b) SO_4^{2-} ($\mu g/m^3$) c) NO_3^- ($\mu g/m^3$) d) TNO_3 ($\mu g/m^3$) e) TC ($\mu g/m^3$) and f) total $PM_{2.5}$ mass ($\mu g/m^3$) for January 2006.

Fig. 7. Diurnal domain-wide average O₃ for August 2006 for AQS observed (black solid crosses; light gray shading), MM5-CMAQ predicted (dashed blue triangles; medium gray shading) and WRF-CMAQ predicted (dashed red plus signs; dark gray shading). The solid and dashed lines represent the average hourly O₃ concentration, while the shading represents the 25th to 75th percentiles.

Fig. 8. Average difference in the mean bias between the MM5-CMAQ and WRF-CMAQ simulations for a) maximum 8-hr average O_3 (ppb) at the AQS sites and b) SO_4^{2-} ($\mu g/m^3$) at IMPROVE (circle), CSN (triangle) and CASTNet (square) for August 2006. Warmer shading represents higher bias in the WRF-CMAQ simulation; cooler shading represents lower bias in the WRF-CMAQ simulation; gray shading represents a difference in mean bias of less than 2 ppb or $0.2~\mu g/m^3$ between the two simulations.

Fig. 9. Difference in monthly average a) O_3 mixing ratios (ppb) and b) TNO_3 ($\mu g/m^3$) between the MM5 simulation using u_* values calculated by WRF and the original MM5 simulation.

Fig. 10. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) precipitation (cm) b) SO₄²⁻ wet deposition (kg/ha) c) NO₃⁻ wet deposition (kg/ha) and d) NH₄⁺ wet deposition (kg/ha) for August 2006.

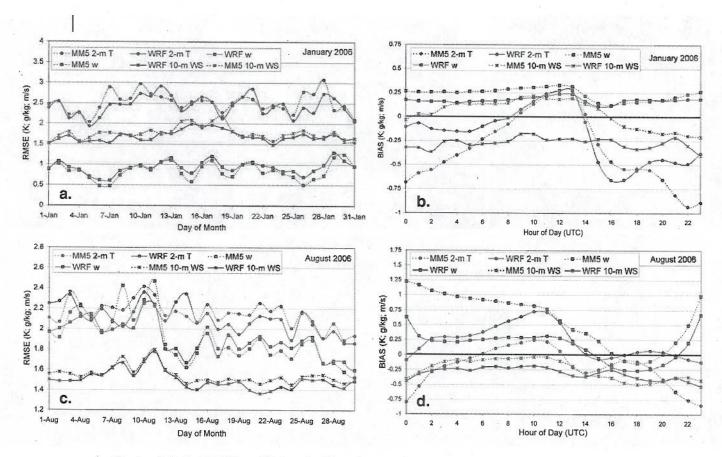


Fig. 1. a) Daily RMSE and b) hourly bias of 2-m T (red; K), w (green; g/kg) and 10-m WS (blue; ms⁻¹) for the MM5 (dashed) and WRF (solid) model simulations for January 2006. c) As in (a), except for August 2006. d) As in (b), except for August 2006.

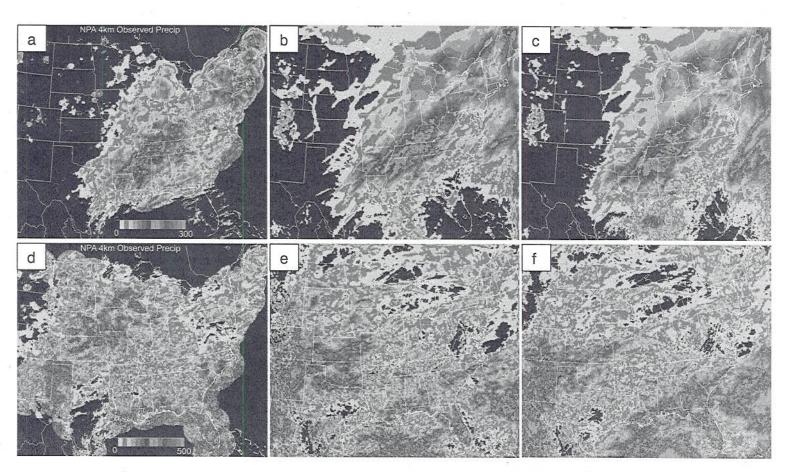


Fig 2. January 2006 monthly accumulated precipitation (cm) for a) NPA observed, b) MM5 predicted and c) WRF predicted. August 2006 monthly accumulated precipitation (cm) for d) NPA observed, e) MM5 predicted and f) WRF predicted.

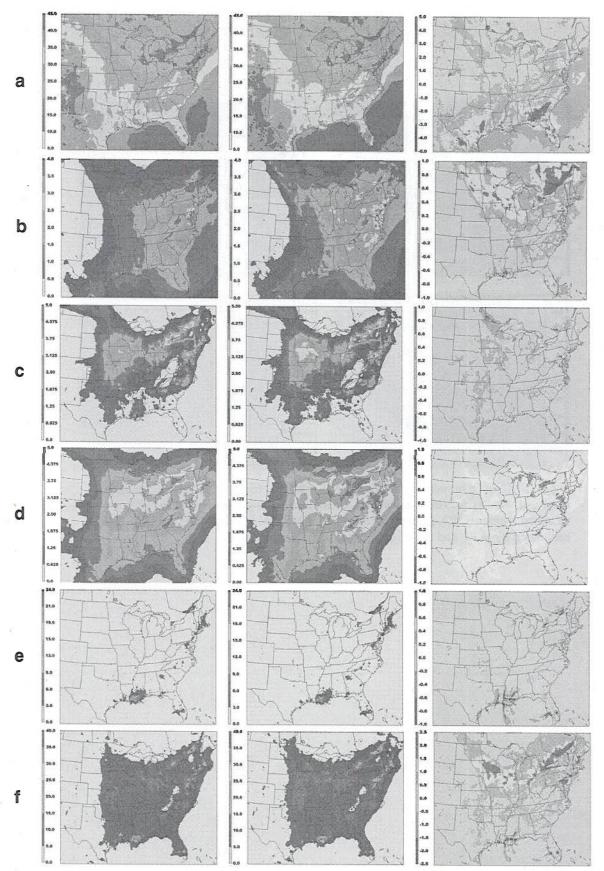


Fig. 3. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) O_3 (ppb) b) SO_4^{2-} ($\mu g/m^3$) c) NO_3^- ($\mu g/m^3$) d) TNO_3 ($\mu g/m^3$) e) TC ($\mu g/m^3$) and f) total $PM_{2.5}$ mass ($\mu g/m^3$) for January 2006.

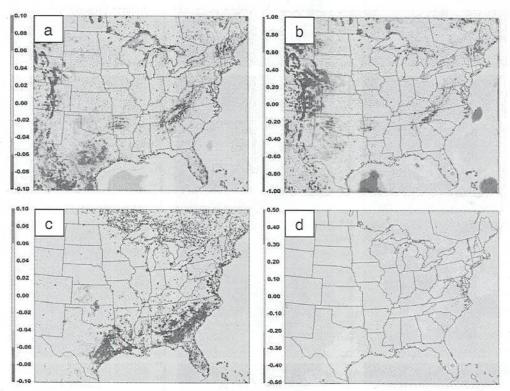


Fig. 4. Difference in monthly average WRF –MM5 a) friction velocity (u*; m/s), b) layer one wind speed (m/s) and c) surface roughness length (m). d) Difference in monthly average TNO $_3$ (µg/m 3) between the MM5-CMAQ simulation using u* values calculated by WRF and the original MM5-CMAQ simulation.

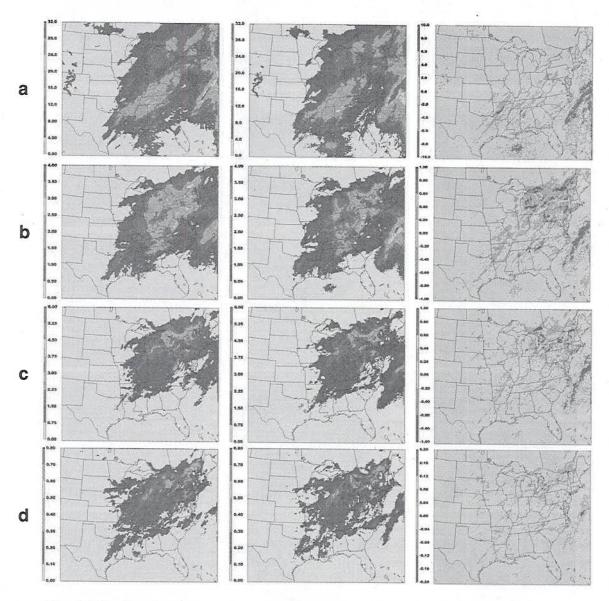


Fig. 5. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) precipitation (cm) b) SO_4^{2-} wet deposition (kg/ha) c) NO_3^{-} wet deposition (kg/ha) and d) NH_4^+ wet deposition (kg/ha) for January 2006.

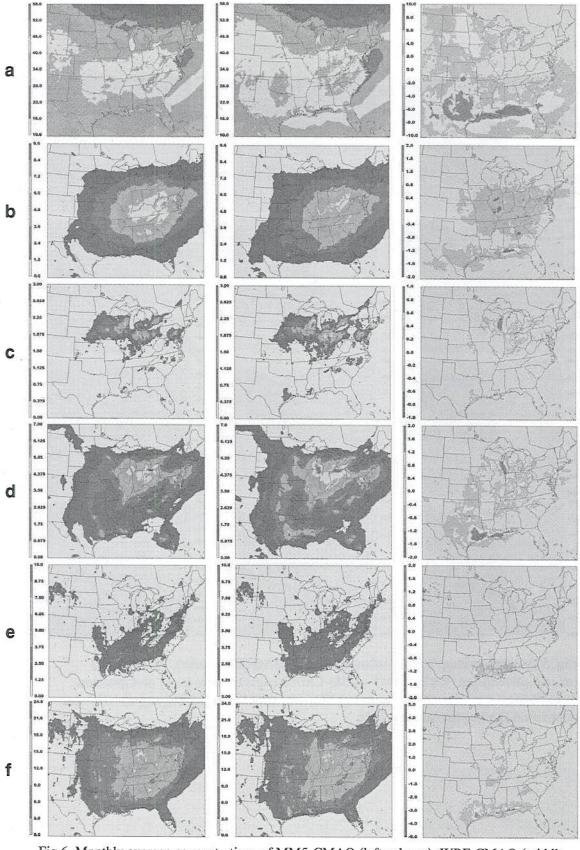


Fig 6. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) O_3 (ppb) b) SO_4^{2-} ($\mu g/m^3$) c) NO_3^{-} ($\mu g/m^3$) d) TNO_3 ($\mu g/m^3$) e) TC ($\mu g/m^3$) and f) total $PM_{2.5}$ mass ($\mu g/m^3$) for August 2006.

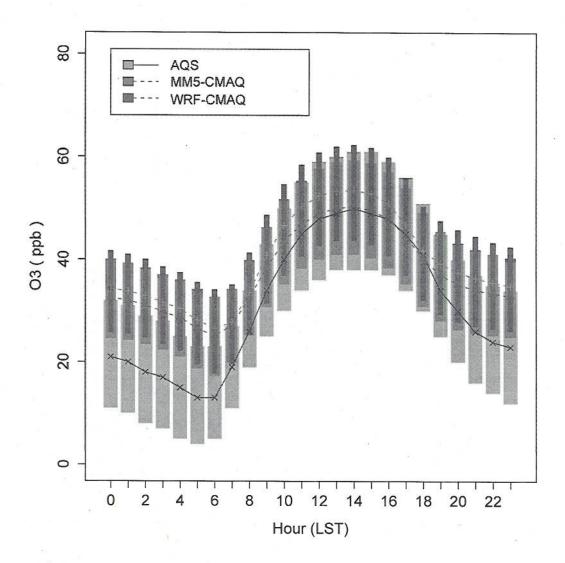


Fig. 7. Diurnal domain-wide average O_3 for August 2006 for AQS observed (black solid crosses; light gray shading), MM5-CMAQ predicted (dashed blue triangles; medium gray shading) and WRF-CMAQ predicted (dashed red plus signs; dark gray shading). The solid and dashed lines represent the average hourly O_3 concentration, while the shading represents the 25^{th} to 75^{th} percentiles.

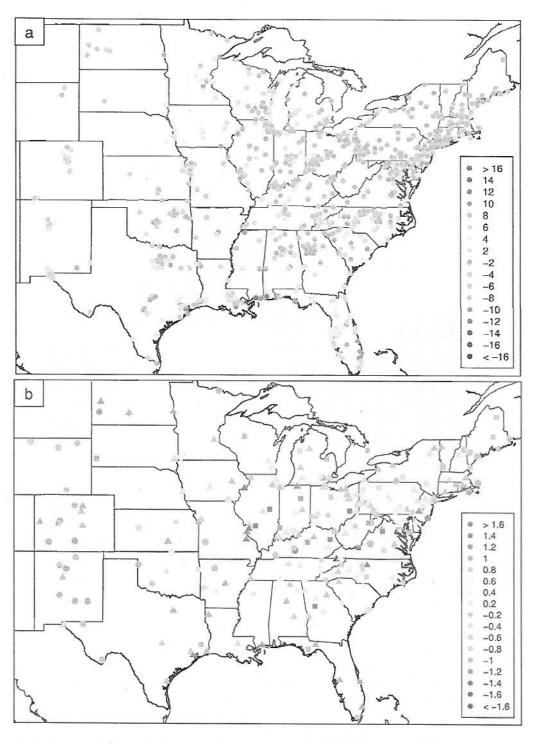


Fig. 8. Average difference in the mean bias between the MM5-CMAQ and WRF-CMAQ simulations for a) maximum 8-hr average O_3 (ppb) at the AQS sites and b) $SO_4^{2^-}$ ($\mu g/m^3$) at IMPROVE (circle), CSN (triangle) and CASTNet (square) for August 2006. Warmer shading represents higher bias in the WRF-CMAQ simulation; cooler shading represents lower bias in the WRF-CMAQ simulation; gray shading represents a difference in mean bias of less than 2 ppb or $0.2~\mu g/m^3$ between the two simulations..

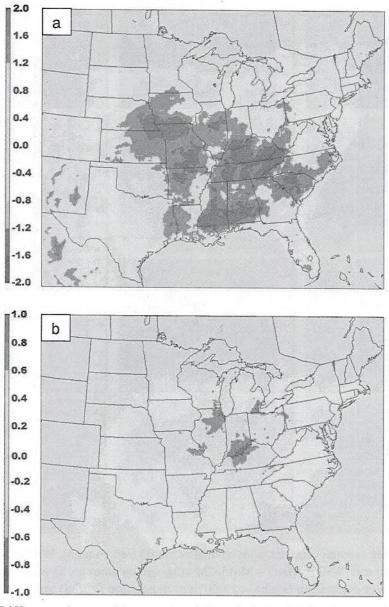


Fig. 9. Difference in monthly average a) O_3 mixing ratio (ppb) and b) TNO₃ ($\mu g/m^3$) between the MM5-CMAQ simulation using u_* values calculated by WRF and the original MM5-CMAQ simulation.

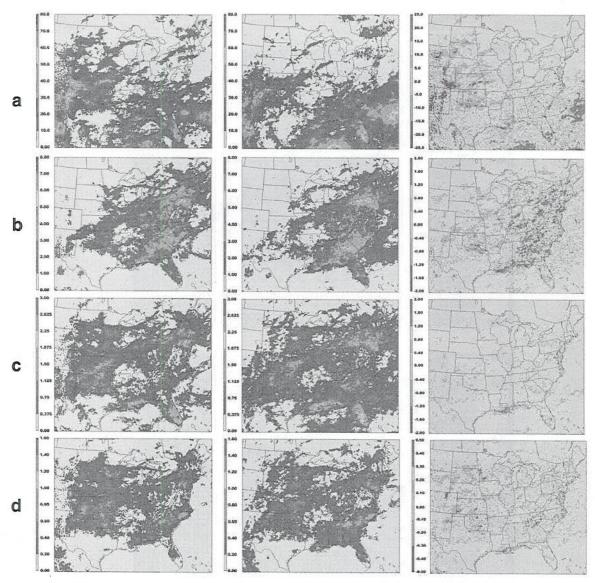


Fig. 10. Monthly average concentrations of MM5-CMAQ (left column), WRF-CMAQ (middle column) and WRF-CMAQ – MM5-CMAQ (right column) for a) precipitation (cm) b) SO_4^{2-} wet deposition (kg/ha) c) NO_3^{-} wet deposition (kg/ha) and d) NH_4^+ wet deposition (kg/ha) for August 2006.