

1 **Title:** Quantitative PCR for genetic markers of human fecal pollution

2

3

4 **Authors:** Orin C. Shanks^{1*}, Catherine A. Kelty¹, Mano Sivaganesan¹, Manju
5 Varma², and Richard A. Haugland²

6

7

8 ¹ U.S. Environmental Protection Agency, Office of Research and Development, National
9 Risk Management Research Laboratory, 26 West Martin Luther King Drive, Cincinnati,
10 OH 45268, USA

11

12 ² U.S. Environmental Protection Agency, Office of Research and Development, National
13 Exposure Research Laboratory, 26 West Martin Luther King Drive, Cincinnati, OH.
14 45268, USA

15

16

17 **Running Title:** qPCR FOR HUMAN FECAL BACTERIA

18

19

20 *Corresponding Author: Tel.: +1 513 569 7314; Fax: +1 513 569 7328;

21 Email address: shanks.orin@epa.gov

22

23

24 **Keywords:** Quantitative PCR, risk assessment, microbial source tracking, water
25 quality

26

ABSTRACT

27 Assessment of health risk and fecal bacteria loads associated with human fecal
28 pollution requires reliable host-specific analytical methods and a rapid quantification
29 approach. We report the development of quantitative PCR assays for quantification of
30 two recently described human-specific genetic markers targeting *Bacteroidales*-like cell
31 surface associated genes. Both assays exhibited a range of quantification from 10 to
32 1×10^6 copies of target DNA. For each assay, internal amplification controls were
33 developed to detect the presence or absence of amplification inhibitors. The assays
34 predominantly detected human fecal specimens and exhibited specificity levels greater
35 than 97% when tested against 265 fecal DNA extracts from 22 different animal species.
36 The abundance of each human-specific genetic marker was measured in primary effluent
37 wastewater samples collected from 20 geographically distinct locations and compared to
38 quantities estimated by real-time PCR assays specific for ribosomal RNA gene sequences
39 from total *Bacteroidales* and enterococci fecal microorganisms. Assay performances
40 combined with the prevalence of DNA targets in sewage samples provide experimental
41 evidence supporting the potential application of these quantitative methods for
42 monitoring fecal pollution in ambient environmental waters.

43

44

INTRODUCTION

45

46 Waterborne diseases that originate from human fecal pollution remain a
47 significant public health issue. As a result, a large number of methods have been
48 developed to detect and quantify human fecal pollution (10, 12, 18, 21). The majority of

49 these methods are based on real-time quantitative PCR (qPCR) assays designed to
50 estimate the concentration of 16S ribosomal RNA gene (rRNA) sequences from various
51 subpopulations within the order *Bacteroidales*. This bacterial order constitutes a large
52 proportion of the normal gut microbiota of most animals, including humans (3, 15, 27).
53 Bacterial 16S rRNA genes are useful as markers because they have relatively low
54 mutation rates (7) and are typically present in multiple operons, increasing template DNA
55 levels available for detection (2, 11, 17, 29). While several studies have demonstrated the
56 value of *Bacteroides* 16S rDNA-based qPCR assays, currently available assays cannot
57 discriminate between several animal sources closely associated with humans, including
58 cats, dogs, and/or swine (10, 12, 18, 21). Alternative qPCR assays targeting genes
59 directly involved in host-specific interactions may be capable of increased discrimination
60 of fecal pollution sources (23, 24) and are needed to complement existing qPCR-based
61 approaches used to identify sources of human fecal pollution.

62 A recent metagenomic survey of a human fecal bacterial community using
63 genome fragment enrichment has led to the identification of hundreds of candidate
64 human fecal bacteria-specific DNA sequences (24). PCR assays targeting two gene
65 sequences encoding for a hypothetical protein potentially involved in remodeling of
66 bacterial surface polysaccharides and lipopolysaccharides (Assay 19) and a putative RNA
67 polymerase extracytoplasmic function sigma factor (Assay 22) from *Bacteroidales*-like
68 microorganisms exhibited a high level of specificity (100%) for human fecal material
69 (24).. However, it remained to be determined whether these reported chromosomal DNA
70 sequences are abundant and uniform enough within human populations to be detected
71 once diluted in the environment. Based on these considerations, the next steps toward the

72 application of these gene sequences for water quality monitoring applications were to
73 design qPCR assays for their detection and then to use these assays to evaluate the overall
74 abundance and distribution of these sequences in human populations relative to those of
75 ribosomal RNA gene sequences from different currently recognized fecal indicator
76 bacteria groups.

77 Here we report the development of two qPCR assays for the quantification of the
78 human-specific DNA sequences targeted by previously reported PCR assays 19 and 22
79 (24). Method performance characteristics including specificity, range of quantification
80 (ROQ), limit of quantification, amplification efficiency, and analytical precision were
81 defined for each assay. An internal amplification control (IAC) was designed to monitor
82 for the presence of inhibitors commonly associated with environmental sampling that can
83 confound DNA target copy number estimations. Finally, the abundance of each DNA
84 target was measured by qPCR analysis in primary effluent wastewater samples
85 representative of 20 geographically distinct human populations. In addition, the
86 abundance of these human-specific DNA genes in wastewater was compared to those
87 rRNA genes from *Bacteroidales* and enterococci, two general fecal indicator bacterial
88 groups that have been widely used for water quality testing.

89

90

MATERIALS AND METHODS

91

92 **Sample collection.** Individual fecal samples (n = 265) and wastewater samples (n = 20)
93 were collected for analysis as previously described (24). Primary effluent wastewater
94 samples were collected on-site from 20 different wastewater treatment facilities across

95 the United States (Tab. 1). Facilities were selected based on population served and
96 geographic location. Briefly, 500 ml of primary effluent was collected from each facility
97 and immediately stored on ice. Samples were then packed in ice and shipped overnight
98 to Cincinnati, OH for laboratory testing. Twenty-five milliliters of primary effluent from
99 each facility was filtered through a 0.2 µm pore size Supor-200 filters (Whatman) and
100 each filter was placed in a sterile 1.5 ml microtube and stored at -80°C (< 6 months) until
101 time of DNA extraction and qPCR amplification.

102 Individual fecal samples were collected over a 12-month period at various
103 locations across the United States from 22 different animal species likely to impact
104 watersheds or beaches including *Homo sapiens* (human, n = 16), *Lama pacos* (alpaca; n =
105 2), *Anser* sp. (Canadian goose; n = 12), *Felis catus* (cat, n = 10), *Gallus gallus* (chicken, n
106 = 10), *Bos taurus* (cow, n = 80), *Odocoileus virginianus* (white-tail deer, n = 15),
107 *Odocoileus hemionus* (mule deer, n = 5), *Cervus elaphus* (elk, n = 5), *Alces alces* (moose,
108 n = 1), *Antilocapra americana* (pronghorn, n = 4), *Canis familiaris* (dog, n = 10), *Anas* sp.
109 (duck, n = 12), *Capra aegagrus* (goat, n = 7), *Equus caballus* (horse, n = 12), *Pelecanus*
110 sp. (pelican, n = 5), *Sus scrofa* (pig, n = 22), Laridae (gull, n = 12), *Ovis aries*, (sheep, n =
111 10), *Zalophus californianus* (sea lion, n = 5), Delphinidae (marine dolphin, n = 3), and
112 *Meleagris* sp. (turkey, n = 7). Each fecal sample was collected from a different
113 individual to maximize the opportunity to observe false positive amplifications.

114

115 **DNA extraction of fecal and primary effluent wastewater samples.** All DNA
116 extractions were performed with the FastDNA kit for soils (Q-Biogene, Carlsbad, CA) as
117 described (24) with the exception that a FastPrep®-24 instrument (MP, Solon, OH) at

118 setting of 6 m/s for 120 sec was used for cell lysis. DNA extraction yields were
119 determined with a NanoDrop ND-1000 UV spectrophotometer (NanoDrop Technologies;
120 Wilmington, DE). Filtration and extraction controls, with purified water substituted for
121 primary effluent, were performed each day samples were received or extracted to monitor
122 for potential contamination.

123

124 **Oligonucleotides and primers.** TaqMan® probe and primer assays targeting the rRNA
125 genes of *Bacteroidales* (GenBac3) and *Enterococcus* (Entero1) are reported elsewhere (8,
126 13). qPCR probe and primer sequences for the putative human-specific HumM2 and
127 HumM3 assays (Tab. 2) were designed with Primer Express software (Applied
128 Biosystems, Foster City, CA) based on the previously reported end-point PCR assays
129 HumM19 and HumM22, respectively (24). Primers and TaqMan® probes were designed
130 using the default parameters of the Primer Express software (Version 1.5; Applied
131 Biosystems). Fluorogenic probes were 5' labeled with 6FAM (6-carboxyfluorescein) or
132 VIC™ and 3' labeled with TAMRA (6-carboxytetramethylrhodamine). Optimal primer
133 and probe reaction concentrations were determined according to a standard Applied
134 Biosystems protocol (1). The HumM2 and HumM3 assay primer and probe sets (Tab. 2)
135 were tested for specificity with animal fecal and wastewater sample composites (5 ng
136 DNA template per PCR assay).

137

138 **DNA preparations from pure bacterial cultures.** American Type Culture Collection
139 (ATCC) bacterial strains were used to prepare DNA standards for the *Bacteroidales* and
140 *Enterococcus* qPCR assays. *E. faecalis* (ATCC #29212) was cultured as previously

141 described (8). *B. thetaiotaomicron* (ATCC # 29741) cells were grown in chopped meat
142 carbohydrate broth (Remel, Lenexa, KS) according to manufacturer's instructions. Both
143 cultures were harvested by centrifugation at 8,000 \times g for 5 min, washed twice using
144 sterile phosphate buffered saline (Sigma, St. Louis, MO) and stored in aliquots at -40°C.
145 Cell concentrations of each organism in the final washed suspensions were determined by
146 bright field microscopy at 40x magnification in disposable hemocytometer chambers
147 (Nexcelom Bioscience, #CP2-002, Lawrence, MA). DNA was isolated from the cell
148 suspensions using a bead beating extraction approach (8) and incubated for one hour at
149 37°C with 0.017 μ g/ μ l RNase A (Gentra Systems, USA). DNA purification was
150 performed using a silica column adsorption kit (DNA-EZ, GeneRite, Kendall Park, NJ.).
151 DNA concentrations of cell extracts were determined by spectrophotometric absorbance
152 readings at 260 nm (A_{260}) and purity of the DNA preparations was determined by
153 A_{260}/A_{280} ratios.

154

155 **Construction of IAC and plasmid DNA standards.** A plasmid DNA construct was
156 developed to function as an IAC DNA target that can be spiked into DNA extracts to
157 monitor for PCR inhibition and also as a plasmid DNA standard for calculation of
158 HumM2 and HumM3 qPCR assay calibration curves. The IAC construct was designed to
159 contain a single site for hybridization of a unique TaqMan® VIC™ labeled probe
160 sequence flanked by multiple primer binding sequences (Tab. 2, Fig. 1). To build the
161 human assay IAC construct, long oligonucleotides (> 100 bp, Tab. 1) containing multiple
162 primer sequences (24) were designed such that their 3' ends overlapped. The overlapping
163 fragments were then combined into a single DNA molecule using overlap extension PCR

164 (9). The IAC construct was inserted into a plasmid vector, purified, linearized,
165 quantified, and diluted to generate samples ranging from approximately 10 to 1×10^6
166 molecules of template DNA as described (22).
167
168 **qPCR assays and quantification.** The four qPCR assays used in this study were
169 HumM2, HumM3, GenBac3, and Entero1 (Tab. 2). Amplification was performed in a
170 7900 HT Fast Real Time Sequence Detector (Applied Biosystems). Reaction conditions
171 and thermal cycling parameters for GenBac3 and Entero1 are described elsewhere (20).
172 For HumM2 and HumM3, reaction mixtures (25 μ l) contained 1X TaqMan® Universal
173 PCR Master Mix with AmpErase® uracil-N-glycosylase (UNG, Applied Biosystems),
174 0.2 mg/ml bovine serum albumin (Sigma), 1 μ M of each primer, 80 nM FAM™ or
175 VIC™ labeled TaqMan® probe (Applied Biosystems), and either 1 to 100 ng genomic
176 DNA (fecal and wastewater samples) or 10 to 1×10^6 target gene copies (human IAC
177 plasmid DNA). Reaction mixtures for multiplex applications were the same as above
178 with the additions of both 80 nM of VIC™ or TET labeled TaqMan® probes for IAC
179 plasmid DNA and 80 nM of 6FAM™ labeled TaqMan® probe for native DNA targets.
180 IAC spike concentrations were either 25 or 50 copies. All reactions were performed in
181 triplicate in MicroAmp Optical 96-well reaction plates with MicroAmp Optical Caps
182 (Applied Biosystems). Thermal conditions were 50°C for 2 min to activate UNG
183 followed by 10 min incubation at 95°C to activate AmpliTaq Gold enzyme, the
184 temperature profile then followed a forty cycle pattern with a short denaturation at 95°C
185 for 15 sec and a combined annealing and primer extension phase at 60°C for 1 min. Data
186 was initially analyzed with Sequence Detector Software (Version 2.2.2) at a threshold

187 determination of 0.08 for human-specific assays (HumM2 and HumM3) and 0.03 for
188 general fecal indicator bacteria assays (GenBac3 and Entero1). Threshold cycle (C_T)
189 values were exported to Microsoft Excel for further statistical analysis. A minimum of
190 three no-template amplifications with purified water substituted for template DNA were
191 performed for each 96-well qPCR experiment to monitor for potential contamination.

192

193 **Calculations and statistical analysis.** The specificity of HumM2 and HumM3 was
194 determined as: $\text{specificity} = d/(b+d)$, where 'b' is false positives and 'd' is true negatives.
195 Master calibration curves, unknown DNA concentration estimates, and credible intervals
196 were determined using a Markov Chain Monte Carlo approach (25). Bayesian
197 calculations were performed using the publicly available software WinBUGS version
198 1.4.1 (<http://www.mrc-bsu.cam.ac.uk/bugs>) (14). See supplemental material for
199 WinBUGS program code and resulting data output used to develop master calibration
200 curves for HumM2 and HumM3 DNA standards. An analysis of covariance (ANCOVA)
201 model was used to compare the intercept and slope of individual standard curves used to
202 calculate the master calibration equations. One-way ANOVA tests comparing C_T values
203 from reactions with known amounts of standard target DNA were used to define the
204 range of quantification (ROQ) for each assay. The precision of C_T measurements
205 determined from DNA standards was expressed as a percent coefficient of variation (CV,
206 standard deviation expressed as a percentage of the mean). A one-way random effect
207 ANOVA model (with location as random factor) was used to test the hypotheses that the
208 variability between untreated wastewater sample locations was zero. A paired two

209 sample t-test was used to compare the overall mean difference in C_T values between
210 HumM2, HumM3, Enterol, and GenBac3 qPCR mean C_T values for wastewater samples.

211

212

RESULTS

213

214 **Master calibration curves and range of quantification.** Overall fitted curves
215 representing multiple independent runs of the DNA standards were compared using
216 ANCOVA tests. Independent fitted curves for each qPCR assay demonstrated a
217 significant difference in intercepts ($p < 0.05$), but no difference in slopes ($p > 0.05$).
218 Calibration curve equations and performance characteristics of the four qPCR assays are
219 shown in Table 3. Calibration curves for GenBac3 and Enterol general fecal indicator
220 bacteria assays were generated from eight independent runs using genomic DNA
221 standards extracted from cultured cell suspensions whereas HumM2 and HumM3 fitted
222 curves were generated from 12 independent fitted curves each using plasmid DNA
223 standards. ROQs spanned the entire range of standard concentrations tested for all qPCR
224 assays including 10 to 1×10^6 copies for human-specific assays and 40 to 4×10^4 copies for
225 general fecal indicator bacteria assays. Precision of C_T measurements across defined
226 ROQs for all assays was less than 3% CV and amplification efficiencies ranged from
227 1.87 to 1.99 (Tab. 3). No-template controls indicated the absence of contamination in
228 98.9% of qPCR experiments and all extraction blanks tested negative for presence of
229 extraneous DNA molecules.

230

231 **Evaluation of multiplex host-specific qPCR application.** A composite synthetic
232 internal control was developed for each host-specific assay to monitor fecal and
233 wastewater DNA extracts for potential PCR inhibition. The IAC construct was designed
234 with the intention to allow target DNA and an IAC to be coamplified with the same set of
235 primers, under the same reaction conditions, in the same PCR tube. The target DNA and
236 IAC product could then be detected and quantified simultaneously with different
237 fluorescently labeled TaqMan® probes provided that (i) there is no significant difference
238 ($p > 0.05$) between simplex and multiplex standard curve intercepts and slopes and (ii) a
239 fixed amount of IAC could be quantified across a range of genomic DNA standard
240 concentrations (22). An IAC spike of 50 copies was undetectable at human fecal DNA
241 concentrations ranging from 1 to 100 ng for the HumM2 assay, while a significant
242 difference between simplex and multiplex curve intercepts and slopes was observed for
243 the HumM3 assay ($p < 0.05$) suggesting that neither of these assays is reliable as a
244 multiplex reaction (data not shown). The failure of both assays to perform in a multiplex
245 environment is most likely due to competition between genomic target DNA (6FAM
246 labeled) and the IAC spike (VIC labeled). Thus, only the HumM3 IAC could be used to
247 monitor for PCR inhibition and only in a simplex application.

248

249 **Monitoring for PCR Inhibition in DNA Extracts.** DNA isolation from wastewater and
250 fecal samples may not remove all substances that can interfere with qPCR and the degree
251 of interference may vary between samples. Therefore, internal controls designed to
252 evaluate the suitability of isolated DNA for quantitative analysis were included for each
253 DNA extract. All fecal DNA extracts were screened for inhibition of the HumM3 IAC

254 assay. The criterion for concluding no significant PCR inhibition of the HumM3 IAC
255 assay by these samples was established as a $C_T = 34.6 \pm 1.65$, based on repeated
256 experiments measuring the simplex mean C_T and standard deviation values for control
257 reactions containing 50 copies of IAC in buffer (Fig. 2, panel A). Wastewater DNA
258 extracts were also tested using the previously reported multiplex Entero1 application with
259 a 25 copy IAC spike. The criterion for concluding no significant PCR inhibition in these
260 assays was defined as a C_T of 34.0 ± 1.41 (Fig.2, panel B). IAC analyses indicated the
261 absence of PCR inhibitors in all fecal and untreated wastewater DNA extracts based on
262 both of these criteria.

263

264 **Specificity of host-specific qPCR assays.** Specificity of the HumM2 and HumM3
265 assays was tested with a reference collection of fecal samples from hundreds of non-
266 target animals (Tab. 4). HumM2 and HumM3 assays exhibited specificity values of
267 99.2% and 97.2%, respectively. HumM2 elicited false positives with two chicken fecal
268 samples ($C_T 29.3 \pm 0.16$ and 29.1 ± 0.14), while HumM3 cross-reacted with a single elk
269 sample ($C_T 33.6 \pm 0.35$) and six sheep samples (C_T ranging from 24.4 ± 0.05 to $36.9 \pm$
270 0.73). Both assays successfully detected respective DNA targets in all human fecal and
271 primary effluent wastewater DNA extracts (Tab. 4).

272

273 **Quantification of fecal bacterial genes in untreated wastewater.** Primary effluent
274 wastewater samples were collected from 20 different geographic locations to characterize
275 target DNA variability between localities and to compare the relative abundance of each
276 target DNA to enterococci and general *Bacteroidales* 16S rRNA genes. A one-way

277 random effect ANOVA model indicated that there is significant variability ($p < 0.05$) in
278 C_T values among all locations, for each assay. Variance (σ^2) between wastewater sample
279 locations ranged from 0.30 for HumM2, 1.06 for HumM3, 1.65 for Entero1 to 0.45 for
280 GenBac3. Target DNA relative abundance for each assay was compared by normalizing
281 data sets to 1 ng of template DNA and plotting \log_{10} mean copy number estimates for
282 each wastewater sample by qPCR assay. A box-and-whisker diagram was used to
283 display differences between wastewater sample DNA target estimates for each qPCR
284 assay including the smallest observation, lower quartile (25th percentile), median, upper
285 quartile (75th percentile), largest observation, and outliers (Fig. 3).

286

287

DISCUSSION

288

289 **Human-specific qPCR.** We report on two qPCR assays that detect predominantly
290 human fecal DNA when tested against a panel of samples representing agriculturally
291 important animals such as cattle, poultry, and swine as well as many wildlife species.
292 These qPCR assays were designed to target the same gene sequences as two end-point
293 PCR assays (Assays 19 and 22) that were previously reported to be 100% human-specific
294 based on a fecal reference collection consisting of 160 individual samples representing 11
295 different animal species (24). The slight decrease in specificity of the real-time qPCR
296 assays compared to the end-point PCR assays may be due to the larger non-target fecal
297 sample reference library used to establish specificity values or factors associated with the
298 TaqMan® qPCR approach such as constraints in primer design, PCR reagent chemistry,
299 thermal cycling settings, and increased number of amplification thermal cycles.

300 Regardless of the reason, the HumM2 and HumM3 qPCR assays exhibit extremely high
301 levels of specificity exceeding 97.5%.

302 Master calibration curves were used in this study due to the large numbers of fecal
303 and wastewater samples processed and the need to maximize the number of samples in
304 each experiment set-up and reduce expenses. Each master curve was compiled from up
305 to 12 independent runs in order to reflect sources of intra- and inter-run variability.

306 Master calibration curves were acceptable in this study because (i) there was no
307 significant differences in the slopes of fitted curves between independent runs ($p > 0.05$),
308 (ii) the analytical precision (%CV) over the ROQ between runs averaged less than 3%,
309 and (iii) the Bayesian approach accounts for run-to-run variability with a 95% credible
310 interval when generating fitted calibration curves (25).

311

312 **Abundance of host-specific and fecal indicator genes.** Little is known regarding the
313 abundance and geographical distribution of human-specific genes in sewage. In this
314 study, we tested 20 primary effluent wastewater samples collected from different
315 geographic locations in the United States ranging from Hawaii to Florida. Wastewater
316 samples were representative of approximately 4.1 million individuals responsible for
317 generating an average of 5,180 million gallons of raw sewage per year and were ideal for
318 estimating the abundance of host-specific gene targets in different human populations.
319 Host-specific and general fecal indicator bacteria qPCR assays successfully detected
320 respective genetic targets from 1 ng of DNA for 100% of the wastewater samples
321 regardless of locality. The general *Bacteroidales* assay (GenBac3) detected the highest
322 target gene concentration in all samples, which supports previous research reporting that

323 *Bacteroidales* often makes up a large portion of the human fecal bacterial community (6,
324 16, 28). The HumM2 and HumM3 gene targets were the next most abundant markers
325 and more prevalent than the enterococcal 23S rRNA genes (Fig. 3). Enterococci are
326 routinely detected in fecal polluted waters (26). The observation that host-specific gene
327 targets are more abundant than enterococcal 23S rRNA genes suggests that detectable
328 quantities of HumM2 and HumM3 gene targets may be present in ambient waters.

329 All qPCR assays exhibited less than 3.9% dispersion of C_T values from an overall
330 wastewater sample mean [(one-way random effect ANOVA qPCR standard
331 deviation/mean) x 100] regardless of gene target. In addition, a significant difference (p
332 < 0.05) was observed in concentrations of all qPCR gene targets between wastewater
333 geographic locations. Fluctuations in relative gene target concentrations between
334 wastewater samples could result from differences in local population diet, age, and/or
335 health, but could also reflect uncertainty associated with single sample events.
336 Regardless of the reason, low dispersion percentages ($< 3.9\%$) suggest that the human-
337 specific gene targets can be detected with a similar level of confidence as 16S rRNA
338 general *Bacteroidales* and 23S rRNA enterococcal gene targets.

339

340 **Implications for microbial source tracking (MST).** Recreational and drinking source
341 waters continue to be impacted by human fecal pollution and can impose a direct threat to
342 human health (4, 5, 19). In addition to human waste, many other agricultural and wildlife
343 animal sources can contribute to the total fecal load. Most MST methods attempt to
344 identify specific fecal sources to help local authorities prioritize polluted areas for
345 restoration. Recent advances in PCR-based methods now allow for the estimation of host-

346 specific DNA target concentrations. These quantitative approaches can extend the utility
347 of MST applications by supplying information regarding the concentration of host-
348 specific fecal pollution sources. To date, no qPCR-based method has been found to be
349 100% specific for human fecal pollution (10, 12, 18, 21). Animals that cohabitate with
350 humans such as cats and dogs, and animals that share similar digestive physiologies, such
351 as pigs, are the most problematic. Fecal pollution originating from pets can confound
352 MST studies where cat and dog waste is mixed with sewage and/or runoff after rain
353 events. A similar problem can arise in watersheds impacted by swine sources of fecal
354 pollution. HumM2 and HumM3 are the first qPCR assays available that can discriminate
355 between all three of these sources of fecal pollution. In addition, these assays can
356 quantify as few as 10 copies of target DNA per reaction with a high degree of precision.
357 DNA targets of these assays were widely distributed among 20 different human
358 populations and more abundant than fecal enterococci in almost all wastewater samples
359 tested.

360 To explore the potential of the HumM2 and HumM3 qPCR assays for
361 environmental monitoring, each assay underwent preliminary testing with DNA isolated
362 from river, stream, and storm water samples (n = 6). All six samples contained general
363 *Bacteroidales* target sequences (GenBac3 C_T values ranging from 34.3 ± 1.26 to 26.2 ±
364 0.10) suggesting the presence of fecal pollution. Two of these samples, both collected
365 from locations situated within 100 m downstream of a wastewater discharge pipe,
366 generated C_T values for both host-specific assays (C_T values ranging from 35.8 ± 0.46 to
367 34.7 ± 0.32). These preliminary results combined with the high levels of specificity and
368 broad distribution of their DNA targets in wastewater samples suggest that the HumM2

369 and HumM3 assays may have future utility in MST applications. However, to realize the
370 full potential of these qPCR assays, several issues remain to be addressed. Future studies
371 characterizing the survival of target DNA molecules through the wastewater treatment
372 process and in the environment are needed to generate reliable estimates of the impact of
373 these sources on ambient water samples. Research projects focusing on the relevance of
374 each qPCR assay to current culture-based and qPCR-based fecal indicator methods (such
375 as *E. coli* and enterococi) are also critical for successful MST applications. Finally,
376 epidemiological studies are necessary to establish any links between the prevalence of
377 host-specific DNA targets and relevant public health risks.

378

379

ACKNOWLEDGEMENTS

380 The U.S. Environmental Protection Agency, through its Office of Research and
381 Development, funded and managed, or partially funded and collaborated in, the research
382 described herein. It has been subjected to the Agency's peer and administrative review
383 and has been approved for external publication. Any opinions expressed in this paper are
384 those of the author (s) and do not necessarily reflect the views of the Agency, therefore,
385 no official endorsement should be inferred. Any mention of trade names or commercial
386 products does not constitute endorsement or recommendation for use.

387

388

389

REFERENCES

390

391

392

393

394

395

396

397

398

399

400

401

402

1. **ABI.** 2002. TaqMan Universal PCR Master Mix Protocol. Applied Biosystems.
2. **Acinas, S. G., L. A. Marcelino, V. Klepac-Ceraj, and M. F. Polz.** 2004. Divergence and redundancy of 16S rRNA sequences in genomes with multiple *rrn* operons. *Journal of Bacteriology* **186**:2629-2635.
3. **Andersson, A. F., M. Lindberg, H. Jakobsson, F. Backhed, P. Nyren, and L. Engstrand.** 2008. Comparative analysis of human gut microbiota by barcoded pyrosequencing. *Plos One* **3**:e2836.
4. **Balarajan, R., V. Soni Raleigh, P. Yuen, D. Wheeler, D. Machin, and R. Caftwright.** 1991. Health risks associated with bathing in sea water. *British Medical Journal* **303**:1444-1445.
5. **Dufour, A. P.** 1984. Health effects criteria for fresh recreational waters. U.S. Environmental Protection Agency.

- 403 6. **Eckburg, P. B., E. M. Bik, C. N. Bernstein, E. Purdom, L. Dethlefsen, M.**
404 **Sargent, S. R. Gill, K. E. Nelson, and D. A. Relman.** 2005. Diversity of human
405 intestinal microbial flora. *Science* **308**:1635-1638.
- 406 7. **Grimont, F., and P. A. Grimont.** 1986. Ribosomal ribonucleic acid gene
407 restriction patterns as potential taxonomic tools. *Annals of the Institute of*
408 *Pasteur in Microbiology* **137B**:165-175.
- 409 8. **Haugland, R. A., S. C. Sieftring, L. J. Wymer, K. P. Brenner, and A. P.**
410 **Dufour.** 2005. Comparison of *Enterococcus* measurements in freshwater at two
411 recreational beaches by quantitative polymerase chain reaction and membrane
412 filter culture analysis. *Water Research* **39**:559-568.
- 413 9. **Higuchi, R., B. Krummel, and R. K. Saiki.** 1988. A general method of in vitro
414 preparation and specific mutagenesis of DNA fragments: study of protein and
415 DNA interactions. *Nucleic Acids Research* **16**:7351-7367.
- 416 10. **Kildare, B. J., C. M. Leutenegger, B. S. McSwain, D. G. Bambic, V. B. Rajal,**
417 **and S. Wuertz.** 2007. 16S rRNA-based assays for quantitative detection of
418 universal, human-, cow-, and dog-specific fecal *Bacteroidales*: A Bayesian
419 Approach. *Water Research* **41**:3701-3715.
- 420 11. **Kuwahra, T., A. Yamashita, H. Hirakawa, H. Nakayama, H. Toh, N. Okada,**
421 **S. Kuhara, M. Hattori, T. Hayashi, and Y. Ohnishi.** 2004. Genomic analysis of
422 *Bacteroides fragilis* reveals extensive DNA inversions regulating cell surface
423 adaptation. *Proceedings of the National Academy of Sciences of the United States*
424 *of America* **101**:14919-149-24.
- 425 12. **Layton, A., L. McKay, D. Williams, V. Garrett, R. Gentry, and G. Sayler.**
426 2006. Development of *Bacteroides* 16S rRNA gene TaqMan-based real-time PCR
427 assays for estimation of total, human, and bovine fecal pollution in water. *Applied*
428 *and Environmental Microbiology* **72**:4214-4224.
- 429 13. **Ludwig, W., and K. H. Schleifer.** 2000. How quantitative is quantitative PCR
430 with respect to cell counts? *Systematic and Applied Microbiology* **23**:556-562.
- 431 14. **Lunn, D. J., A. Thomas, N. Best, and D. Spiegelhalter.** 2000. WinBUGS-A
432 Bayesian modeling framework: Concepts, structure, and extensibility. *Statistics*
433 *and Computing* **10**:325-337.
- 434 15. **Matsuki, T., K. Watanabe, J. Fujimoto, T. Takada, and R. Tanaka.** 2004. Use
435 of 16S rRNA gene-targeted group-specific primers for real-time PCR analysis of
436 predominant bacteria in human feces. *Applied and Environmental Microbiology*
437 **70**:7220-7228.
- 438 16. **Moore, W. E. C., and I. V. Holdeman.** 1974. Human fecal flora: the normal
439 flora of 20 Japanese-Hawaiians. *Applied and Environmental Microbiology*
440 **27**:961-979.
- 441 17. **Nelson, K. E., R. D. Fleischmann, R. DeBoy, I. T. Paulsen, D. E. Fouts, J. A.**
442 **Eisen, S. C. Daugherty, R. J. Dodson, A. S. Durkin, M. Gwinn, D. H. Haft, J.**
443 **Kolonay, W. C. Nelson, T. Mason, L. Tallon, J. Gray, D. Granger, H.**
444 **Tettelin, H. Dong, J. L. Galvin, M. J. Duncan, F. E. Dewhirst, and C. M.**
445 **Fraser.** 2003. Complete genome sequence of the oral pathogenic bacterium
446 *Porphyromonas gingivalis* strain W83. *Journal of Bacteriology* **185**:5591-5601.
- 447 18. **Okabe, S., N. Okayama, O. Savichtcheva, and T. Ito.** 2007. Quantification of
448 host-specific *Bacteroides-Prevotella* 16S rRNA genetic markers for assessment of

- 449 fecal pollution in freshwater. *Applied Microbiology and Biotechnology* **74**:890-
450 901.
- 451 19. **Scott, T. M., J. B. Rose, T. M. Jenkins, S. R. Farrah, and J. Lukasik.** 2002.
452 Microbial source tracking: current methodology and future directions. *Applied*
453 *and Environmental Microbiology* **68**:5796-5803.
- 454 20. **Seifring, S. C., M. Varma, E. Atikovic, L. J. Wymer, and R. A. Haugland.**
455 2008. Improved real-time PCR assays for the detection of fecal indicator bacteria
456 in surface waters with different instrument and reagent systems. *Journal of Water*
457 *and Health* **6**:225-237.
- 458 21. **Seurinck, S., T. Defoirdt, W. Verstraete, and S. Siciliano.** 2005. Detection and
459 quantification of human-specific HF183 *Bacteroides* 16S rRNA genetic marker
460 with real-time PCR for assessment of human faecal pollution in freshwater.
461 *Environmental Microbiology* **7**:249-259.
- 462 22. **Shanks, O. C., E. Atikovic, A. D. Blackwood, J. Lu, R. T. Noble, J. Santo**
463 **Domingo, S. Seifring, M. Sivaganesan, and R. P. Haugland.** 2008. Quantitative
464 PCR for detection and enumeration of genetic markers of bovine fecal pollution.
465 *Applied and Environmental Microbiology* **74**:745-752.
- 466 23. **Shanks, O. C., J. W. Santo Domingo, R. Lamendella, C. A. Kelty, and J. E.**
467 **Graham.** 2006. Competitive metagenomic DNA hybridization identifies host-
468 specific microbial genetic markers in cow fecal samples. *Applied and*
469 *Environmental Microbiology* **72**:4054-4060.
- 470 24. **Shanks, O. C., J. W. Santo Domingo, J. Lu, C. A. Kelty, and J. E. Graham.**
471 2007. Identification of bacterial DNA markers for the detection of human fecal
472 pollution in water. *Applied and Environmental Microbiology* **73**:2416-2422.
- 473 25. **Sivaganesan, M., S. Seifring, M. Varma, R. A. Haugland, and O. C. Shanks.**
474 2008. A Bayesian method for calculating real-time quantitative PCR calibration
475 curves using absolute plasmid DNA standards. *BMC Bioinformatics* **9**:120.
- 476 26. **USEPA.** 2000. Improved enumeration methods for recreational water quality
477 indicators: enterococci and *Escherichia coli* Report EPA-821-R-97-004. United
478 States Environmental Protection Agency.
- 479 27. **Wang, R. F., M. L. Beggs, B. D. Erickson, and C. E. Cerniglia.** 2004. DNA
480 microarray analysis of predominant human intestinal bacteria in fecal samples.
481 *Molecular and Cellular Probes* **18**:223-234.
- 482 28. **Wood, J., K. P. Scott, G. Avgustin, C. J. Newbold, and H. J. Flint.** 1998.
483 Estimation of the relative abundance of different *Bacteroides* and *Prevotella*
484 ribotypes in gut samples by restriction enzyme profiling of PCR-amplified 16S
485 rRNA gene sequences. *Applied and Environmental Microbiology* **64**:3683-3689.
- 486 29. **Xu, J., M. K. Bjursell, J. Himrod, S. Deng, L. K. Carmichael, H. C. Chiang,**
487 **L. V. Hooper, and J. I. Gordon.** 2003. A genomic view of the human-
488 *Bacteroides thetaiotaomicron* symbiosis. *Science* **299**:2074-2076.
- 489 30. **Zhang, Y., D. Zhang, L. Wenquan, J. Chen, Y. Peng, and W. Cao.** 2003. A
490 novel real-time quantitative PCR method using attached universal template probe.
491 *Nucleic Acids Research* **31**:e123.
492

493 **Table 1:** Primary effluent wastewater sample information.
 494

Facility	Location	Population Served	Inflow (mgd)
Sacramento RWTP	Sacramento, CA	1,200,000	168
Clarksburg WWTP	Clarksburg, WV	24,498	8.65
Lincoln Northeast WWTF	Lincoln, NE	55,000	5
Lower East Fork WWTP	Milford, OH	55,000	6.53
West Point WWTP	Seattle, WA	1,400,000	98.1
Crystal Lake WWTP No.2	Crystal Lake, IL	38,600	5.8
Little Falls WWTP	Little Falls, NY	49,000	5.14
Wildcat Hill WWTP	Flagstaff, AZ	60,000	3.3
Northwest Bergen County WWTP	Waldwick, NJ	102,448	10
Moorehead WWTP	Moorehead, KY	20,454	2.5
Buffalo WWTP	Buffalo, MO	6,000	0.72
Saginaw WWTP	Saginaw, MI	57,523	25
Bonner Springs WWTP	Bonner Springs, KS	7,500	0.53
Frankurt Sewer Department	Frankfurt, KY	48,000	6.69
Old Town PCF	Old Town, ME	9,500	1.2
Rutland WWTP	Rutland, VT	22,000	5.85
Maui County Kahului WWTF	Kahului, HI	41,720	4.3
City of St. Peter WWTP	St. Peter, MN	10,850	1.1
Las Vegas WWTP	Las Vegas, NV	815,207	68
Marshall St. Advanced WWTP	Clearwater, FL	65,000	5.3
Totals		4,088,300	431.7

495
 496 Inflow indicates the average rate of sewage influent at each treatment facility reported in million gallons
 497 per day (mgd).

498 **Table 2:** Oligonucleotides, primers, and probes.

499

Assay	Primer and probe sequences (5' to 3')	Size (bp)	Reference
GenBac3	Forward: GGGGTTCTGAGAGGAAGGT; Reverse: AGTAGCGGAAGGATGACGG; Probe: FAM-CAATATTCCTCACTGCTGCCTCCCGTA-TAMRA	129	(20)
Enterol	Forward: AGAAATTCCAAACGAACTTG; Reverse: AATGATGGAGGTAGAGCAC; Probe: FAM-TGGTTCTCTCCGAAATAGCTTTAGGGCTA-TAMRA	92	(13)
HumM2	Hum2F: CGTCAGGTTTGTTCGGTATTG; Hum2R: TCATCACGTAACCTATTTATATGCATTAGC; Probe: FAM-TATCGAAAATCTCACGGATTAACCTTTGTGTACGC-TAMRA	101	This Study
HumM3	Hum3F: GTAATTCGCGTTCTTCCTCACAT; Hum3R: GGAGGAAACAAGTATGAAGATAGAAGAATTAA; Probe: FAM-AGGTCTGTCCTTCGAAATAGCGGT-TAMRA	83	This Study
Human IAC	Frag1:GATCATGAGTTCACATGTCCGAGTAATTCGCGTTCTTCCTCACATACGTCAGGTTT GTTTCGGTATTG AGTTAGGAACAGGCGGCGACGAATG TTAATCTTCTATCTTC; Frag2:TCCGGTGATGTCTCGAGAGTGTCTCATCACGTAACCTATTTATATGCATTAGCGGT GAAGGTCTGGGAGGAAACAAGTATGAAGATAGAAGAATTAACATTCGTCGCCGC; Frag3:AGTTAGGAACAGGCGGCGACGAATGTTAATTCCTTCTATCTTCATACTTGTTTCCTC CCAGACCTTCACCGCTAATGCATATAAATAAGTTACGTGATGAGACACTCTCGA; Frag4:CCGTCATCCTTCACGCTACTGATGTCTGCATGGTATATGTTGAGTGCAATGGGATT TTATCCGGTGAATCCGGTGATGTCTCGAGAGTGTCTCATCACGTAACCTATTTA; Probe: VIC-TAGGAACAGGCGGCGACGA-TAMRA ^a	258	(30); This Study

500

501

502

^a The TaqMan probes was modified from the previously reported UT probe (30).

503 **Table 3:** Calibration curve equations and performance characteristics of qPCR assays.
 504

Assay	Calibration Equation	Amplification Efficiency ^a	ROQ (copies) for target DNA	%CV across ROQ	Method ^b
Entero1	Y = 38.0 -3.42X	1.96	40-4x10 ⁴	2.24	Multiplex
GenBac3	Y = 38.1 -3.34X	1.99	40-4x10 ⁴	2.92	Simplex
HumM2	Y = 41.8 -3.67X	1.87	10-1x10 ⁶	2.46	Simplex
HumM3	Y= 41.9 -3.66X	1.88	10-1x10 ⁶	2.40	Simplex

505
 506
 507
 508

^a Amplification efficiency = 10^(1/-slope).

^b Either a simplex approach or multiplex strategy where the target DNA was simultaneously detected with an IAC.

509 **Table 4:** Specificity of HumM2 and HumM3 qPCR assays.
 510

Animal Source	No.	HumM2		HumM3	
		Average C_T	St. Dev.	Average C_T	St. Dev.
Alpaca	2
Cow	80
Goat	7
Sheep	10	34.1	0.03	.	.
Horse	12
Pig	22
Antelope	4
Whitetail Deer	15
Mule Deer	5
Moose	1
Elk	5	35.7	0.24	.	.
Canadian Goose	12
Duck	12
Pelican	5
Gull	12
Turkey	7
Chicken	10	.	.	32.1	0.30
Marine Dolphin	3
California Sea Lion	5
Cat	10
Dog	10
Human	16	29.7	0.03	30.3	0.10
Wastewater	20	31.8	0.54	32.8	1.01
Total	285				

511
 512 C_T values generated from 1 ng of total DNA.

513 **Fig. Captions:**

514

515 **Fig1:** Diagram of human-specific plasmid DNA IAC composite construct. The IAC
516 (258 bp) consists of a VIC-labeled universal probe binding site (30) flanked by primer
517 sequences for HumM2 (101 bp) and HumM3 (83 bp) qPCR assays.

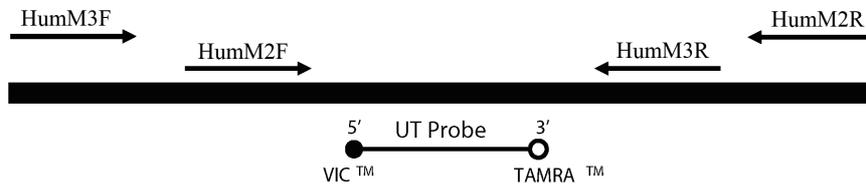
518

519 **Fig2:** Results of qPCR IAC inhibition tests for fecal and wastewater DNA extracts.
520 Scatter plots show IAC (VIC® or TET probes) and genomic DNA (6FAM probe) C_T
521 values from analyses of fecal DNA extracts using HumM3 (Panel A) and wastewater
522 DNA extracts using Enterol (Panel B). Confidence intervals (dashed lines) represent
523 three standard deviations of the mean IAC C_T (solid lines; HumM3 $C_T = 34.6$ and
524 Enterol $C_T = 34.0$) established from repeated control experiments.

525

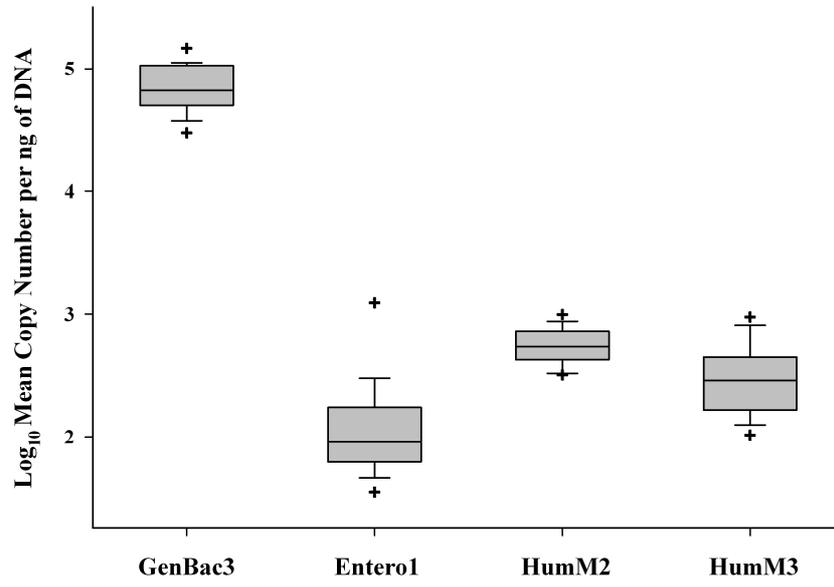
526 **Fig3:** Box-and-whisker diagram depicting the relative abundance of gene targets from
527 HumM2, HumM3, Enterol, and GenBac3 qPCR assays from all primary effluent sewage
528 sample locations. Estimated gene target concentrations are reported as \log_{10} mean copy
529 number per ng of total DNA. The boundary of the box closest to zero indicates the 25th
530 percentile, the line within the box represents the median, and the boundary of the box
531 farthest from zero indicates the 75th percentile. Whiskers above and below the box
532 indicate the 10th and 90th percentiles. A “+” denotes outlier measurements.

533 **Fig. 1:**
534



535
536

541 **Fig. 3:**
542



543
544